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Synergetic Adsorption of Polymers on Montmorillonite: Insights from Molecular Dynamics Simulations

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Abstract

8 In the treatment of industrial tailings, fine clay minerals are generally flocculated using 9 polymer flocculants. Interestingly, applications of multiple types of polymers together were 10 reported to enhance the flocculation, although the underlying mechanism was still unclear. In 11 this work, a series of molecular dynamics simulations were performed to investigate the 12 synergetic adsorption of two types of polymers on the basal surface of montmorillonite (Mt) in 13 aqueous environment. Cationic chitosan molecules were first introduced, followed by the 14 addition of either neutral polyacrylamide (PAM) or anionic Magnafloc (MF). A monolayer 15 coating of chitosan was observed to form quickly on the Mt surface. The coating was stable and 16 facilitated the subsequent adsorption of PAM or MF which by themselves did not show effective 17 adsorption. Adsorption of PAM and MF showed different configurations. PAM could adsorb 18 either directly on Mt or on chitosan, in the form of clusters or individual molecules. On the other 19 hand, MF only adsorbed on chitosan, forming a two-layer structure above the Mt surface. 20 Adding chitosan simultaneously with PAM or MF, instead of sequentially, did not change the 21 characteristics of the adsorption. The synergetic adsorption of polymers was attributed to the 22 interplay of electrostatic attraction between Mt and chitosan, hydrogen bonding between chitosan

and PAM, as well as electrostatic attraction between chitosan and MF. Together, the simulation
 results allowed us to provide a mechanistic explanation for the experimentally observed
 synergetic flocculation enabled by adding multiple polymers.

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27 Keywords: montmorillonite; chitosan; polyacrylamide; magnafloc; synergy; adsorption

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29 **1. Introduction**

30 Removing fine particles and recycling water from industrial wastewater especially tailings 31 have become an urgent need in the past decades (Wang et al., 2014). After the coarse fraction of 32 tailings segregated, a gel-like dispersion could form with very slow rate of consolidation 33 (Bakhshi et al., 1975), which made the removal of fine particles very difficult. As a result, the 34 tailings required long-term storage structures causing huge wasteland (Watson et al., 2010). The 35 slow dewatering of industrial tailings, continuous exploitation of natural resources, and growing 36 industrial manufacturing led to global shortage of clean water and lands (Moreno and Neretnieks, 37 2006; Watson et al., 2010). Consequently, environmental impacts brought by the fine particles 38 have been gaining more attention in recent years (Moreno and Neretnieks, 2006; Kaniki and 39 Tumba, 2019).

The fine particles in mineral tailings are mainly clay minerals such as kaolinite, illite, and montmorillonite (Mt) (Omotoso and Mikula, 2004; Wang et al., 2014; Gorakhki and Bareither, 2015). It was found that basal surfaces of the clay minerals often carried permanent negative charges, while the protonation of hydroxyl groups made the charge on the edge surfaces pHdependent (Bourg et al., 2007; Zeitler et al., 2017; Chen et al., 2020). The negative charges were reported to be balanced by interlayer exchangeable cations, e.g., Na⁺, K⁺, Ca²⁺ (Theng, 2012). 46 Distributions of counter-ions and co-ions around the solid surface lead to the formation of the
47 diffuse electric double layer (EDL) (Israelachvili, 2011) which contributes to the stability of fine
48 clay minerals.

49 To reduce the stability of these mineral tailings, many chemical treatments were introduced 50 towards solid-liquid separation via flocculation (Long et al., 2006; Alamgir et al., 2012; Robert 51 et al., 2012; Nasim et al., 2018; Shaikh et al., 2018; Kang et al., 2019). Natural or synthetic 52 polymers and inorganic salts are considered as major types of flocculants. Many natural 53 polymers such as chitosan were shown to flocculate fine particles by adsorbing on the solid 54 surface through electrostatic attraction (Yang et al., 2011; Lu et al., 2016; Molatlhegi and Alagha, 55 2017). For example, Lu et al. (2016) reported that the zeta potential of particles in the 56 supernatants of mature fine tailings (MFT, concentrated dispersion of fine clay minerals 57 generated from water-based mining of oil sands) gradually increased from -38.6 mV to +30.2 58 mV when treated with increasing dosage of chitosan. Before reaching an optimum dosage (zeta 59 potential = 0 mV), as more chitosan was added, more particles were flocculated and settled 60 downwards, represented by the increasing initial settling rate (ISR) and decreasing turbidity of 61 the supernatant. On the other hand, hydrogen bonding (H-bonding) was also demonstrated to 62 facilitate the adsorption of polymers, such as neutral polyacrylamide (PAM) (Ji et al., 2013) as 63 well as its anionic derivatives Magnafloc (MF) (Ji et al., 2013), Zetag (Wang et al., 2016) and 64 poly (N-isopropyl acrylamide) (Zhang et al., 2017). These polymers have been extensively used 65 for solid-water separation since they are effective, inexpensive and have low toxicity. For instance, Li et al. (2005) observed an increase in ISR when tailing samples were treated by 66 67 anionic HPAM (hydrolyzed PAM). It was proposed that H-bonding between the hydroxyl groups

of HPAM and clay mineral surfaces could induce bridging interactions between the clay mineral
 particles, thus enhancing the phase separation.

70 Interestingly, instead of using a single type of polymers as flocculant, recent experimental 71 studies reported enhanced flocculation performance when two types of polymers were applied. 72 For example, in the work of Lu et al. (2016) the ISR reached as high as 10.26 m/h (meters per 73 hour) when the MFT were treated by MF at its optimum dosage, but the clarity of the supernatant 74 had no improvement compared with the untreated one. On the other hand, chitosan could render 75 a high clarity of the release supernatant, with turbidity of 15 NTU (nephelometric turbidity unit, 76 an indication of the relative clarity of water), but the highest ISR was only 1 m/h. Both high ISR 77 (7.7 m/h) and clear supernatant (turbidity of 71 NTU) were achieved by treating the MFT first by 78 MF and then by chitosan. Similar phenomena were also reported by other researchers (Zhang et 79 al., 2017).

80 Despite the observations from experiments, very limited theoretical work has been conducted 81 to investigate the mechanisms behind the solid-liquid separation, and as such there were 82 discrepancies between explanations proposed for the same observations (Theng, 2012). While 83 the adsorption of a single type of polymers onto clay mineral layers was examined by theoretical 84 simulations (Cygan et al., 2009; Cheng et al., 2015; Wei et al., 2017; Azevedo Rios Silva et al., 85 2018; Ren et al., 2020), there has been no theoretical studies on the possible synergy between 86 multiple types of polymer flocculants in the vicinity of clay mineral surfaces. The effects of 87 polymer charges, adding sequence, and how the synergy is tied to the flocculation capability are 88 among the most interesting but unaddressed questions in the field.

To fill this gap, in the present work a series of molecular dynamics (MD) simulations were performed to reveal the synergetic behaviors of two types of polymer flocculants when they were

91 added to the vicinity of a clay mineral surface. The clay mineral was represented by Mt with 92 negative charges. Three polymers were simulated: chitosan (cationic), PAM (neutral) and MF 93 1011 (anionic). Chitosan was added with either PAM or MF 1011 to study potential synergetic 94 effects. For comparison, simulations were also conducted on systems containing a single type of 95 polymers. As well, the effect of adding sequence was addressed by adding the two types of 96 polymers in different manners: simultaneously and sequentially.

97 **2. Simulation methods**

98 **2.1 Molecular models**

99 Mt belongs to the 2:1 mineral class, with a central sheet of octahedrally coordinated Al atoms, 100 sandwiched between two sheets consisting of SiO_4 tetrahedra (Viani et al., 2002). The unit cell of 101 Mt used in this work was based on the structure available in the American Mineralogist Crystal 102 Structure Database (Downs and Hall-Wallace, 2003), and their atomic coordinates were derived by Viani et al. (2002). The unit cell parameters were a = 5.18 Å, b = 8.95 Å, c = 15 Å, and $\alpha = \beta$ 103 104 $= \gamma = 90^{\circ}$, where a, b and c were lattice constants in the three directions respectively, and α , β 105 and γ were the angles between the lattice sides. Hydrogen atoms were manually added to oxygen 106 atoms in the octahedral sheet. A neutral $2 \times 2 \times 1$ supercell was then built by closely stacking the 107 unit cells. Three Al atoms in octahedral sheet were isomorphically substituted by Mg, resulting 108 in a negative layer charge, as the tetrahedral sheets were less favorable to be substituted than the 109 octahedral sheet. The negative layer charge was neutralized by sodium ions in the system, 110 yielding a unit cell formula of Na_{0.75}Si₈(Al_{3.25}Mg_{0.75})O₂₀(OH)₄. The Na-Mt supercell was 111 subsequently expanded to a $16 \times 10 \times 1$ supercell (Figure 1a) to accommodate the polymers.

112 Chitosan, PAM, and MF 1011 were all modeled based on compounds in the PubChem 113 database (Kim et al., 2019). A 10-mer chitosan ($C_{60}H_{112}N_{10}O_{41}$, MW 1638 g/mol, Figure 1b) was

built and all amino groups were protonated. The protonation ratio represented the condition of 114 115 pH = 8; such a pH was typical in the treatment of process water as a result of adding alkaline 116 reagents such as sodium hydroxide (Allen, 2008). Previous experimental works (Wang et al., 117 2006; Fernandes et al., 2012) also adopted the same pH value. MF 1011 was a sodium salt of 118 hydrolyzed PAM (Lu et al., 2016). MF ionomers with 9 units (C₂₇H₄₄N₆O₁₂, MW 641 g/mol, 119 Figure 1c) was simulated with a charge density of 33%. A neutral 10-mer PAM ($C_{30}H_{52}N_{10}O_{10}$, 120 MW 712 g/mol, Figure 1d) was used, corresponding to the same pH condition (Fraenkel and 121 Franconi, 1960; Grant et al., 2010).

122 2.2 Systems simulated

123 A total of nine systems (see Table 1) were simulated to study the adsorption of polymers on 124 the surface of Mt. System Mt-H₂O, with neutralizing sodium ions but without any polymers, was 125 first simulated to allow water and ions to equilibrate around the basal surface of Mt. Three 126 systems, namely Mt-CT, Mt-PAM, and Mt-MF, were simulated next to probe the adsorption of 127 single types of polymers on Mt. These simulations also allowed for comparison with systems 128 that involve two types of polymers. In each of these three systems, twelve molecules of chitosan, 129 PAM or MF were introduced into the equilibrated aqueous solution above the Mt, forming a 130 $3 \times 2 \times 2$ array as shown in Figure 1a. The number of sodium ions was adjusted accordingly in each 131 system to maintain charge neutrality.

The following two simulations, Mt-(CT-PAM) and Mt-(CT-MF), were performed where chitosan and PAM (in system Mt-(CT-PAM)) or chitosan and MF (in system Mt-(CT-MF)) were added simultaneously above the clay mineral surface. The polymers were also added in the form of a $3\times2\times2$ array as shown in Figure 1a, with one type of polymers forming the $3\times1\times2$ array on the left and the other on the right. Lastly, to study the effect of adding sequence, twelve MF molecules were placed above the Mt previously coated by chitosan, which resulted from the Mt-CT simulation. This formed the initial configuration for system (Mt-CT)-MF. System (Mt-CT)-PAM was constructed in the same manner, with MF replaced by PAM. One additional system, (Mt-PAM)-CT, was simulated by using the configuration of system Mt-PAM at 100 ns and adding twelve chitosan molecules, in the form of a $3 \times 2 \times 2$ array, above the Mt surface pre-coated by PAM.

143 **2.3 Simulation details**

144 The force field parameters for Mt were adopted from the CLAYFF force field (Cygan et al., 145 2004; Pouvreau et al., 2017). CLAYFF was designed for clay mineral models and has a good 146 compatibility with organic force fields. The partial charges of all the atoms remained the same as 147 in CLAYFF, however the potentials for van der Waals were converted into forms compatible 148 with GROMOS96 53A6 (Oostenbrink et al., 2004). Validation of the force field parameters is 149 given in Supporting Information (SI, Section S1). Force field parameters for chitosan, PAM and 150 MF were first obtained from the GlycoBioChem PRODRG server (Schüttelkopf and Van Aalten, 151 2004) by submitting the initial atomic coordinates and invoking the GROMOS96 force field 152 parameter set 53A6. The partial atomic charges were then manually adjusted based on the 153 density functional theory (DFT) calculation from Gaussian 16 (Frisch et al., 2016). More details 154 about the calculation of partial atomic charges of the three types of polymers are given in the SI 155 (Section S2).

All MD simulations were carried out using GROMACS (Berendsen et al., 1995; Pronk et al., 2013; Abraham et al., 2015) with periodic boundary conditions applied in all directions. As such, the polymers were essentially located between two basal surfaces of Mt. Water and sodium ions were first introduced and equilibrated for 2.5 ns in NPT ensemble. Then, polymers were solvated

into the equilibrated $Mt-H_2O$ system with the number of sodium ions adjusted accordingly. Each system was subjected to a geometrical optimization followed by NPT equilibration. Due to the different nature and strength of interactions between the polymers and Mt, different systems required different time to reach equilibrium (see Table 1, more details in SI Section S3).

The temperature was controlled at 300 K by Nose-Hoover thermostat (Nosé, 1984; Hoover, 1985), and pressure was controlled at 1 bar by Parrinello–Rahman barostat (Parrinello and Rahman, 1981). Leap-frog algorithm (Hockney et al., 1974) was used for the integration of equations of motion, with a time step of 1 fs. The bonds within all molecules were constrained by LINC algorithm (Hess et al., 1997) except those in H₂O which were constrained by SETTLE (Miyamoto and Kollman, 1992). Long-range electrostatic interactions were handled by particle mesh Ewald (PME) method (Darden et al., 1993).

171 **3. Results and discussion**

172 **3.1 Adsorption of a single type of polymers**

173 The final configurations of systems Mt-CT, Mt-PAM, and Mt-MF are displayed in Figure 2. 174 In Figure 2a, the adsorbed chitosan molecules formed a stable monolayer coating on each Mt 175 surface. In Figure 2b, only the chitosan molecules adsorbed on the upper surface are shown, 176 which adhered to the surface in a dispersed manner. Detailed examination of the adsorption sites 177 revealed a close correlation with the locations of the isomorphic substitutions (see more details in 178 SI, section S3). Interestingly, Willemsen et al., 2019 reported the anticorrelation between the 179 adsorption sites of phthalate esters with locations of isomorphic substitutions, which was caused 180 by the hydrophobic nature of phthalate esters and the resulting preference of adhering to 181 uncharged patches. Different from chitosan, PAM molecules as shown in Figure 2c and Figure

2d tended to form clusters among themselves. All the MF molecules remained dispersed in bulk
water without adsorbing to the clay mineral surfaces, as shown in Figure 2e and Figure 2f.

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186 The vertical distances (D), as a function of time, between the center of geometry (COG) of 187 Mt and the COG of one representative polymer molecules in each of the three systems are shown 188 in Figure 3. The results for the other polymer molecules in each system are shown in SI (Section 189 S3), which shared similar characteristics. All chitosan molecules moved towards Mt quickly, 190 with a fast decay in D in the first 2 ns, and then stayed on the surfaces without detaching again. 191 The adsorption of chitosan was driven by the strong electrostatic attraction between the 192 positively charged chitosan and the negatively charged Mt. Meanwhile, the electrostatic 193 repulsion among the chitosan molecules caused them to adsorb separately on the surface and 194 form a monolayer coating. Consistent with the snapshots shown in Figures 2e and 2f, MF 195 molecules were dispersed in the bulk without adsorption, and the value of D was above 2 nm for 196 most of the simulation time. The electrostatic repulsions between the negatively charged MF and 197 Mt as well as among MF molecules kept them apart from approaching each other. The 198 adsorption of PAM molecules was much slower and weaker, as D was below 1 nm at ~ 10 ns for 199 the first time and increased again. Throughout the simulation, D underwent large fluctuations, 200 corresponding to frequent exchange between adsorbed and desorbed states. Such unstable 201 adsorption was driven by the H-bonds between amides of PAM and bridging O in the basal 202 surfaces of Mt. In addition, PAM could adsorb on Mt in the form of either a single molecule or 203 an aggregate. More details on determining the aggregation state of PAM molecules are given in 204 SI (Section S3). For example, the representative PAM molecule shown in Figures 3 was

adsorbed as a single molecule (square symbols) at around 10 ns, 25 ns, and 60 ns, whereas
between 90 ns and 105 ns it was adsorbed as a molecule in an aggregate ("×" symbols).

207 Quantitative analysis on H-bonding provided further assessment on the characteristics of 208 PAM adsorption. The total numbers of H-bonds between PAM and Mt $(n_{\rm H})$ was calculated and then normalized with respect to the number of donor/acceptor pairs available (N_0 , more details in 209 210 SI, section S4). This ratio, $p = n_{\rm H}/N_0$, represented the percentage of available donor/acceptor 211 pairs that actually contributed to H-bonding. The same calculation was done for the H-bonding 212 between PAM molecules, and both results are shown in Figure 4. In the first 8 ns, p between 213 PAM and Mt remained zero while the number among PAM molecules quickly increased, 214 corresponding to the clustering of PAM molecules in the bulk solution. The value of p among 215 PAM molecules fluctuated steadily between 7% and 16% for the rest of the simulation time, 216 indicating the stability of the formed aggregates. Meanwhile, participating percentage of 217 donor/acceptor pairs between PAM and Mt remained less then 2%, and frequently reduced to 218 zero, suggesting reversible adsorption/desorption of the PAM molecules. The increases of p219 between PAM and Mt from 50 to 80 ns, and from 90 to 110 ns, were caused by PAM adsorption 220 in either individual or aggregated forms, and examples are shown in the snapshots in Figure 4.

221 **3.2 Adsorption of simultaneously added polymers**

The configurations of systems Mt-(CT-PAM) and Mt-(CT-MF) at different time are shown in Figure 5. Consider first system Mt-(CT-PAM) in Figure 5a-d. At t = 3 ns (Figure 5b), all the chitosan molecules became adsorbed on Mt while the PAM molecules were still in the bulk. Because chitosan molecules were initially placed on the right side of the simulation box, the adsorption occurred on the right side as well. That is, the chitosan molecules found the shortest path to form contact with Mt. As the simulation continued, the chitosan molecules started spreading over the entire Mt surface, forming a thin monolayer similar to what was seen in Figure 2a. Meanwhile, several PAM molecules started adhering to the adsorbed chitosan while some of them also adsorbed directly on Mt surrounded by chitosan molecules, as illustrated in Figure 5c. At the end of the simulation (Figure 5d), all the polymers were adsorbed, and the PAM molecules conformed much better to the surface than in Figure 2b where a bulky PAM cluster adhered to the surface with significant exposure to the solvent. Clearly, the presence of chitosan greatly enhanced the adsorption of PAM.

235 The behaviors in system Mt-(CT-MF) as depicted in Figure 5e-h were similar to those in 236 system Mt-(CT-PAM): the process also began by attachment and subsequent spreading of 237 chitosan, followed by the adsorption of MF via chitosan, and all the polymer molecules were 238 adsorbed at the end. The enhancement of MF adsorption was significant, considering that in 239 Figure 2c all MF molecules were dispersed in the bulk with no adsorption at all. Compared with 240 system Mt-(CT-PAM), there were a few differences in system Mt-(CT-MF). First of all, in the 241 initial stage of the simulation some chitosan could be "captured" by the negatively charged MF 242 in the bulk (see Figure 5f at t = 3 ns) and their adsorption to Mt was delayed (not until t = 30 ns 243 in Figure 5g where the clustered chitosan and MF adsorbed together). On the contrary, for 244 system Mt-(CT-PAM), there was little interaction between the positively charged chitosan and 245 neutral PAM before surface adsorption of chitosan. Secondly, while PAM molecules could 246 adsorb on Mt via two modes (directly and indirectly through chitosan bridges), all MF adsorption 247 occurred via binding with chitosan.

The different modes of adsorption could be further confirmed by the vertical distances (D)between the COG of polymer molecules and COG of Mt, as plotted in Figure 6, against simulation time. Results are only shown for one chitosan (labeled as CT_1), two PAM molecules

(labeled as PAM_1, PAM_2) in Mt-(CT-PAM), and one MF molecule (labeled as MF 1) in Mt-251 252 (CT-MF), while the rest are given in the SI (Section S3). Chitosan in both systems displayed the 253 same trend as CT 1: D reduced to about 0.5 nm after 2 ns and remained constant, corresponding 254 to fast, direct and stable adsorption. All MF in Mt-(CT-MF) also exhibited the same mode of 255 adsorption, as D reduced to 1.5 nm at \sim 25 ns in Figure 6 and was stable afterwards. This larger 256 steady-state value of D, compared to chitosan (0.5 nm), suggests that MF adsorbed on top of the 257 chitosan monolayer and could not move closer to Mt. The fast movement and stable coating of 258 chitosan on Mt were due to the strong electrostatic attraction between positively charged 259 chitosan and negatively charged Mt. Such attraction also existed between chitosan and negative 260 MF, while MF and Mt were mutually repulsive. These interactions determined the mode of 261 adsorption in Mt-(CT-MF), where chitosan molecules were sandwiched between Mt and MF 262 serving as bridges.

263 PAM in system Mt-(CT-PAM) displayed two modes of adsorption. The first was represented 264 by PAM_1, for which D maintained a value of ~ 0.5 nm after 25 ns, representing a stable 265 adsorption directly on the surface of Mt. Contrary to PAM_1, PAM_2 displayed a more dynamic 266 motion. At ~20 ns, it approached Mt with D staying at ~1.5 nm for about 10 ns. During this time, 267 PAM_2 adsorbed to Mt via a chitosan bridge, similar to MF molecules in Mt-(CT-MF). 268 Afterwards D increased rapidly corresponding to departure of PAM_2 from Mt. It adhered to Mt 269 again after 45 ns with D fluctuating between 0.5 nm and 1.5 nm. Such fluctuation represented the 270 switch between the two adsorption modes: directly on Mt and adsorption via chitosan.

Overall charge neutral PAM molecules interacted with chitosan via the formation of H-bonds between the hydroxyl groups of chitosan and amides of PAM. The ability of PAM to form Hbonds with chitosan as well as with Mt led to both direct adsorption and indirect adsorption via 274 chitosan bridges. Similar to Figure 4, the percentages of available donor/acceptor pairs that 275 contributed to H-bonding, p, were calculated for system Mt-(CT-PAM) and shown in Figure 7. 276 From 0 to 10 ns, p value was highest among PAM molecules, slightly above zero between 277 chitosan and PAM, and zero between PAM and Mt. Therefore, similar to system Mt-PAM, 278 during the initial stage clustering of PAM was favored and no immediate adsorption onto Mt was 279 observed. Meanwhile, PAM started interaction with the adsorbed chitosan layer. From 10 to 25 280 ns, p among PAM molecules reduced to one half of the original value while p between PAM and 281 chitosan increased. H-bonds between PAM and Mt also began to form. The results suggested 282 conversion of some PAM-PAM H-bonds to PAM-chitosan and PAM-Mt H-bonds. PAM 283 aggregates previously formed in the bulk broke into smaller ones to facilitate the adsorption. 284 From 25 to 80 ns, the value of p among PAM first increased and then decreased, indicating the 285 reversibility of PAM aggregation. The value of p between PAM and chitosan fluctuated at a 286 lower level than the curve between PAM and Mt which increased significantly, suggesting that 287 direct adsorption was more favorable compared to indirect adsorption via chitosan bridges.

288 Comparing the data in the last 20 ns of Figure 4 and Figure 7, there was a 50% reduction in 289 the p value among PAM molecules, from $\sim 12\%$ in Figure 4 to $\sim 6\%$ in Figure 7. Meanwhile, 290 there was a 20-fold increase in p between PAM and Mt, from $\sim 0.2\%$ in Figure 4 to $\sim 4\%$ in 291 Figure 7. Chitosan not only served as bridges for the indirect adsorption of PAM, but also 292 suppressed the aggregation of PAM molecules and greatly promoted the direct adsorption of 293 PAM on Mt. As PAM molecules were brought closer to Mt via the chitosan bridges, their short-294 ranged attractions with Mt were enabled, which drove them to continue moving towards Mt until 295 direct H-bonding was established.

296 **3.3 Adsorption of sequentially added polymers**

The configurations of systems (Mt-CT)-PAM and (Mt-PAM)-CT at different time are shown in Figure 8. Consider system (Mt-CT)-PAM as shown in Figure 8a-d, the adsorption showed no difference compared with system Mt-(CT-PAM) shown in Figure 5a-d. Since chitosan moved much faster towards Mt than PAM, adding chitosan first or adding them simultaneously did not affect the adsorption. Similar observations were found for system (Mt-CT)-MF (more details in SI Section S3).

303 What is more interesting is system (Mt-PAM)-CT where chitosan molecules were added to 304 Mt-PAM with an adsorbed PAM cluster on Mt. Even though in this case PAM adsorption had 305 already occurred before the introduction of chitosan, the stronger long-range electrostatic 306 attraction between chitosan and Mt still drove the formation of a chitosan monolayer on Mt at the 307 end of the simulation (Figure 8h), which was similar to systems (Mt-CT)-PAM and Mt-(CT-308 PAM). During this process, however, the PAM aggregate was never "released" from the surface 309 into the bulk. Instead, chitosan maintained their interactions with PAM while forming intimate 310 contact with Mt (see Figure 8f and 8g). Consequently, the PAM aggregate was anchored to the 311 vicinity of the Mt, gradually allowed chitosan to insert between PAM and Mt, and finally 312 adsorbed on the formed chitosan monolayer. PAM initially adsorbed as individual molecules 313 (Figure 8e) remained adsorbed directly, and were not disrupted by chitosan.

314 **3.4 Implications**

As discussed in Section 3.1, chitosan molecules adsorbed quickly and formed a monolayer coating on Mt. Though the two clay mineral surfaces could not be brought together by chitosan due to the application of periodic boundary conditions, the ability of chitosan to bridge two bare clay mineral surfaces were proved in other contexts such as fabrication of nanocomposites 319 (Razmimanesh et al., 2015; Wang et al., 2015). In industrial tailings, a clay mineral particle 320 coated by chitosan could have neutral or even reversed zeta potential (Lu et al., 2016; Zvulunov 321 et al., 2019). As such, it could attract another bare clay mineral surface, flocculating them 322 together. However, if both particles were coated by chitosan and possessed positive surface 323 charges, the electrostatic repulsion between them would hinder their aggregation and limit the 324 growth of flocs, as illustrated in Figure 9a. As a result, it can be hypothesized that the flocs will 325 settle slowly due to their small size. Since the adsorption of chitosan is highly stable, clay 326 mineral particles captured in the flocs are less likely to be released back to the supernatant, 327 resulting in high clarity. This result is consistent with the experimental observation that the ISR 328 of chitosan or other cationic polymers treated wastewater/tailings showed a low ISR but high 329 clarity of supernatant (Lu et al., 2016; Zhang et al., 2017).

330 PAM molecules by themselves could adsorb on Mt, but the adsorption tended to be in the 331 form of an aggregate and was not stable. It could then be inferred that in macroscopic systems, 332 PAM alone might induce multi-layer adsorption with reversible desorption. Two clay mineral 333 particles with adsorbed PAM can further aggregate since the neutral PAM does not introduce 334 repulsion between them while allowing for H-bonding among the adsorbed PAM molecules. 335 This interaction mechanism can lead to large flocs that settle quickly in clay mineral dispersion, 336 consistent with experimental observation (Wang et al., 2010; Zhang et al., 2017). On the other 337 hand, the unstable adsorption can allow a large amount of clay minerals to be released back to 338 the bulk solution, as illustrated in Figure 9b. Consequently, the clarity of clay mineral dispersion 339 is expected to be low, which was observed by various experiments (Liu et al., 2016; Lu et al., 340 2016).

341 In experiments, when two types of flocculants were added simultaneously, the floc size 342 displayed a two-stage growth (Wang et al., 2016) which was consistent with the two-step 343 adsorption as shown in the system Mt-(CT-PAM). Regardless of the adding sequence, our 344 simulations showed that the cationic polymers formed the most intimate adsorption on the clay 345 mineral followed by the neutral polymers. The cationic coating stabilized the adsorption of 346 neutral flocculants, while the outer layer of neutral polymers promoted the local attractive forces 347 such as H-bonding, bridging them together to form larger aggregates, as illustrated in Figure 9c. 348 Compared with the application of a single type of flocculants, such synergetic effect is expected 349 to not only increase the ISR by forming larger clay mineral aggregates but also enhance the 350 clarity of the supernatant by reducing the number of bare clay mineral particles in it. The result is 351 consistent with previous reports where the addition of two types of polymers showed both high 352 ISR and clarity (Lu et al., 2016; Zhang et al., 2017). The synergetic effect was also present when 353 a cationic and an anionic flocculants were added, as our simulations showed a similar two-stage 354 adsorption process for chitosan and MF. The bonding strength between chitosan and MF was in 355 fact larger than that between chitosan and PAM. Although if two particles both possessed an 356 outer MF layer they would repel, in macroscopic systems it is possible to have multi-layer 357 adsorption in the form of CT-MF-CT, CT-MF-CT-MF, etc., which would allow particles with 358 opposite apparent surface charges to aggregate. To the best of our knowledge, this is the first 359 theoretical work that investigated the adsorption of polymer flocculants from molecular level and 360 accounted for the different flocculation behaviors. The results provided molecular evidence that 361 applying polymers with different charge properties would promote the flocculation and the 362 applying sequences had limited influence. Our findings may also be of value to other applications, such as capture and storage of organic resources, as well as the design of claypolymer nanocomposite.

365 **3.5 Limitations**

366 In the present work, the basal surface of Mt was chosen as a representative clay mineral 367 surface, while actual clay particles also contain edge surfaces and interlayers. When it comes to 368 polymer flocculation, the likelihood of adsorption of polymer flocculants in the interlayers is 369 expected to be low. This is because the interlayer spacing is on the order of nanometers (Suter et 370 al., 2011), which is difficult to accommodate polymer flocculants whose molecular mass is usually in the range of 10^{6} - 10^{7} g/mol (Alamgir et al., 2012; Lu et al., 2015). The relative 371 372 importance of basal and edge surfaces is an interesting but complex problem. For example, 373 characterization results from XRD and SAXS showed preferential interactions of polymers with 374 basal surface (Li et al., 2019; Willemsen et al., 2019), with edge surface (Song and Sandí, 2001), 375 or comparable affinities to both surfaces (Ngnie et al., 2018). The distinct results were likely 376 caused by the different charges and terminating groups on the basal and edge surfaces, different 377 polymer properties, as well as different solution chemistries (e.g. pH, ions). Since the focus of 378 this study was to investigate the synergy of polymer adsorption arising from interactions such as 379 electrostatics and hydrogen bonding, the use of a single-layer basal surface served the purpose. 380 The mechanism discovered in this work should also be applicable to an edge surface if its 381 interactions with the polymer flocculants were governed by the same types of intermolecular 382 forces, although confirmation of this hypothesis would require addition investigations.

In addition to Mt, other types of clay minerals also exist in industrial tailings (Theng, 2012), for example, kaolinite and illite. Kaolinite is a 1:1 type planar phyllosilicate with one tetrahedral siloxane sheet and one octahedral alumina sheet (Zeitler et al., 2017). Under the same pH (= 8)

386 condition as considered in this work, it tends to have a neutral basal surface as atom substitutions 387 are less likely to happen. However, its edge surface can carry negative charges as the physically 388 bonded water would get deprotonated under the pH of 8 (Zeitler et al., 2017). Thus, adsorption of 389 cationic chitosan on kaolinite is expected to occur but mostly on the edge surface. PAM and MF 390 may display more adsorption on kaolinite than on Mt because of the exposed hydroxyl groups on 391 the alumina basal and edge surfaces of kaolinite, which promote hydrogen bonding. The 392 synergetic effect demonstrated in this work is expected to be present for kaolinite, since all three 393 types of polymers are expected to adsorb on it and the interactions driving adsorption are similar. 394 Illite is a dioctahedral mica-like mineral (Theng, 2012) which has similar molecular structure 395 with Mt. Its basal surface carries permanent negative charges due to atom substitutions while the 396 charges on the edge surface are pH dependent. Therefore, the synergetic polymer adsorption 397 observed for Mt is expected to be similar for illite. Slight difference might come from the fact 398 that atom substitution in Mt usually occurs only in the octahedral sheet while it can occur in both 399 tetrahedral and octahedral sheets in illite (Theng, 2012; Hao et al., 2018). Because of this, the 400 charge density on the basal surface of illite can be higher which may lead to enhanced adsorption 401 of cationic chitosan.

Due to the size limitation in MD simulations, it was not possible to reproduce the bulk concentration of polymer molecules in experiments. In fact, typical bulk concentrations used in experiments (Lu et al., 2016; Nittala et al., 2019) would correspond to at most one polymer molecule in the simulation cell, which would not have allowed us to investigate the interactions among the polymer molecules, a key aspect of this work. On the other hand, because this work focused on polymer adsorption, the higher concentration used mimicked the fact that the concentration of the polymers near the clay surface should be much higher than that in the bulk 409 solution. Different loadings of polymers can affect polymer adsorption. As some of the 410 experimental works suggested (Lu et al., 2016; Zhang et al., 2017), there could be an optimal 411 dosage of polymers for which the extent of flocculation was maximized. The mechanism behind 412 the observations is unclear and is a very interesting question to address in the future.

413 Furthermore, the ions included in the simulation systems were monovalent (Na⁺). Multiple types of ions can exist in waste-water, such as Na⁺, Ca²⁺, Mg²⁺, Cl⁻, HCO₃⁻, SO₄²⁻, NH₄⁺ (Allen, 414 2008). There was evidence that divalent ions, such as Ca^{2+} , could form more stable bridges 415 416 between clay mineral and anionic polymers which enhanced their adsorption (Li et al., 2005). In 417 fact, in our simulations MF by themselves did not exhibit any adsorption, which appeared 418 contradictory to earlier experimental studies where MF and other anionic polymers showed 419 flocculating abilities. The presence of divalent ions in the experiments (Li et al., 2005; Lu et al., 420 2016) could be a reason for the discrepancy; another contributing factor could be the hydroxyl 421 groups on the edge surfaces of Mt and basal/edge surfaces of kaolinite (Bish, 1993) in the 422 experiments. It is worth mentioning that some ions could also compete with polymers in the 423 adsorption and even formed surface complexes on Mt (Bourg and Sposito, 2011) that hindered 424 polymer adsorption. Effect of different ions are being investigated and will be reported elsewhere. Finally, the sizes of clay mineral particles (~1 μ m) and polymers (molar mass 10^{6~7} g/mol) in 425 426 practical settling (Wang et al., 2014) are much larger than the molecular models studied here. As 427 with any molecular simulations, the length and time scales of our simulations are limited by the 428 state-of-the-art computing capabilities. On the other hand, the atomistic simulations revealed the 429 intermolecular interactions that give rise to synergetic adsorption, which could not be obtained 430 with lab-scale resolution. The good agreement found here between mechanistic insights from

431 simulations and synergetic flocculation observed in experiments supports the complementarity of432 the two approaches, which have great potential for other advanced studies in clay science.

433 **4. Conclusion**

434 Molecular dynamics simulations revealed different characteristics for the adsorptions of 435 cationic (chitosan), neutral (PAM), and anionic (MF) polymers on the negatively charged basal 436 surface of montmorillonite. When a single type of polymers was added, chitosan showed fast and 437 stable monolayer adsorption, driven by electrostatic attraction; PAM displayed weak adsorption 438 driven by hydrogen bonding, mainly in an aggregated form; and MF showed no adsorption. 439 When chitosan was added to PAM or MF, their adsorption was significantly enhanced. Chitosan 440 not only served as bridges between PAM (or MF) and montmorillonite, but also promoted direct 441 adsorption of PAM on montmorillonite. The findings correlated well with the synergetic effect 442 observed in flocculation experiments, and have direct implications on the selection of flocculants 443 as well as the design of flocculation procedure.

444 **Supporting Information**

- 445 S1. Validation of force fields for clay minerals and organic molecules
- 446 S2. Partial charge calculation
- 447 S3. Additional results for the simulated systems
- 448 **S4. Calculation of hydrogen bonds**

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454 **Reference**

- 455 Abraham, M. J., Murtola, T., Schulz, R., Páll, S., Smith, J. C., Hess, B., Lindahl, E., 2015.
- 456 GROMACS: High performance molecular simulations through multi-level parallelism from
- 457
 laptops
 to
 supercomputers.
 SoftwareX
 1–2,
 19–25.

 458
 https://doi.org/10.1016/j.softx.2015.06.001
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- 459 Alamgir, A., Harbottle, D., Masliyah, J., Xu, Z., 2012. Al-PAM assisted filtration system for
- 460 abatement of mature fine tailings. Chem. Eng. Sci. 80, 91–99.
 461 https://doi.org/10.1016/j.ces.2012.06.010
- Allen, E.W., 2008. Process water treatment in Canada's oil sands industry: I. Target pollutants
 and treatment objectives. J. Environ. Eng. Sci. 7, 123–138. https://doi.org/10.1139/S07-038
- 464 Azevedo Rios Silva, F., Araújo Sales, M.J., Ghoul, M., Chebil, L., Duarte Ramos Matos, G.,
- 465 Maia, E.R., 2018. Molecular dynamics simulations of montmorillonite reinforcing amylose
- 466 plasticized by Brazilian Cerrado oils: Polymer-clay nanocomposite. MRS Commun. 8, 266–
- 467 274. https://doi.org/10.1557/mrc.2018.41
- Bakhshi, N.N., Gillies, R.G., Khare, P., 1975. Treatment of tar sands tailings with fly ash.
 Environ. Sci. Technol. https://doi.org/10.1021/es60102a013
- 470 Berendsen, H.J.C., van der Spoel, D., van Drunen, R., 1995. GROMACS: a message-passing
- 471 parallel molecular dynamics implementation. Comput. Phys. Commun. 91, 43–56.
 472 https://doi.org/10.1016/0010-4655(95)00042-E
- 473 Bish, D.L., 1993. Rietveld refinement of the kaolinite structure at 1.5 K. Clays Clay Miner. 41,
- 474 738–744. https://doi.org/10.1346/CCMN.1993.0410613
 - 21

Bourg, I.C., Sposito, G., 2011. Molecular dynamics simulations of the electrical double layer on
 smectite surfaces contacting concentrated mixed electrolyte (NaCl-CaCl2) solutions. J.

477 Colloid Interface Sci. 360, 701–715. https://doi.org/10.1016/j.jcis.2011.04.063

- 478 Bourg, I.C., Sposito, G., Bourg, A.C.M., 2007. Modeling the acid-base surface chemistry of
- 479 montmorillonite. J. Colloid Interface Sci. 312, 297–310.
 480 https://doi.org/10.1016/j.jcis.2007.03.062
- Chen, J., Min, F., Liu, L., Cai, C., 2020. Systematic exploration of the interactions between Fedoped kaolinite and coal based on DFT calculations. Fuel 266, 117082.
 https://doi.org/10.1016/j.fuel.2020.117082
- Cheng, H., Zhang, S., Liu, Q., Li, X., Frost, R.L., 2015. The molecular structure of kaolinitepotassium acetate intercalation complexes: A combined experimental and molecular
 dynamic simulation study. Appl. Clay Sci. 116–117, 273–280.
 https://doi.org/10.1016/j.clay.2015.04.008
- 488 Cygan, R.T., Greathouse, J.A., Heinz, H., Kalinichev, A.G., 2009. Molecular models and
 489 simulations of layered materials. J. Mater. Chem. 19, 2470.
 490 https://doi.org/10.1039/b819076c
- 491 Cygan, R.T., Liang, J.J., Kalinichev, A.G., 2004. Molecular models of hydroxide, oxyhydroxide,
 492 and clay phases and the development of a general force field. J. Phys. Chem. B 108, 1255–
 493 1266. https://doi.org/10.1021/jp0363287
- 494 Darden, T., York, D., Pedersen, L., 1993. Particle mesh Ewald: An N · log(N) method for Ewald
 495 sums in large systems. J. Chem. Phys 98, 10089–10092. https://doi.org/10.1063/1.464397
- 496 Downs, R.T., Hall-Wallace, M., 2003. The American Mineralogist crystal structure database.
- 497 Am. Mineral. 88, 247–250.

- 498 Fernandes, D., Conway, W., Wang, X., Burns, R., Lawrance, G., Maeder, M., Puxty, G., 2012.
- 499 Protonation constants and thermodynamic properties of amines for post combustion capture
- 500 of CO 2. J. Chem. Thermodyn. 51, 97–102. https://doi.org/10.1016/j.jct.2012.02.031
- 501 Fraenkel, G., Franconi, C., 1960. Protonation of Amides. J. Am. Chem. Soc. 82, 4478–4483.
 502 https://doi.org/10.1021/ja01502a010
- 503 Gorakhki, M.H., Bareither, C.A., 2015. Salinity effects on sedimentation behavior of kaolin, 504 bentonite, soda ash mine tailings. Appl. Clay Sci. 114. 593-602. and 505 https://doi.org/10.1016/j.clay.2015.07.018
- Grant, H., Mctigue, P., Ward, D., 2010. The basicities of aliphatic amides. Aust. J. Chem. 36,
 2211. https://doi.org/10.1071/ch9832211
- Hao, Y., Yuan, L., Li, P., Zhao, W., Li, D., Lu, D., 2018. Molecular Simulations of Methane
 Adsorption Behavior in Illite Nanopores Considering Basal and Edge Surfaces. Energy
 Fuels 32. https://doi.org/10.1021/acs.energyfuels.8b00070
- 511 Hess, B., Bekker, H., Berendsen, H.J.C., Fraaije, J.G.E.M., 1997. LINCS: A linear constraint
- 512 solver for molecular simulations. J. Comp. Chem 18, 1463–1472.
 513 https://doi.org/10.1002/(SICI)1096-987X(199709)18:12<1463::AID-JCC4>3.0.CO;2-H
- 514 Hockney, R.W., Goel, S.P., Eastwood, J., 1974. Quiet high resolution computer models of a
- 515 plasma. J. Comp. Phys 14, 148–158. https://doi.org/10.1016/0021-9991(74)90010-2
- 516 Hoover, W.G., 1985. Canonical dynamics: equilibrium phase-space distributions. Phys. Rev. A
- 517 31, 1695–1697. https://doi.org/10.1103/PhysRevA.31.1695
- Israelachvili, J.N., 2011. Intermolecular and Surface Forces, third ed. Elsevier Inc. Waltham
 https://doi.org/10.1016/C2011-0-05119-0
- 520 Ji, Y., Lu, Q., Liu, Q., Zeng, H., 2013. Effect of solution salinity on settling of mineral tailings

- by polymer flocculants. Colloids Surfaces A Physicochem. Eng. Asp. 430, 29–38.
 https://doi.org/10.1016/j.colsurfa.2013.04.006
- Kang, X., Xia, Z., Chen, R., Sun, H., Yang, W., 2019. Effects of inorganic ions, organic
 polymers, and fly ashes on the sedimentation characteristics of kaolinite suspensions. Appl.
- 525 Clay Sci. 181, 105220. https://doi.org/10.1016/j.clay.2019.105220
- Kaniki, A.T., Tumba, K., 2019. Management of mineral processing tailings and metallurgical
 slags of the Congolese copperbelt: Environmental stakes and perspectives. J. Clean. Prod.
 210, 1406–1413. https://doi.org/10.1016/j.jclepro.2018.11.131
- 529 Kim, S., Chen, J., Cheng, T., Gindulyte, A., He, J., He, S., Li, Q., Shoemaker, B.A., Thiessen,
- P.A., Yu, B., Zaslavsky, L., Zhang, J., Bolton, E.E., 2019. PubChem 2019 update: Improved
 access to chemical data. Nucleic Acids Res. 47, D1102–D1109.
 https://doi.org/10.1093/nar/gky1033
- Li, H., Long, J., Xu, Z., Masliyah, J.H., 2005. Synergetic Role of Polymer Flocculant in LowTemperature Bitumen Extraction and Tailings Treatment. Energy Fuels 19, 936–943.
 https://doi.org/10.1021/ef049744e
- Li, P., Khan, M.A., Xia, M., Lei, W., Zhu, S., Wang, F., 2019. Efficient preparation and
 molecular dynamic (MD) simulations of Gemini surfactant modified layered
 montmorillonite to potentially remove emerging organic contaminants from wastewater.
 Ceram. Int. 45, 10782–10791. https://doi.org/10.1016/j.ceramint.2019.02.152
- Liu, Y., Lv, C., Ding, J., Qian, P., Zhang, X., Yu, Y., Ye, S., Chen, Y., 2016. The use of the
 organic-inorganic hybrid polymer Al(OH)3-polyacrylamide to flocculate particles in the
 cyanide tailing suspensions. Miner. Eng. 89, 108–117.
 https://doi.org/10.1016/j.mineng.2016.01.018

- Long, J., Li, H., Xu, Z., Masliyah, J.H., 2006. Role of colloidal interactions in oil sand tailings
 treatment. AIChE J. 52, 371–383. https://doi.org/10.1002/aic.10603
- 546 Lu, H., Wang, Y., Li, L., Kotsuchibashi, Y., Narain, R., Zeng, H., 2015. Temperature- and pH-
- 547 Responsive Benzoboroxole-Based Polymers for Flocculation and Enhanced Dewatering of
- 548 Fine Particle Suspensions. ACS Appl. Mater. Interfaces 7, 27176–27187.
 549 https://doi.org/10.1021/acsami.5b09874
- Lu, H., Xiang, L., Cui, X., Liu, J., Wang, Y., Narain, R., Zeng, H., 2016. Molecular Weight
 Dependence of Synthetic Glycopolymers on Flocculation and Dewatering of Fine Particles.
 Langmuir 32, 11615–11622. https://doi.org/10.1021/acs.langmuir.6b03072
- Lu, Q., Yan, B., Xie, L., Huang, J., Liu, Y., Zeng, H., 2016. A two-step flocculation process on
 oil sands tailings treatment using oppositely charged polymer flocculants. Sci. Total
 Environ. 565, 369–375. https://doi.org/10.1016/j.scitotenv.2016.04.192
- 556 Frisch, M. J., Trucks, G. W., Schlegel, H. B., Scuseria, G. E., Robb, M. A., Cheeseman, J. R.,
- 557 Scalmani, G., Barone, V., Petersson, G. A., Nakatsuji, H., Li, X., Caricato, M., Marenich, A.
- 558 V., Bloino, J., Janesko, B. G., Gomperts, R., Mennucci, B., Hratchian, H. P., Ortiz, J. V.,
- 559 Izmaylov, A. F., Sonnenberg, J. L., Williams-Young, D., Ding, F., Lipparini, F., Egidi, F.,
- 560 Goings, J., Peng, B., Petrone, A., Henderson, T., Ranasinghe, D., Zakrzewski, V. G., Gao, J.,
- 561 Rega, N., Zheng, G., Liang, W., Hada, M., Ehara, M., Toyota, K., Fukuda, R., Hasegawa, J.,
- 562 Ishida, M., Nakajima, T., Honda, Y., Kitao, O., Nakai, H., Vreven, T., Throssell, K.,
- 563 Montgomery, J. A., Jr., Peralta, J. E., Ogliaro, F., Bearpark, M. J., Heyd, J. J., Brothers, E.
- 564 N., Kudin, K. N., Staroverov, V. N., Keith, T. A., Kobayashi, R., Normand, J.,
- 565 Raghavachari, K., Rendell, A. P., Burant, J. C., Iyengar, S. S., Tomasi, J., Cossi, M., Millam,
- 566 J. M., Klene, M., Adamo, C., Cammi, R., Ochterski, J. W., Martin, R. L., Morokuma, K.,

- 567 Farkas, O., Foresman, J. B., Fox, D. J., 2016. Gaussian 16 Revision B. 01. Wallingford CT.
- 568 Miyamoto, S., Kollman, P.A., 1992. SETTLE: An analytical version of the SHAKE and
- 569 RATTLE algorithms for rigid water models. J. Comp. Chem 13, 952–962.
 570 https://doi.org/10.1002/jcc.540130805
- Molatlhegi, O., Alagha, L., 2017. Adsorption characteristics of chitosan grafted copolymer on
 kaolin. Appl. Clay Sci. 150, 342–353. https://doi.org/10.1016/j.clay.2017.09.032
- 573 Moreno, L., Neretnieks, I., 2006. Long-term environmental impact of tailings deposits.
 574 Hydrometallurgy 83, 176–183. https://doi.org/10.1016/j.hydromet.2006.03.052
- 575 Nasim, T., Pal, A., Bandyopadhyay, A., 2018. Flocculation of aqueous kaolin suspension using a
 576 biodegradable flocculant system of poly (vinyl alcohol)-Acacia nilotica gum blends. Appl.
 577 Clay Sci. 152, 83–92. https://doi.org/10.1016/j.clay.2017.10.035
- Ngnie, G., Baitan, D., Dedzo, G.K., Detellier, C., 2018. Sedimentation of fine particles of
 kaolinite and polymer-coated kaolinite in cyclohexane: Implications for fines removal from
 extracted bitumen in non-aqueous processes. Fuel 234, 218–224.
 https://doi.org/10.1016/j.fuel.2018.07.032
- 582 Nittala, A.K., Gumfekar, S.P., Soares, J.B.P., 2019. Multifunctional CO 2 -switchable polymers
- for the flocculation of oil sands tailings. J. Appl. Polym. Sci. 136, 1–9.
 https://doi.org/10.1002/app.47578
- Nosé, S., 1984. A molecular dynamics method for simulations in the canonical ensemble. Mol.
 Phys 52, 255–268. https://doi.org/10.1080/00268978400101201
- 587 Omotoso, O.E., Mikula, R.J., 2004. High surface areas caused by smectitic interstratification of
 588 kaolinite and illite in Athabasca oil sands. Appl. Clay Sci. 25, 37–47.
 589 https://doi.org/10.1016/j.clay.2003.08.002

- 590 Oostenbrink, C., Villa, A., Mark, A.E., Van Gunsteren, W.F., 2004. A biomolecular force field
 591 based on the free enthalpy of hydration and solvation: The GROMOS force-field parameter
 592 sets 53A5 and 53A6. J. Comput. Chem. 25, 1656–1676. https://doi.org/10.1002/jcc.20090
- 593 Parrinello, M., Rahman, A., 1981. Polymorphic transitions in single crystals: A new molecular
 594 dynamics method. J. Appl. Phys 52, 7182–7190. https://doi.org/10.1063/1.328693
- 595 Pouvreau, M., Greathouse, J.A., Cygan, R.T., Kalinichev, A.G., 2017. Structure of Hydrated
 596 Gibbsite and Brucite Edge Surfaces: DFT Results and Further Development of the ClayFF
- 597 Classical Force Field with Metal-O-H Angle Bending Terms. J. Phys. Chem. C 121, 14757–
- 598 14771. https://doi.org/10.1021/acs.jpcc.7b05362
- Pronk, S., Páll, S., Schulz, R., Larsson, P., Bjelkmar, P., Apostolov, R., Shirts, M. R., S., J. C.,
 Kasson, P. M., van der Spoel, D., Hess, B., Lindahl, E., 2013. GROMACS 4.5: a
 highthroughput and highly parallel open source molecular simulation toolkit.
 Bioinformatics 29, 845–854. https://doi.org/10.1093/bioinformatics/btt055
- Razmimanesh, F., Amjad-Iranagh, S., Modarress, H., 2015. Molecular dynamics simulation
 study of chitosan and gemcitabine as a drug delivery system. J. Mol. Model. 21.
 https://doi.org/10.1007/s00894-015-2705-2
- Ren, B., Min, F., Liu, L., Chen, J., Liu, C., Lv, K., 2020. Adsorption of different PAM structural
 units on kaolinite (0 0 1) surface: Density functional theory study. Appl. Surf. Sci. 504.
 https://doi.org/10.1016/j.apsusc.2019.144324
- 609 Robert, T., Mercer, S.M., Clark, T.J., Mariampillai, B.E., Champagne, P., Cunningham, M.F.,
- 610 Jessop, P.G., 2012. Nitrogen-containing polymers as potent ionogens for aqueous solutions
- of switchable ionic strength: application to separation of organic liquids and clay particles
- 612 from water. Green Chem. 14, 3053. https://doi.org/10.1039/c2gc36074h

- Schüttelkopf, A.W., Van Aalten, D.M.F., 2004. PRODRG: A tool for high-throughput
 crystallography of protein-ligand complexes. Acta Crystallogr. Sect. D Biol. Crystallogr. 60,
 1355–1363. https://doi.org/10.1107/S0907444904011679
- 616 Shaikh, S.M.R., Nasser, M.S., Magzoub, M., Benamor, A., Hussein, I.A., El-Naas, M.H.,
- 617 Qiblawey, H., 2018. Effect of electrolytes on electrokinetics and flocculation behavior of
 618 bentonite-polyacrylamide dispersions. Appl. Clay Sci. 158, 46–54.
 619 https://doi.org/10.1016/j.clay.2018.03.017
- Song, K., Sandí, G., 2001. Characterization of montmorillonite surfaces after modification by
 organosilane. Clays Clay Miner. 49, 119–125.
 https://doi.org/10.1346/CCMN.2001.0490202
- Suter, J.L., Coveney, P. V., Anderson, R.L., Greenwell, H.C., Cliffe, S., 2011. Rule based design
 of clay-swelling inhibitors. Energy Environ. Sci. 4, 4572–4586.
 https://doi.org/10.1039/c1ee01280k
- Theng, B.K.G., 2012. Formation and properties of clay-polymer complexes. Elsevier,
 Developmnets of Clay Science, Volume 4, Amsterdam https://doi.org/10.1016/B978075068620-4.50036-8
- Viani, A., Gualtieri, A.F., Artioli, G., 2002. The nature of disorder in montmorillonite by
 simulation of X-ray powder patterns. Am. Mineral. 87, 966–975.
- Wang, C., Han, C., Lin, Z., Masliyah, J., Liu, Q., Xu, Z., 2016. Role of Preconditioning Cationic
- Zetag Flocculant in Enhancing Mature Fine Tailings Flocculation. Energy and Fuels 30,
 5223–5231. https://doi.org/10.1021/acs.energyfuels.6b00108
- Wang, C., Harbottle, D., Liu, Q., Xu, Z., 2014. Current state of fine mineral tailings treatment: A
 critical review on theory and practice. Miner. Eng. 58, 113–131.

- 636 https://doi.org/10.1016/j.mineng.2014.01.018
- Wang, Q.Z., Chen, X.G., Liu, N., Wang, S.X., Liu, C.S., Meng, X.H., Liu, C.G., 2006.
 Protonation constants of chitosan with different molecular weight and degree of
 deacetylation. Carbohydr. Polym. 65, 194–201.
 https://doi.org/10.1016/j.carbpol.2006.01.001
- Wang, X.W., Feng, X., Xu, Z., Masliyah, J.H., 2010. Polymer aids for settling and filtration of
 oil sands tailings. Can. J. Chem. Eng. 88, 403–410. https://doi.org/10.1002/cjce.20283
- 643 Wang, Y., Wohlert, J., Bergenstråhle-Wohlert, M., Tu, Y., Ågren, H., 2015. Molecular
- mechanisms for the adhesion of chitin and chitosan to montmorillonite clay. RSC Adv. 5,
 54580–54588. https://doi.org/10.1039/c5ra06424d
- Watson, A.H., Corser, P.G., Garces Pardo, E.E., Lopez Christian, T.E., Vandekeybus, J., 2010. A
 comparison of alternative tailings disposal methods—the promises and realities. In: R
 Jewell & AB Fourie (eds) Proceedings of the First International Seminar on the Reduction
 of Risk in the Management of Tailings and Mine Waste, Australian Centre for
 Geomechanics, Perth, pp. 499-514. https://doi.org/10.36487/ACG_rep/1008_41_Watson
- 651 Wei, Q., Wang, Y., Wang, S., Zhang, Y., Chen, X., 2017. Investigating the properties and interaction mechanism of nano-silica in polyvinyl alcohol/polyacrylamide blends at an 652 653 level. J. Mech. Behav. Biomed. 75, 529-537. atomic Mater. 654 https://doi.org/10.1016/j.jmbbm.2017.08.027
- 655 Willemsen, J.A.R., Myneni, S.C.B., Bourg, I.C., 2019. Molecular Dynamics Simulations of the
- Adsorption of Phthalate Esters on Smectite Clay Surfaces. J. Phys. Chem. C 123, 13624–

657 13636. https://doi.org/10.1021/acs.jpcc.9b01864

458 Yang, Z., Shang, Y., Lu, Y., Chen, Y., Huang, X., Chen, A., Jiang, Y., Gu, W., Qian, X., Yang,

- H., Cheng, R., 2011. Flocculation properties of biodegradable amphoteric chitosan-based
 flocculants. Chem. Eng. J. 172, 287–295. https://doi.org/10.1016/j.cej.2011.05.106
- 661 Zeitler, T.R., Greathouse, J.A., Cygan, R.T., Fredrich, J.T., Jerauld, G.R., 2017. Molecular
- Dynamics Simulation of Resin Adsorption at Kaolinite Edge Sites: Effect of Surface
 Deprotonation on Interfacial Structure. J. Phys. Chem. C 121, 22787–22796.
 https://doi.org/10.1021/acs.jpcc.7b06688
- Zhang, D., Thundat, T., Narain, R., 2017. Flocculation and dewatering of mature fine tailings
 using temperature-responsive cationic polymers. Langmuir 33, 5900–5909.
 https://doi.org/10.1021/acs.langmuir.7b01160
- Zvulunov, Y., Ben-Barak-Zelas, Z., Fishman, A., Radian, A., 2019. A self-regenerating claypolymer-bacteria composite for formaldehyde removal from water. Chem. Eng. J. 374,
 1275–1285. https://doi.org/10.1016/j.cej.2019.06.017

system name	polymers	number of Na ⁺	simulated time
Mt-H ₂ O	0	120	2.5 ns
Mt-CT	chitosan (3×2×2)	0	7 ns
Mt-PAM	PAM (3×2×2)	120	110 ns
Mt-MF	MF (3×2×2)	156	50 ns
Mt-(CT-PAM)	chitosan $(3 \times 1 \times 2)$ + PAM $(3 \times 1 \times 2)$	60	80 ns
Mt-(CT-MF)	chitosan $(3 \times 1 \times 2) + MF (3 \times 1 \times 2)$	78	80 ns
(Mt-CT)-MF	Mt-CT + MF ($3 \times 2 \times 2$)	$0 \rightarrow 36$	80 ns
(Mt-CT)-PAM	Mt-CT + PAM $(3 \times 2 \times 2)$	$0 \rightarrow 0$	80 ns
(Mt-PAM)-CT	Mt-PAM + chitosan ($3 \times 2 \times 2$)	$120 \rightarrow 0$	80 ns

Table 1. Details of Simulated Syster	ms
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Figure 1. Initial configuration of system Mt-PAM (a), chitosan (b), MF 1011 (c), and PAM (d). In (a), the Mt supercell is represented by ball and stick. PAM is in licorice representation. Water as solvent is in line representation.



Figure 2. Snapshots of the final configuration of systems Mt-CT (a, b), Mt-PAM (c, d), and Mt-MF (e, f). Side view (a, c, e) and top view (b, d, f) are shown for each system. In top view, the alumina octahedron in Mt is not displayed, and water as well as ions are removed for clarity.



Figure 3. Vertical distances (D) between COG of Mt and COG of one representative chitosan molecule in system Mt-CT, one representative MF molecule in Mt-MF, and one representative PAM molecule in Mt-PAM, plotted against simulation time. For the representative PAM molecule in Mt-PAM, the " \times " symbol corresponds to it being aggregated with other PAM molecules while the square symbol corresponds to it being in an unaggregated state.



Figure 4. Vertical distances (D) between COG of Mt and COG of one representative chitosan molecule in system Mt-CT, one representative MF molecule in Mt-MF, and one representative PAM molecule in Mt-PAM, plotted against simulation time. For the representative PAM molecule in Mt-PAM, the " \times " symbol corresponds to it being aggregated with other PAM molecules while the square symbol corresponds to it being in an unaggregated state.



Figure 5. Snapshots of system Mt-(CT-PAM) at t = (a) 0 ns, (b) 3 ns, (c) 31 ns and (d) 80 ns; and of system Mt-(CT-MF) at t = (e) 0 ns, (f) 10 ns, (g) 40 ns and (h) 80 ns. At t = 0 ns, chitosan molecules are on the right while PAM and MF molecules are on the left. Water is removed for clarity.



Figure 6. Vertical distances (D) between COG of Mt and COG of representative polymer molecules, plotted against simulation time.



Figure 7. Percentage of available donor/acceptor pairs that contributed to H-bonding, among PAM molecules, between PAM and chitosan, and between PAM and Mt in system Mt-(CT-PAM).



Figure 8. Snapshots of system (Mt-CT)-PAM at t = (a) 0 ns, (b) 3 ns, (c) 20 ns, and (d) 80 ns; and of system (Mt-PAM)-CT at t = (e) 0 ns, (f) 0.2 ns, (g) 0.6 ns, and (h) 80 ns. The $3 \times 2 \times 2$ arrays of molecules above Mt in (a) and (e) correspond to PAM and chitosan, respectively. Water is removed for clarity.



Figure 9. Proposed mechanism for the macroscopic flocculation of Mt by the addition of chitosan (a), PAM (b), chitosan and PAM (c). Plates represent the clay mineral particles; chitosan and PAM are represented by strings. Refer to Figure 1 for detailed molecular structures.