

1 INTRODUCTION

2 The spread, establishment and impacts of the spiny water
3 flea, *Bythotrephes longimanus*, in temperate North America:
4 a synopsis of the special issue

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9 Abstract More than most sub-disciplines of ecol-
10 ogy, the study of biological invasions is characterized
11 by breadth rather than by depth. Studies of expanding
12 ranges of invaders are common, as are post-invasion
13 case studies, but we rarely have a deep understanding
14 of the dynamics and regulators of the processes of
15 invasion and resultant ecological transformations.
16 This is unfortunate because such depth may well be
17 needed to develop targeted, knowledge-based, man-
18 agement plans. In this collection we provide this
19 needed depth of study of the key aspects of the
20 invasion process for the spiny water flea, *Bythotre-*
21 *phes longimanus*. We do so by presenting the results
22 of the work conducted by researchers in the Canadian
23 Aquatic Invasive Species Network (CAISN), and

several of their American and European collaborators 24
over the past half decade. Given its rapid spread in 25
the Great Lakes basin in North America, and the 26
decreases in pelagic biodiversity that have ensued, 27
the last decade has witnessed a surge of research on 28
Bythotrephes. In this collection we learn much about 29
mechanisms and dynamics of its spread, about the 30
key role of humans in that spread, about the 31
importance of Allee effects to establishment and 32
persistence, about choices and parameterization of 33
risk assessment models, about the value of comparing 34
“effects” in native and invaded regions, about 35
complex probable interactions of the invasion with 36
impending changes in the climate, and about the 37
regulators of the invader’s abundance and impacts. 38
There should be much of interest in the collection for 39

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40 aquatic ecologists and invading species biologists
41 alike.

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44

46 Introduction: on the relevance of *Bythotrephes*
47 *longimanus* to invasion biologists

48 Most invading species biologists work on the land, or
49 in the littoral regions of lakes and oceans, ecozones
50 that together form roughly a third of the planet's
51 surface. Pelagic ecosystems, both fresh and saline,
52 blanket the remaining two-thirds of the earth, and the
53 key biota that underpin the productivity of these
54 waters are the plankton. Based on their areal coverage
55 alone, it should come as no surprise, then, that the
56 services provided by plankton are important to
57 mankind. A healthy global plankton community
58 supplies humankind with services we either cannot
59 do without, e.g. atmospheric oxygen, or without
60 which our lives would be greatly impoverished, e.g.
61 essential fatty acids (Arts et al. 2001). The provision

62 of these and many other services from the pelagian
63 relies on its continued productivity and function, both
64 of which are underpinned by planktonic biodiversity
65 (Dodson et al. 2000; Cardinale 2011). Thus, any
66 serious anthropogenic threat to the biodiversity of
67 pelagic waters deserves our scrutiny, followed,
68 hopefully, by our enlightened management (Vander
69 Zanden and Olden 2008). Planktonic invaders are
70 now quite common in lakes and oceans (Bollens et al.
71 2002), and some of these species may pose a serious
72 threat to pelagic biodiversity. Unfortunately these
73 invaders have rarely received much scrutiny, but one
74 exception to this pattern is the spiny water flea,
75 *Bythotrephes longimanus* (Crustacea, Onychopoda,
76 Cercopagidae)—the world's best studied invasive
77 zooplankton (Bollens et al. 2002; Strecker in press).
78 There has been a surge of recent interest in the
79 impacts of *Bythotrephes* on pelagic freshwaters, and
80 we highlight this research in this special issue.

81 *Bythotrephes longimanus* (Fig. 1a) was more than
82 likely introduced to North America via ballast water
83 discharged from ships that picked it up in ports in the
84 northwest (Berg et al. 2002), or perhaps other regions
85 (Colautti et al. 2005) of Europe. It was misidentified

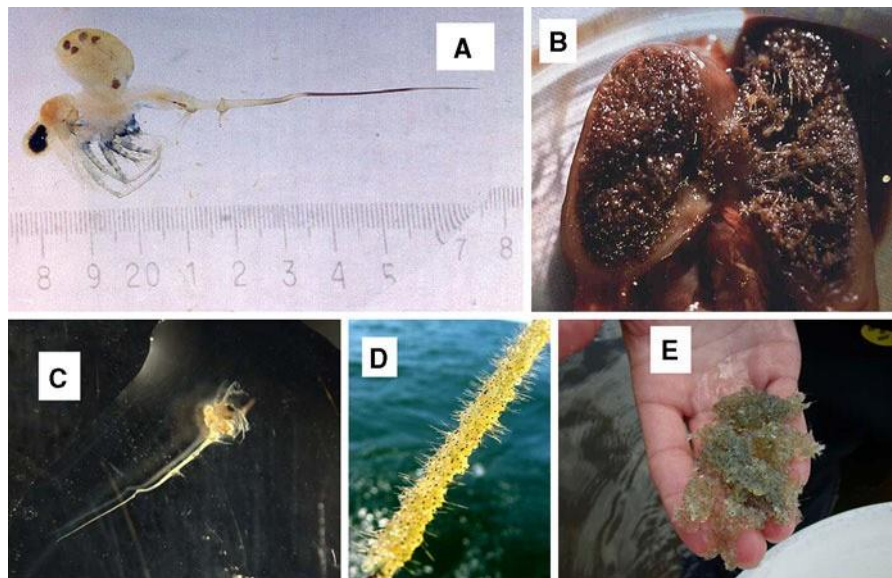


Fig. 1 a Photograph of a mature *Bythotrephes* with 5 late-stage embryos in her brood pouch. Animal was collected from Harp Lake, in Muskoka, Ont, Canada (photograph by Bill O'Neill), b photograph of *Bythotrephes* collected from the ballast tank of a ship in transit in the Great Lakes (photograph provided by Hugh MacIsaac, University of Windsor), c thousands of *Bythotrephes* in the stomach of a lake herring

(*Coregonus artedii*) from Lake Rosseau, District of Muskoka, Ontario (photograph by Bev Clarke), d a handful of *Bythotrephes* collected in a larval fish drift net in the Rainy River in northwestern Ontario, Canada (photograph provided by Ont. Min. Natural Resources), and e *Bythotrephes* on a downrigger fishing cable in Lake Erie (photograph by A. Jaeger)

86 in the earlier literature as *B. cederstræmi*, before the
 87 great polymorphism of *Bythotrephes* was recognized
 88 (Berg and Garton 1994; Therriault et al. 2002), and
 89 following prior naming conventions, *B. longimanus*
 90 was accepted as the proper binomial. It is a Ponto-
 91 Caspian zooplanktivore that has been established for
 92 millennia in large, temperate, nutrient-poor lakes in
 93 Eurasia (Grigorovich et al. 1998; MacIsaac et al.
 94 2000). By many criteria, it is an important member of
 95 its native pelagic ecosystems, for example, inhabiting
 96 about 20% of lakes in Norway (Hessen et al. 2011),
 97 contributing to salmonid fish diets out of proportion
 98 to its contribution to planktonic biomass (Nilsson
 99 1979, and Fig. 1b), and functioning as a key regulator
 100 of plankton composition (Manca et al. 2000). While
 101 *Bythotrephes* is not considered problematic in
 102 Europe, the situation is quite different in North
 103 America, where it has proven to be a serious threat to
 104 pelagic biodiversity in both large and small lakes
 105 (Yan et al. 2002; Barbiero and Tuchman 2004;
 106 Strecker et al. 2006). Its damaging effects cascade
 107 below its immediate crustacean prey to pelagic
 108 rotifers (Hovius et al. 2006), and likely to phyto-
 109 plankton (Strecker et al. 2011), and also up the food
 110 chain to competing macro-invertebrate predators
 111 (Foster and Sprules 2009; Weisz and Yan 2011)
 112 and fish (Parker-Stetter et al. 2005).

113 Students of biological invasions can learn much of
 114 general value from a deep examination of particular
 115 invaders. For example, we have learned much about
 116 the mechanisms of spread and establishment of
 117 invaders, about their ecological and socio-economic
 118 impacts, and about challenges and approaches to their
 119 management from focused research on *Caulerpa*, the
 120 “killer algae”, in the Mediterranean Sea (Meinesz
 121 1999), American comb jelly in the Black Sea (Kideys
 122 2002), zebra mussels in the Laurentian Great Lakes
 123 (Claudi and Mackie 1993), and Nile perch in Lake
 124 Victoria (Goldschmidt et al. 1993). Many of the key
 125 issues of interest to invading species biologists also
 126 apply to planktonic invaders, i.e. the mechanisms and
 127 dynamics of spread, the regulation of establishment
 128 success and post-establishment population growth,
 129 the subsequent ecological changes, their site speci-
 130 ficity, and their effects on ecological services (e.g.
 131 Myers and Bazely 2003; Lockwood et al. 2007). We
 132 deal with all of these issues in this collection. Our
 133 collective goal is to present to invading species
 134 biologists the latest knowledge on the mechanisms

and models of the spread, establishment, and impacts
 of *Bythotrephes* on freshwater ecosystems, princi-
 pally in eastern, temperate, North America. There are
 four specific reasons why *Bythotrephes* deserves such
 attention: (1) the apparent enormous threat it poses to
 North American pelagic biodiversity; (2) the many
 gaps in understanding of this threat which recent
 research can now plug; (3) its rapidity of spread,
 which lead to its selection by CAISN (the Canadian
 Aquatic Invading Species Network) as its model
 pelagic invader, thus providing us the opportunity to
 compare risk assessment models with different
 underlying drivers on a common data set; and 4)
 the need to better inform managers of best options to
 reduce the spread of this and other pelagic invaders.
 We consider each of these reasons in the following
 few paragraphs.

First, we believe *Bythotrephes* represents a wide-
 spread threat to pelagic biodiversity in temperate
 North America. It is spreading rapidly and widely, and
 severely damaging at least its planktonic prey. *Bytho-*
trephes was first identified in North America in Lake
 Ontario in the early 1980s (Johannsson et al. 1991). It
 has since spread rapidly colonizing all of the Laurentian
 Great Lakes by the end of the 1980s (Bur et al.
 1986; Lange and Cap 1986; Lehman 1987; Evans
 1988; Cullis and Johnson 1988), likely moved in
 ballast among the lakes by the Great Lakes shipping
 fleet (Fig. 1c). By the late 1980s and early 1990s, the
 invader appeared in a few inland lakes in Michigan,
 USA, and more than a dozen inland lakes in Ontario,
 Canada (Yan et al. 1992). During the 1990s it spread
 rapidly in Ontario, especially among recreational lakes
 in the District of Muskoka, a few hours north of
 Toronto (Yan and Pawson 1997; Therriault et al. 2002;
 Muirhead and MacIsaac 2005; Weisz and Yan 2010).
 By 2010, there were 150 known invaded lakes spread
 over a 1,300 km range from south-central to north-
 western Ontario, and in Canada the invader had spread
 beyond the Great Lakes watershed into the Hudson
 Bay drainage. During the same time period, many
 invasions were also documented in lakes and reser-
 voirs in Michigan, Minnesota, Wisconsin, Ohio and
 New York (Branstrator et al. 2006; Johnson et al.
 2008; Strecker et al. 2011, and Fig. 1 in Kerfoot et al.
 2011). Given the similar climates and water chemistry
 of Shield lakes in Canada and northern Europe, the
 20% prevalence of *Bythotrephes* in lakes in Norway
 (Hessen et al. 2011), and its rapid recent spread

184 (Kerfoot et al. 2011), we hypothesize that many
 185 thousands of lakes in temperate North America will
 186 eventually come to support this invader. As planktonic
 187 crustacean species richness typically falls by some
 188 20% after North American *Bythotrephes* invasions
 189 (Lehman and Caceres 1993; Schulz and Yurista 1999;
 190 Yan et al. 2002; Barbiero and Tuchman 2004; Strecker
 191 et al. 2006), we believe the eventual impacts of
 192 *Bythotrephes* on zooplankton biodiversity in N. Amer-
 193 ica will be enormous, assuming the initial impacts are
 194 long-lasting, which, to date, they appear to be (Yan
 195 et al. 2008).

196 Our second reason for assembling this special issue
 197 on *Bythotrephes* was that the work of a large number
 198 of researchers that entered the field in the last decade
 199 was nearing completion, and its collective publication
 200 could build the deep knowledge that the field needs.
 201 There is a reasonably large body of published
 202 *Bythotrephes* research on which to build (Fig. 2),
 203 but predictably, much of the early North American
 204 work is limited to reports of range expansions (e.g.
 205 Yan et al. 1992), and descriptive case studies (e.g. Yan
 206 and Pawson 1997), or what we might term first
 207 generation models of spread, which are not mecha-
 208 nistically-based (MacIsaac et al. 2000), nor built on
 209 data derived from probability-based surveys (Muir-
 210 head and MacIsaac 2005). However, since 2005,
 211 much has changed, especially with the Canadian
 212 Aquatic Invasive Species Network (CAISN) adopting

213 *Bythotrephes* as their model pelagic invader. CAISN
 214 recognized the need for in depth analysis of model
 215 systems to identify key issues underlying the mech-
 216 anisms of spread, establishment and impact of
 217 biological invaders. *Bythotrephes* was an obvious
 218 choice given their detrimental effects, the current
 219 concerns with respect to their spread, and the back-
 220 ground research that had already been completed that
 221 would facilitate more general advances in invasive
 222 species knowledge. Finally, CAISN together with the
 223 Ontario Ministry of the Environment hosted an
 224 international *Bythotrephes* workshop in Dorset,
 225 Ontario, Canada, in the fall of 2009, to bring together
 226 CAISN researchers and their North American and
 227 European colleagues, resulting in the development of
 228 this special issue. Here, we fill several fundamental
 229 holes in understanding about *Bythotrephes*. In terms
 230 of population and community dynamics, Brown and
 231 Branstrator (2011) and Wittmann et al. (2011) dem-
 232 onstrate the role of the resting egg biology of
 233 *Bythotrephes* on its invasion success, while Pich-
 234 lová-Ptáčníková and Vanderploeg (2011), Bourdeau
 235 et al. (2011), and Young et al. (2011), respectively,
 236 consider how differences in prey avoidance abilities,
 237 migration tendencies, and spring abundances can
 238 explain the invader's abundance, and its differential
 239 impacts on specific taxa. Hessen et al. (2011) and
 240 Jokela et al. (2011) compare the invader's interactions
 241 with native macro-invertebrate, holoplanktonic pre-
 242 dators in Norway and Canada; Kerfoot et al. (2011)
 243 prove the role of fish in its dispersal; while Rennie
 244 et al. (2011) document the overall changes in trophic
 245 structuring of food webs that follow invasion.

246 Our third reason for assembling this special issue
 247 is that *Bythotrephes* provides an excellent model for
 248 the study of the secondary spread of invaders. The
 249 CAISN initiative produced a common data set, which
 250 yielded the opportunity to compare alternative for-
 251 mulations of models of spread, produced by inde-
 252 pendent labs. Such comparisons are rarely possible,
 253 but are very useful for consideration of the conse-
 254 quences of subtle differences in model structure (i.e.,
 255 analysis of model uncertainty), for identification of
 256 the potential importance of different underlying
 257 invasion processes, and for testing alternative hypothe-
 258 ses when multiple processes or model structures
 259 yield similar fits to the data. We assemble that
 260 research here, with four papers focused on modeling
 261 the growth and spread of the invader on the south-

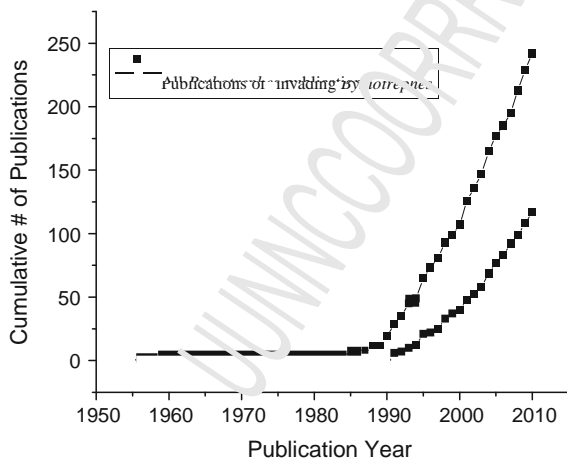


Fig. 2 Growth of the *Bythotrephes* literature, distinguishing all ISI-tracked publications with the keyword *Bythotrephes* from those specifically concerned with *Bythotrephes* as an invading species

- 356 2. Propagule pressure, linked to human recreational
 357 activity including fishing (Jarnagin et al. 1999),
 358 is likely the major determinant of the spread of
 359 *Bythotrephes* (e.g. Muirhead and MacIsaac 2005;
 360 Weisz and Yan 2010, e.g. Fig. 1d), but habitat
 361 quality may also affect establishment success of
 362 propagules (MacIsaac et al. 2000). Research in
 363 this collection dramatically enriches this under-
 364 standing. In independent modeling efforts,
 365 Potapov et al. (2011); Gertzen and Leung (2011),
 366 and Muirhead and MacIsaac (2011) all demon-
 367 strate the central role of propagule pressure in
 368 explaining the current pattern of *Bythotrephes*
 369 presence on the Canadian Shield. Further Gert-
 370 zzen and Leung (2011) prove that the component
 371 of propagule pressure contributed by stream
 372 connections in this landscape is so low it can
 373 be practically ignored, while it certainly can be
 374 high in much larger rivers (e.g. Fig. 1e). Wang
 375 and Jackson (2011) and Potapov et al. (2011)
 376 demonstrate that habitat information can improve
 377 predictions of invader prevalence, with consid-
 378 eration, respectively, of sport fish composition
 379 and habitat acidity, while Jokela et al. (2011)
 380 prove that interactions with numerous native
 381 macro-invertebrate predators will not slow the
 382 spread of the invader.
- 383 3. The collection proves that the regulators of
 384 establishment of *Bythotrephes* may vary from
 385 place to place in North America. On the Cana-
 386 dian Shield, lake connections in landscapes do
 387 not appear to influence the spread of the invader
 388 (Gertzen and Leung 2011), suggesting *Bythotre-*
 389 *phes* does not move between lakes in water
 390 masses. In Lake Superior, in contrast, Kerfoot
 391 et al. (2011) prove that currents may well control
 392 spread along coastlines and into embayments,
 393 while local temperature regimes may well control
 394 persistence.
- 395 4. There is a growing interest in the effects of
 396 climatic change on the spread of invaders. For
 397 *Bythotrephes*, it appears that present and future
 398 water temperatures may have a complex effect
 399 on the spread of *Bythotrephes*. Wittmann et al.
 400 (2011) predict that small increases in temperature
 401 should increase the probability of establishment
 402 of *Bythotrephes* by increasing rates of population
 403 growth of founding propagules to Allee effect
 404 thresholds that will lead to establishment.
- 405 However, *Bythotrephes* is a cool-water species,
 406 dying at temperatures just above 25°C (Grigorovich
 407 et al. 1998; Yurista 1999; Kim and Yan 2010);
 408 hence, climate warming should eventually alter
 409 the invader's spread and its eventual distribution.
- 410 5. Brown and Branstrator (2011) provide strong
 411 evidence that early seasonal introductions and
 412 large propagule sizes promote establishment of
 413 *Bythotrephes*, because the over-wintering sur-
 414 vival of its resting eggs can be surprisingly low,
 415 and turnover of resting eggs within a year can be
 416 surprising high. Persistence may well be depen-
 417 dent on the production of a great many resting
 418 eggs.
- 419 6. We learn much about Allee effects in this
 420 collection. Potapov et al. (2011), Wittmann
 421 et al. (2011), and Brown and Branstrator (2011)
 422 all provide evidence for a strong Allee effect
 423 influencing *Bythotrephes* establishment success,
 424 (see also Gertzen et al. 2011). Underlying
 425 mechanisms of Allee effects were also identified,
 426 in particular bottom-up control and starvation
 427 (Young et al. 2011) controlling summer popula-
 428 tion size, the rapid turnover and relative low
 429 survival rate of resting eggs (Brown and Bran-
 430 strator 2011), and temperature-limited growth
 431 (Wittmann et al. 2011), below thermal thresh-
 432 olds. Even relatively well established populations
 433 may fail in particularly hot years (Kerfoot et al.
 434 2011).
- 435 7. *Bythotrephes* are planktivorous, and influence
 436 prey populations directly by increasing their
 437 death rates, but they are also known to influence
 438 at least their daphnid prey indirectly, by altering
 439 their migratory behaviour and subsequent growth
 440 rates (Pangle et al. 2007). In this collection we
 441 learn more about such indirect effects. Jokela
 442 et al. (2011) demonstrate alterations in the
 443 vertical distributions of the invader's macro-
 444 invertebrate competitors, while Bourdeau et al.
 445 (2011) used chemical cues from the invader to
 446 induce alterations in the diel vertical distribution
 447 of copepods in Lake Michigan waters.
- 448 8. Much of the published work on the effects of
 449 *Bythotrephes* has been focused on alterations
 450 in pelagic structure, with limited work on
 451 function (Strecker and Arnott 2008), or on the
 452 determinants of *Bythotrephes* population size.
 453 In this collection, we learn that *Bythotrephes*,

- 454 by reducing abundances of herbivorous Clado- 503
 455 cera, alters trophic positioning of the entire 504
 456 pelagic assemblage (Rennie et al. 2011). We 505
 457 learn from an examination of Norwegian lakes 506
 458 that the ongoing replacement of *Leptodora* by 507
 459 *Bythotrephes* in North America (Foster and 508
 460 Sprules 2009; Weisz and Yan 2011) might well 509
 461 have been predicted from their co-occurrence 510
 462 patterns in Europe (Hessen et al. 2011). Finally, 511
 463 we learn that spring prey abundance may well 512
 464 be the prime determinant of *Bythotrephes* 513
 465 population size (Young et al. 2011), and perhaps, 514
 466 establishment success, given the large Allee 515
 467 effect. 516
- 468 9. Of the many species of *Daphnia* found in North 517
 469 American Lakes, only *D. mendotae* appears to 518
 470 thrive in the presence of *Bythotrephes*. In this 519
 471 collection, Hessen et al. (2011) demonstrate that 520
 472 the related *D. galeata* is one of few species that is 521
 473 actually positively associated with *Bythotrephes* 522
 474 in Norway. Pichlová-Ptácníková and Vander- 523
 475 ploeg (2011) provide compelling evidence to 524
 476 explain this persistence of *D. mendotae* in Lake 525
 477 Michigan with their demonstration that *D. men-* 526
 478 *dotae* has much faster escape responses to the 527
 479 invader than other daphniids, allowing it to 528
 480 prosper from the increased availability of 529
 481 resources left behind by its slower competitors. 530
- 482 10. Much has also been learned about modeling the 531
 483 risk of spread and establishment of invaders in 532
 484 this body of work (see especially point 2 533
 485 above). There are methodological advances, 534
 486 regarding the maximal usage of incomplete 535
 487 spatial and temporal information (Gertzen and 536
 488 Leung 2011), and the influence of the underly- 537
 489 ing structure of gravity models on their predic- 538
 490 tive ability. Production-constrained gravity 539
 491 models may well be the best overall choice 540
 492 (Muirhead and MacIsaac 2011). More funda- 541
 493 mentally we learn that the ongoing invasion of 542
 494 CAISN's key 1600-lake watershed is actually 543
 495 slowing, despite increased discovery rates, 544
 496 likely because of saturation of optimal sites 545
 497 (Gertzen and Leung 2011). 546
- 498 11. Beyond efforts to educate the public, there is 547
 499 currently no management directed specifically 548
 500 at *Bythotrephes*; hence, there is no article on 549
 501 *Bythotrephes* management in this collection. 550
 502 Nonetheless, there are many implications for 551
- management in the knowledge assembled in 503
 this collection. First, with a single sampling of 504
 300 of the 1600 lakes in an invaded watershed, 505
 it was possible to produce risk assessment 506
 models of several types that had a high prob- 507
 ability of predicting the pattern of occurrence of 508
Bythotrephes in a landscape. Clearly, such 509
 models can be developed from incomplete data 510
 sets for this invader, and likely for other 511
 invaders with similar life histories, such as 512
Cercopagis (Panov et al. 2007). Propagule 513
 pressure from humans emerged as the single 514
 best predictor of spread on the Shield in the 515
 work of Muirhead and MacIsaac (2011), Pota- 516
 pov et al. (2011), Gertzen and Leung (2011), 517
 and Kerfoot et al. (2011). This strongly suggests 518
 that management efforts are best directed at 519
 recreational lake users, especially boaters and 520
 anglers that are moving from invaded to non- 521
 invaded lakes. The recognition of strong Allee 522
 effects in several papers in this collection (and 523
 in Gertzen et al. 2011) counters earlier sugges- 524
 tions that only a few *Bythotrephes* colonists 525
 might found permanent populations (Drake 526
 et al. 2006), and clearly indicates that efforts 527
 to reduce propagule size and number, at least 528
 via public communication programs are justi- 529
 fied. We also learn from the collection that 530
 long-term establishment is not guaranteed, even 531
 if initial colonization success appears high, e.g. 532
 Portage Lake (Kerfoot et al. 2011). Hence, 533
 managers should endeavour to reduce propagule 534
 supply to lakes even after establishment, espe- 535
 cially for relatively shallow lakes that suffer 536
 occasional hot summers that may decimate the 537
 established population of invaders. 538
- In summary, the research contained in this collec- 539
 tion has taught us that, despite complex dynamics and 540
 interactions, the North American *Bythotrephes* estab- 541
 lishment, spread and impacts, can be understood in 542
 terms of key drivers. These drivers are the essential 543
 determinants of invasion outcomes. Establishment 544
 depends crucially upon dispersal at a level sufficient 545
 to overcome Allee effects. These Allee effects are, in 546
 turn, dependent upon local environmental conditions 547
 such as temperature. Once Allee effects are over- 548
 come, spread is quite predictable over broad spatial 549
 scales, determined first by anthropogenic dispersal in 550

552 Shield lakes, or anthropogenic dispersal coupled with
 553 water mass movements in the Great Lakes, and
 554 second by local environmental conditions. As with
 555 many invading species, impacts on biodiversity are
 556 fundamentally different in endemic and invaded
 557 ranges; hence, we take insights from work in endemic
 558 ranges, but not necessarily specific predictions of
 559 impacts. The impacts of the invader scale with its
 560 abundance, and the key driver that influences the
 561 invader's abundance and phenology in many, newly
 562 invaded North American lakes appears to be vernal
 563 prey density. The impacts may also be site-specific
 564 influenced by the capacity of native residents to avoid
 565 the predator either by changing their diel migratory
 566 behaviour, or, for a few taxa, having inherent escape
 567 abilities good enough to avoid capture. The regula-
 568 tion of impact is thus complex, including both direct,
 569 predatory drivers, and indirect behavioural drivers
 570 that differ among the invader, its prey, and likely its
 571 predators. A full unraveling of the food web inter-
 572 actions that govern these impacts is, perhaps unsur-
 573 prisingly, not yet available. Much has been learned,
 574 as the collection demonstrates. The threat to pelagic
 575 biodiversity that *Bythotrephes* represents should
 576 motivate continued research. We advise plankton
 577 ecologists and fisheries biologists that work in
 578 temperate lakes in North America to watch for
 579 *Bythotrephes* in their plankton and fish diet samples,
 580 given the rapid spread of this invader, and the damage
 581 to pelagic ecosystems that it causes. Importantly,
 582 many of these key drivers and issues are applicable to
 583 planktonic invaders in general. To the extent that a
 584 deep knowledge of one invader can inform the study
 585 and management of others, we hope that of the
 586 readers of the journal will benefit from this focused
 587 examination of one invader, the spiny water flea,
 588 *B. longimanus*.

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 592 research in Canada, and for funding the 2009 *Bythotrephes*
 593 workshop at Dorset, Ontario, at which the majority of the
 594 authors of this special issue compared their *Bythotrephes*
 595 research, and at which the idea for this special issue
 596 blossomed. We thank Julia Dewing for her secretarial,
 597 administrative, and organizational work. The co-editors'
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References

- Arts MT, Ackman RG, Holub BJ (2001) Essential fatty acids in
 aquatic ecosystems: a crucial link between diet and human
 health and evolution. *Can J Fish Aquat Sci* 58:122–137
- Barbiero RP, Tuchman ML (2004) Changes in the crustacean
 communities of Lakes Michigan, Huron, and Erie fol-
 lowing the invasion of the predatory cladoceran *Bytho-
 trephes longimanus*. *Can J Fish Aquat Sci* 61:2111–2125
- Berg DJ, Garton DW (1994) Genetic differentiation in North
 American and European populations of the cladoceran
Bythotrephes. *Limnol Oceanogr* 39:1503–1516
- Berg DJ, Garton DW, MacIsaac HJ, Panov VE, Telesh IV
 (2002) Changes in genetic structure of North American
Bythotrephes populations following invasion from Lake
 Ladoga, Russia. *Freshw Biol* 47:275–282
- Bollens SM, Cordell JR, Arent S, Hooff R (2002) Zooplankton
 invasions: a brief review, plus two case studies from the
 northeast Pacific. *Hydrobiol* 480:87–110
- Bourdeau PE, Pangle KL, Peacor SD (2011) The invasive
 predator *Bythotrephes* induces vertical migration in native
 copepods of Lake Michigan. *Biol Invasions* (this issue)
- Branstrator DK, Brown ME, Shannon LJ, Thabes M, Heimgartner
 K (2006) Range expansion of *Bythotrephes longimanus* in
 North America: evaluating habitat characteristics in the
 spread of an exotic zooplankton. *Biol Invasions* 8:1367–1379
- Brown M, Branstrator D (2011) Patterns in the abundance,
 phenology and hatching of the resting egg stage of the
 invasive zooplankton *Bythotrephes longimanus*. *Biol
 Invasions* (this issue)
- Bur MT, Klarer DM, Krieger KA (1986) First records of a
 European cladoceran, *Bythotrephes cederstroemi*, in lakes
 Erie and Huron. *J Great Lakes Res* 12:144–146
- Cardinale BJ (2011) Biodiversity improves water quality
 through niche partitioning. *Nature* 472:86–89
- Claudi R, Mackie GL (1993) Practical manual for zebra mussel
 monitoring and control. Lewis, London, pp 227
- Colautti RI, Manca M, Viljanen M, Ketelaars HAM, Burgi HR,
 MacIsaac HJ, Heath DH (2005) Invasion genetics of the
 Eurasian spiny waterflea: evidence for bottlenecks and
 gene flow using microsatellites. *Mol Ecol* 14:1869
- Cullis KI, Johnson GE (1988) First evidence of the cladoceran
Bythotrephes cederstroemi schoedler in lake superior.
J Great Lakes Res 14:524–525
- Dodson SI, Arnott SE, Cottingham KL (2000) The relationship
 in lake communities between primary productivity and
 species richness. *Ecology* 81:2662–2679
- Drake JM, Drury KLS, Lodge DM, Blukacz A, Yan ND, Dwyer
 G (2006) Demographic stochasticity, environmental vari-
 ability, and windows of invasion risk for *Bythotrephes
 longimanus* in North America. *Biol Invasions* 8:843–861
- Dumitru C, Sprules WG, Yan ND (2001) Impact of *Bythot-
 rephes cederstroemi* on zooplankton assemblages of Harp
 Lake, Canada: an assessment based on predator con-
 sumption and prey production. *Freshw Biol* 46:241–251
- Evans MS (1988) *Bythotrephes cederstroemi*: its new appear-
 ance in Lake Michigan. *J Great Lakes Res* 14:234–240
- Foster SE, Sprules WG (2009) Effects of the *Bythotrephes*
 invasion on native predatory invertebrates. *Limnol Oce-
 anogr* 54:757–769

- 661 Gertzen E, Leung B (2011) Predicting the spread of invasive
662 species in an uncertain world: accommodating multiple
663 vectors and gaps in temporal and spatial data for *Bytho-*
664 *trephes longimanus*. Biol Invasions (this issue)
- 665 Gertzen E, Leung B, Yan ND (2011) Propagule pressure,
666 stochasticity, and allee effects in relation to the proba-
667 bility of establishment of invasive species: an enclosure
668 study and population model of *Bythotrephes longimanus*.
669 Ecosphere 2(3):art30. doi:10.1890/ES10-00170.1
- 670 Goldschmidt T, Witte F, Wanink J (1993) Cascading effects of
671 the introduced nile perch on the detritivorous/phyto-
672 planktivorous species in the sublittoral areas of lake
673 Victoria. Conserv Biol 7:686–700
- 674 Grigorovich IA, Pashkova OV, Gromoca YF, van Overdijk
675 CDA (1998) *Bythotrephes longimanus* in the common-
676 wealth of independent states: variability, distribution and
677 ecology. Hydrobiol 379:183–198
- 678 Hessen D, Bakkestuen V, Walseng B (2011) Ecological niches
679 of *Bythotrephes* and *Leptodora*: lessons for predicting
680 long-term effects of invasion. Biol Invasions (this issue)
- 681 Hovius J, Beisner B, McCanmn KS (2006) Epilimnetic rotifer
682 community responses to bythotrephes longimanus inva-
683 sion in Canadian shield lakes. Limnol Oceanogr
684 51:1004–1012
- 685 Hovius JT, Beisner BE, McCann KS, Yan ND (2007) Indirect
686 food web effects of *Bythotrephes* invasion: responses by
687 the rotifer *Conochilus* in Harp Lake, Canada. Biol Inva-
688 sions 9:233–243
- 689 Jarnagin ST, Swan BK, Kerfoot WC (1999) Fish as vectors in
690 the dispersal of *Bythotrephes cederstroemi*: diapausing
691 eggs survive passage through the gut. Freshw Biol 43: 692
692 579–589
- 693 Johannsson OE, Mills EL, O’Gorman R (1991) Changes in the
694 nearshore and offshore zooplankton communities in Lake
695 Ontario: 1981–1988. Can J Fish Aquat Sci 48:1546–1557
- 696 Johnson PTJ, Olden JD, Vander Zanden MJ (2008) Dam
697 invaders: impoundments facilitate biological invasions
698 into freshwaters. Front Ecol Environ 6:357–363
- 699 Jokela A, Arnott S, Beisner B (2011) Patterns of *Bythotrephes*
700 *longimanus* distribution relative to native macroinverte-
701 brates and zooplankton prey. Biol Invasions (this issue)
- 702 Kerfoot WC, Yousef F, Hobmeier M, Maki RP, Jarnagin T,
703 Churchill JH (2011) Temperature, recreational fishing and
704 diapause egg connections: dispersal of spiny water fleas
705 (*Bythotrephes longimanus*). Biol Invasions (this issue)
- 706 Kideys AE (2002) Fall and rise of the black sea ecosystem.
707 Science 297:1482–1484
- 708 Kim N, Yan ND (2010) Methods for rearing the invasive
709 zooplankter *Bythotrephes* in the laboratory. Limnol Oeo-
710 anogr Methods 8:552–561
- 711 Lange C, Cap R (1986) *Bythotrephes cederstræmi* (Schœdler),
712 (Cercopagidae: Cladocera): a new record for lake Ontario.
713 J Great Lakes Res 12:142–143
- 714 Lehman JT, Caceres CE (1993) Food-web responses to species
715 invasion by an predatory invertebrate: *Bythotrephes* in
716 lake Michigan. Limnol Oceanogr 38:879–891
- 717 Lockwood JL, Hoopers MF, Marchetti MP (2007) Invasion
718 ecology. Blackwell, Oxford, p 304
- 719 MacIsaac HJ, Grigorovich IA, Hoyle JA, Yan ND, Panov VE
720 (1999) Invasion of lake Ontario by the Ponto-Caspian
721 predatory cladoceran *Cercopagis pengoi*. Can J Fish
722 Aquat Sci 56:1–5
- 723 MacIsaac HJ, Ketelaars HAM, Grigorovich IA, Ramcharan
724 CW, Yan ND (2000) Modeling *Bythotrephes longimanus*
725 invasions in the great lakes basin based on its European
726 distribution. Arch Hydrobiol 149:1–21
- 727 Makarewicz JD, Bertram P, Lewis T, Brown EH Jr (1995) A
728 decade of predatory control of zooplankton species of lake
729 Michigan. J Great Lakes Res 21:620–640
- 730 Manca M, Ramoni C, Comollie P (2000) The decline of
731 *Daphnia hyalina galeata* in Lago Maggiore: a comparison
732 of the population dynamics before and after oligotrophica-
733 tion. Aquat Sci 62:142–153
- 734 Meinesz A (1999) Killer algae: the true tale of a biological
735 invasion. University of Chicago, Chicago, p 360 (Press)
- 736 Mills EL, O’Gorman R, Degisi J, Heberger RF, House RA
737 (1992) Food of the alewife (*Alosa pseudoharengus*) in lake
738 Ontario before and after the establishment of *Bythotrephes*
739 *cederstroemi*. Can J Fish Aquat Sci 49:2009–2019
- 740 Muirhead JR, MacIsaac HJ (2005) Development of inland
741 lakes as hubs in an invasion network. J Appl Ecol
742 42:80–90
- 743 Muirhead JR, MacIsaac HJ (2011) Evaluation of stochastic
744 gravity model selection for use in estimating non-indige-
745 nous species dispersal and establishment. Biol Invasions
746 (this issue)
- 747 Myers JH, Bazely DR (2003) Ecology and control of intro-
748 duced plants. Cambridge University Press, p 313
- 749 Nilsson NA (1979) Food and habitat of the fish community of
750 the offshore region of lake Vanern, Sweden. Inst Freshw
751 Res Drottningholm 58:126–139
- 752 Palmer A, Stich HB, Maier G (2001) Distribution patterns and
753 predation risk of the coexisting cladocerans *Bythotrephes*
754 *longimanus* and *Leptodora kindtii* in a large lake—lake
755 constance. Hydrobiologia 442:301–307
- 756 Pangle KL, Peacor SD, Johannsson O (2007) Large nonlethal
757 effects of an invasive invertebrate predator on zooplank-
758 ton population growth rate. Ecology 88:402–412
- 759 Panov VE, Rodionova NV, Bolshagin PV, Bychek EA (2007)
760 Invasion biology of Ponto-Caspian onychopod cladocers-
761 ans (Crustacea: Cladocera: Onychopoda). Hydrobiol
762 590:3–14
- 763 Parker Stetter SL, Witzel LD, Rudstam LG, Einhouse DW,
764 Mills EL (2005) Energetic consequences of diet shifts in
765 lake Erie rainbow smelt (*Osmerus mordax*). Can J Fish
766 Aquat Sci 62:145–152
- 767 Pichlová-Ptáčnicková R, Vanderploeg HA (2011) The quick and
768 the dead: might differences in escape rates explain the
769 changes in the zooplankton community composition of
770 Lake Michigan after invasion by *Bythotrephes*? Biol
771 Invasions (this issue)
- 772 Potapov A, Muirhead J, Yan N, Lele S, Lewis M (2011)
773 Models of lake invasibility by *Bythotrephes longimanus*, a
774 non-indigenous zooplankton. Biol Invasions (this issue)
- 775 Rennie MD, Strecker AL, Palmer ME (2011) *Bythotrephes*
776 invasion elevates trophic position of zooplankton and fish:
777 Implications for contaminant biomagnification. Biol
778 Invasions (this issue)
- 779 Schulz KL, Yurista PM (1999) Implications of an invertebrate
780 predator’s (*Bythotrephes cederstroemi*) atypical effects on

- 781 a pelagic zooplankton community. *Hydrobiol* 380:179–
782 193
- 783 Strecker AL (In press) An overview of invasive freshwater
784 cladocerans: *Bythotrephes longimanus* as a case study. In:
785 Francis R (ed) Handbook of global freshwater invasive
786 species. Earthscan, London
- 787 Strecker AL, Arnott SE (2008) Invasive predator, *Bythotre-*
788 *phes*, has varied effects on ecosystem function in fresh-
789 water lakes. *Ecosystems* 11:490–503
- 790 Strecker AL, Arnott SE, Yan ND, Girard R (2006) Variation in
791 the response of crustacean zooplankton species richness
792 and composition to the invasive predator *Bythotrephes*.
793 *Can J Fish Aquat Sci* 63:2126–2136
- 794 Strecker AL, Beisner BE, Arnott SE, Paterson AM, Winter JG,
795 Johannsson OE, Yan ND (2011) Direct and indirect
796 effects of an invasive planktonic predator on pelagic food
797 webs. *Limnol Oceanogr* 56:179–192
- 798 Therriault TW, Grigorovich IA, Cristescu ME, Ketelaars
799 HAM, Viljanen M, Heath DD, MacIsaac HJ (2002) Tax-
800 onomic resolution of the genus *Bythotrephes* Leydig,
801 using molecular markers and a re-evaluation of its global
802 distribution, with notes on factors affecting dispersal,
803 establishment and abundance. *Divers Distrib* 8:67–84
- 804 Vander Zanden MJ, Olden JD (2008) A management frame-
805 work for preventing the secondary spread of aquatic
806 invasive species. *Can J Fish Aquat Sci* 65:1512–1522
- 807 Wang L, Jackson DA (2011) Modeling the establishment of
808 invasive species: habitat and biotic interactions influenc-
809 ing the establishment of *Bythotrephes longimanus*. *Biol*
810 *Invasion* (this issue)
- 811 Weisz EJ, Yan ND (2010) Relative value of limnological,
812 geographic and human use variables as predictors of the
813 presence of *Bythotrephes longimanus* in Canadian shield
814 lakes. *Can J Fish Aquat Sci* 67:462–472
- 815 Weisz EJ, Yan ND (2011) Shifting invertebrate zooplanktivores:
816 watershed-level replacement of the native *Leptodora* by the
non-indigenous *Bythotrephes* in Canadian shield lakes. *Biol*
Invasions 13:115–123
- Wittmann MJ, Lewis MA, Young JD, Yan ND (2011) Tem-
perature-dependent Allee effects in a stage-structured
model for *Bythotrephes* establishment. *Biol Invasions*
(this issue)
- Yan ND, Pawson TW (1997) Changes in the crustacean zoo-
plankton community of Harp Lake, Canada, following the
invasion by *Bythotrephes cederstræmi*. *Freshw Biol*
37:409–425
- Yan ND, Dunlop W, Pawson TW, Mackay LE (1992) *Bytho-*
trephes cederstræmi (Schœdler) in Muskoka lakes: first
records of the European invader in inland lakes in Canada.
Can J Fish Aquat Sci 49:422–426
- Yan ND, Blukacz A, Sprules WG, Kindy PK, Hackett D,
Girard R, Clark BJ (2001) Changes in the zooplankton
and the phenology of the spiny water flea, *Bythotrephes*,
following its invasion of Harp Lake, Ontario, Canada. *Can*
J Fish Aquat Sci 58:2341–2350
- Yan ND, Girard R, Boudreau S (2002) An introduced predator
(*Bythotrephes*) reduces zooplankton species richness. *Ecol*
Lett 5:481–485
- Yan ND, Somers KM, Girard RE, Paterson A, Keller B,
Ramcharan C, Rusak J, Ingram R, Morgan G, Gunn JM
(2008) Long-term changes in crustacean zooplankton
communities of Dorset, Ontario lakes: the probable
interactive effects of changes in pH, TP, dissolved organic
carbon, and predators. *Can J Fish Aquat Sci* 65:862–877
- Young JD, Strecker AL, Yan ND (2011) Increased abundance
of the non-indigenous zooplanktivore, *Bythotrephes lon-*
gimanus, is strongly correlated with greater spring prey
availability in Canadian shield lakes. *Biol Invasions* (this
issue)
- Yurista PM (1999) A model for temperature correction of size-
dependent respiration in *Bythotrephes cederstræmi* and
Daphnia middendorfianna. *J Plank Res* 21:721–734