

Assessing uncertainties related to the use of satellite remote sensing indices to estimate Gross Primary Production

By

Ronny A. Hernández Mora

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**Department of Earth and Atmospheric Sciences
University of Alberta**

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Abstract

Methods to quantify Gross Primary Production (GPP) are classified into two categories: Eddy Covariance techniques (EC) and satellite data-driven. EC techniques can measure carbon fluxes directly, albeit with spatial constraints. Satellite data-driven methods are promising because they overcome spatial constraints associated with EC techniques. However, satellite-driven products have potentially greater uncertainty than EC methods for GPP estimation such as mixed pixels, cloud cover, and the ability of the sensor to retrieve vegetation under saturation conditions in high biomass environments. Therefore, an effort to analyze and quantify the uncertainty of GPP products derived from satellite platforms is needed. This study quantifies the uncertainty of commonly used satellite vegetation indices such as Normalized Difference Vegetation Index (NDVI), Enhance Vegetation Index (EVI), Chlorophyll/Carotenoid Index (CCI), and Near-Infrared Reflectance Index (NIRv) for GPP estimation compared with direct methods such as EC measurements. We conduct this study on three different sites: the University of Michigan Biological Station (USA), the Borden Forest Research Station flux-site (Canada), and Bartlett Experimental Forest (USA) using traditional regression methods and ML approaches.

Preface

This thesis is an original work by Ronny A. Hernández Mora. No part of this thesis has been previously published.

Dedication

*If you can keep your leaves when all about you, are losing theirs and blaming it on drought;
If you can photosynthesize when others shut their stomates and stop transpiring; yours will
be the canopy and all beneath it and -which is more-you'll be a tree, my seed.*

- Jennifer L. Funk digression from Rudyard Kipling "If" poem

"No one gets anywhere without the help of someone else."

- Franklin Chang Díaz

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Table of contents

| | |
|--|------------|
| Abstract | ii |
| Preface | iii |
| Dedication | iv |
| Acknowledgements | v |
| List of Abbreviations | xv |
| 1. Introduction | 1 |
| 1.1. References | 6 |
| 2. Assessing uncertainties related to satellite remote sensing indices to estimate Gross Primary Production | 10 |
| 2.1. Introduction | 10 |
| 2.2. Methods | 15 |
| 2.2.1. Eddy Covariance sites | 15 |
| 2.2.2. Satellite imagery | 17 |
| 2.2.3. Data Preparation | 20 |
| 2.2.4. Data Analysis | 21 |
| 2.3. Results | 23 |
| 2.3.1. Analysis of GPP-Vegetation Index Relationships Using Linear Models . . | 23 |
| 2.3.2. Analysis of GPP-Vegetation Index Relationships Using GAM Models . . | 24 |
| 2.3.3. LM vs GAM | 30 |
| 2.4. Discussion | 34 |
| 2.5. Conclusions | 36 |
| 2.6. References | 38 |

| | |
|---|-----------|
| 3. A data driven approach to predict GPP from VIs through machine learning methods | 45 |
| 3.1. Introduction..... | 45 |
| 3.2. Methods | 48 |
| 3.2.1. Eddy Covariance sites | 48 |
| 3.2.2. Satellite imagery | 48 |
| 3.2.3. Random Forests..... | 49 |
| 3.2.4. AutoML..... | 50 |
| 3.3. Results..... | 52 |
| 3.3.1. Data-Driven GPP Prediction: A Regression Random Forest Approach . | 52 |
| 3.3.2. The potential of AutoML approaches for GPP predictions | 58 |
| 3.4. Discussion..... | 64 |
| 3.5. Conclusions | 66 |
| 3.6. References | 67 |
| | |
| 4. Conclusions and Future Work | 70 |
| 4.1. Synthesis | 70 |
| 4.2. Limitations..... | 71 |
| 4.3. Future work..... | 72 |
| 4.4. References | 74 |
| | |
| References | 76 |
| | |
| Appendices | 88 |
| | |
| A. Appendices | 88 |
| A.1. Bitstrings tables..... | 88 |
| A.1.1. MODIS State 1km..... | 88 |
| A.1.2. MODIS QC Scan | 90 |
| A.2. Complete LM metrics | 91 |
| A.2.1. Monthly LM metrics | 91 |
| A.2.2. Weekly LM metrics | 92 |
| A.2.3. Daily LM metrics | 93 |
| A.3. Complete GAM metrics | 94 |
| A.3.1. Monthly GAM metrics..... | 94 |

| | |
|-----------------------------------|-----|
| A.3.2. Weekly GAM metrics..... | 96 |
| A.3.3. Daily GAM metrics..... | 98 |
| A.4. Residuals distributions..... | 100 |

List of tables

| | |
|--|----|
| 2.1. ONEFlux sites datasets description | 16 |
| 2.2. ONEFlux Site characteristics overview | 17 |
| 2.3. MODIS (MOD09GA.061 product) bands used to calculate the VIs | 17 |
| 2.4. Summary of Linear models for GPP estimation using the vegetation indices on a monthly (a), weekly (b), and daily (c) basis. MAE and RMSE metrics units are $\text{gC m}^{-2} \text{d}^{-1}$ | 25 |
| 2.5. Summary of GAM models for GPP estimation using the vegetation indices on a monthly (a), weekly (b), and daily (c) basis. MAE and RMSE metrics units are $\text{gC m}^{-2} \text{d}^{-1}$ | 28 |
| 3.1. MODIS (MOD09GA.061 product) bands used for ML methods | 49 |
| 3.2. Summary ML metrics in $\text{gC m}^{-2} \text{d}^{-1}$ | 63 |
| A.1. state_1km bit strings..... | 88 |
| A.2. qc_scan bit strings | 90 |
| A.3. Summary of Linear models for GPP estimation using the vegetation indices on a monthly basis (per site)..... | 91 |
| A.4. Summary of Linear models for GPP estimation using the vegetation indices on a weekly basis (per site). | 92 |
| A.5. Summary of Linear models for GPP estimation using the vegetation indices on a daily basis (per site). | 93 |
| A.6. Summary of the GAM models output for monthly GPP estimation using the every VI as an individual non-linear term..... | 94 |
| A.7. Summary of the GAM models output for monthly GPP estimation using the all VIs per site category as non-linear terms covariates | 95 |
| A.8. Summary of the GAM models output for weekly GPP estimation using the every VI as an individual non-linear term..... | 96 |

| | |
|---|----|
| A.9. Summary of the GAM models output for weekly GPP estimation using the all VIs per site category as non-linear terms covariates..... | 97 |
| A.10. Summary of the GAM models output for daily GPP estimation using the every VI as an individual non-linear term..... | 98 |
| A.11. Summary of the GAM models output for daily GPP estimation using the all VIs per site category as non-linear terms covariates..... | 99 |

List of figures

| | |
|---|----|
| 2.1. Sites locations..... | 15 |
| 2.2. Reference in-situ GPP time series from the study sites on a daily (a), weekly (b), and monthly (c) basis for the University of Michigan Biological Station, Bartlett experimental forest, and the Borden Forest Research Station | 18 |
| 2.3. Total number of observations (pixels) from MODIS classified as high quality (used in the analysis) or other quality (filtered out from the analysis) per site. | 19 |
| 2.4. Total number of observations (pixels) from MODIS classified as high quality (used in the analysis) or other quality (filtered out from the analysis) | 20 |
| 2.5. Monthly high-quality MODIS observations after joining with flux observations containing Gross Primary Productivity (GPP) values higher than 1..... | 22 |
| 2.6. Summary of Linear models for GPP estimation using the vegetation indices on a monthly (a), weekly (b), and daily (c) basis. MAE and RMSE metrics units are $\text{gC m}^2 \text{d}^{-1}$ | 26 |
| 2.7. Summary of GAM models for GPP ($\text{gC m}^{-2} \text{d}^{-1}$) estimation using the vegetation indices. Column A represents the metrics for the monthly models, B the weekly, and C the daily metrics. MAE and RMSE metrics units are $\text{gC m}^2 \text{d}^{-1}$ | 29 |
| 2.8. Scatterplot of MODIS 500m derived VIs and GPP with daily values. Every observation corresponds to the observed GPP from a flux tower site. Total observations corresponds to the number of observations used to obtain the mean of the vegetation index (NDVI, NIRv, CCI and EVI). The red line indicates the GAM fit | 31 |
| 2.9. Scatterplot of MODIS 500m derived VIs and GPP with weekly values. Every observation corresponds to the observed GPP from a flux tower site. Total observations corresponds to the number of observations used to obtain the mean of the vegetation index (NDVI, NIRv, CCI and EVI). The red line indicates the GAM fit | 32 |

| | |
|---|----|
| 2.10. Scatterplot of MODIS 500m derived VIs and GPP with monthly values. Every observation corresponds to the observed GPP from a flux tower site. Total observations corresponds to the number of observations used to obtain the mean of the vegetation index (NDVI, NIRv, CCI and EVI). The red line indicates the GAM fit | 33 |
| 3.1. GPP observed and predicted values from the Random Forest for all the sites at a monthly basis. The red line represents a 1:1 relation. Metrics units are $\text{gC m}^2 \text{d}^{-1}$ | 52 |
| 3.2. GPP observed and predicted values from the Random Forest model for all the sites at a daily basis. The red line represents a 1:1 relation. Metrics units are $\text{gC m}^2 \text{d}^{-1}$ | 53 |
| 3.3. GPP observed and predicted values from the Random Forest for all the sites at a weekly basis. The red line represents a 1:1 relation. Metrics units are $\text{gC m}^2 \text{d}^{-1}$ | 53 |
| 3.4. Variable of importance derived from the Random forest model for the monthly values at 500 m spatial resolution model..... | 54 |
| 3.5. Variable of importance derived from the Random forest model for the weekly values at 500 m spatial resolution model..... | 54 |
| 3.6. Variable of importance derived from the Random forest model for the daily values at 500 m spatial resolution model..... | 55 |
| 3.7. Shapley values derived from the Random forest model for the daily values at 500 m spatial resolution model. Predicted value for the low GPP value is 0.59 (A) and 12.7 for the selected high GPP value (B). | 56 |
| 3.8. Shapley values derived from the Random forest model for the weekly values at 500 m spatial resolution model. Predicted value for the low GPP value is 1.59 (A) and 11.0 for the selected high GPP value (B). | 57 |
| 3.9. Shapley values derived from the Random forest model for the monthly values at 500 m spatial resolution model. Predicted value for the low GPP value is 3.31 (A) and 10.6 for the selected high GPP value (B). | 57 |
| 3.10. GPP observed and predicted values from the autoML for all the sites at a monthly basis. The red line represents a 1:1 relation. Metrics units are $\text{gC m}^2 \text{d}^{-1}$ | 58 |

| | |
|---|-----|
| 3.11. GPP observed and predicted values from the autoML for all the sites at a weekly basis. The red line represents a 1:1 relation. Metrics units are $\text{gC m}^{-2} \text{d}^{-1}$ | 59 |
| 3.12. GPP observed and predicted values from the autoML for all the sites at a daily basis. The red line represents a 1:1 relation. Metrics units are $\text{gC m}^{-2} \text{d}^{-1}$ | 59 |
| 3.13. Variable of importance derived from the autoML model for the daily values at 500 m spatial resolution model. | 60 |
| 3.14. Variable of importance derived from the autoML model for the weekly values at 500 m spatial resolution model. | 61 |
| 3.15. Variable of importance derived from the autoML model for the monthly values at 500 m spatial resolution model. | 62 |
| A.1. Residuals distributions for each of the LMs for GPP estimation using the vegetation indices on a monthly (a), weekly (b), and daily (c) basis. | 100 |
| A.2. Residuals distributions for each of the GAMs for GPP estimation using the vegetation indices on a monthly (a), weekly (b), and daily (c) basis. | 101 |

List of Abbreviations

| Abbreviation | Full phrase |
|--------------|---|
| AIC | Akaike Information Criterion |
| APAR | Absorbed Photosynthetically Active Radiation |
| BIC | Bayesian Information Criterion |
| CCI | Chlorophyll/Carotenoid Index |
| CRI | Chlorophyll Index Red |
| DT | Day Time method |
| EC | Eddy Covariance |
| EO | Earth Observation |
| EVI | Enhanced Vegetation Index |
| GAM | Generalized Additive Model |
| GEE | Google Earth Engine |
| GPP | Gross Primary Production |
| LAI | Leaf Area Index |
| LM | Linear Model |
| LUE | Light Use Efficiency |
| MODIS | Moderate Resolution Imaging Spectroradiometer |
| NDVI | Normalized Difference Vegetation Index |
| NEE | Net Ecosystem Exchange |
| NIR | Near Infrared |
| NIRv | Near-Infrared Reflectance Index |
| PAR | Photosynthetically Active Radiation |
| QA | Quality Assurance |
| QC | Quality Control |
| RECO | Ecosystem Respiration |

| | |
|-------|---|
| RMSE | root mean square error |
| SIF | Solar-Induced Fluorescence |
| SWIR | Short wave infrared |
| VI | Vegetation Index |
| fAPAR | fraction of Photosynthetically Active Radiation |

1. Introduction

The study and assessment of ecosystem dynamics has greatly benefited from the development and advances of the remote sensing field (Running et al. 2004; Baldocchi 2020a) coupled with the growing freely available remote sensing data (Xiao et al. 2019). Elements on the Earth's surface exhibit distinctive spectral signatures arising from their physical and chemical properties and interactions with electromagnetic radiation (Montero et al. 2023). Environmental factors (fires, floods, drought, etc) can affect these interactions modifying the spectral signatures which can provide insights about surface processes when measured by remote sensing instruments (Montero et al. 2023).

The launch in the 1970s of the Earth Resources Technology Satellite (ERTS), later known as Landsat-1 (Donato 1997), initiated a new era in the remote sensing field of the terrestrial biosphere, with a special focus on vegetation monitoring (Montero et al. 2023). With information available about spectral bands, studies about their relations to biophysical characteristics became an object of study to estimate biomass and photosynthetic activity (Myneni et al. 1997).

In fact, one application of remote sensing has been the development of vegetation indices (VI) to study the characteristics of vegetation canopies and their functioning (Houborg, Fisher, and Skidmore 2015). VIs are a function of two or more spectral bands, typically derived from satellite images (Zeng et al. 2022) that can contain information about the status of the vegetation, biochemical characteristics, and structural characteristics such as Leaf Area Index (LAI), among others (Huete et al. 2002).

Early studies about spectral bands relations, led to the development of the Normalized Difference Vegetation Index (NDVI), which showed how the Near Infrared (NIR) and Red ratio had advantages over other bands ratios for monitoring vegetation biomass and structure given surface reflectance measurements in the shortwave spectrum (Tucker 1979; A. J. Richardson and Wiegand 1977). The basis of NDVI rests in the fact that chlorophyll absorption leads to

a decrease in red reflectance with increasing green vegetation relative to the NIR reflectance, that is minimally impacted by chlorophyll absorption (Huete et al. 2010).

The use of a relative indicator serves to reduce the effect of factors such as canopy spatial structure and non-photosynthetic vegetation on relationships between NDVI and biophysical variables such as biomass or absorbed photosynthetically active radiation variances (Sellers 1987; Huete et al. 2010). Nonetheless, NDVI is sensitive to soil background and saturation at high biomass (Huete et al. 2002). To overcome this problem, the Enhance Vegetation Index (EVI) was designed to improve vegetation characterizations such as canopy greenness without soil background sensitivity and aerosol variations by combining the blue, red, and NIR bands (Huete 1988). EVI is also less prone to saturate at high biomass instances (Gao 2000).

The ability to combine reflectance from multiple spectral bands, coupled with advancements and the development of new sensor technologies introducing additional bands or different spectral characteristics, has created opportunities to develop multiple VIs and the total number of indices has grown steadily (Zeng et al. 2022). These equations can function as proxies for various vegetation canopy properties or enhance existing estimations across different ecosystems (Huete et al. 2010).

According to the Awesome Spectral Indices open community catalogue (version 4.0.0), there are 127 spectral indices designed to monitor vegetation. The aim behind the ongoing development of new VIs is to achieve a more accurate mapping of ecosystem variables such as GPP (Montero et al. 2023). These indices can be classified into 3 main categories: biophysical, biochemical and physiological properties (Zeng et al. 2022).

One important variable in understanding ecosystem processes is GPP. At present, VIs derived from satellites cannot directly estimate GPP (Gensheimer et al. 2022) and instead, instruments like eddy covariance have been the traditional method to estimate GPP more directly (Baldocchi 2020a).

GPP represents the total amount of carbon compounds produced by plant photosynthesis in a given period (Ashton et al. 2012). GPP plays a crucial role in the carbon cycle since it quantifies the ability of vegetation to fix carbon from the atmosphere using only solar energy and nutrients (Xiao et al. 2019). GPP is responsive to a range of factors, including abiotic elements such as radiation, temperature, and precipitation (Beer et al. 2010), as well as biotic factors like vegetation type, leaf chemical traits, and species composition (Musavi et al. 2017).

Some of the carbon taken in by GPP is released back to the atmosphere through plant respiration (autotrophic respiration), and the difference between GPP and plant respiration is called Net Primary Productivity (NPP) (Lieth 1975; Xiao et al. 2019). GPP, along with Ecosystem Respiration (ER), which includes both autotrophic and heterotrophic respiration, together determines Net Ecosystem Exchange (NEE) (Xiao et al. 2019). NEE is fundamental to the planetary carbon cycle as it represents the net accumulation or loss of carbon by the ecosystem. Measurement of GPP and NEE is critical to understanding the role of the biosphere in the carbon cycle as well as the status and trends of vegetation productivity (Schimel 1995).

Eddy Covariance (EC) is a method used to measure the exchange of energy and materials between ecosystems and the atmosphere within an area that can range from a few hundred meters to a few kilometers (Pabon-Moreno et al. 2022). Specifically, EC methods can estimate GPP in-situ at a temporal resolution of seconds, by subtracting estimates of respiration from chamber measurements or models from direct measurements of NEE (Reichstein et al. 2005). However, EC estimates are both, costly and limited in spatial coverage (Baldocchi 2020a). Also, uncertainties from this method lie in the typical increase in spatial and temporal variability in GPP over heterogeneous landscapes such as forest boundaries (Reinmann and Hutrya 2017). The prevalent practice of taking measurements primarily from intact forests introduces the risk of potentially misleading estimations for the broader forest being quantified, given the potential differences in dynamics between edge and interior environments (Smith et al. 2018).

Since direct GPP measurements from satellites are not presently available, one of the primary objectives of remote sensing is to enhance the spatial coverage of EC GPP estimates (Xiao et al. 2019). In this context, EC values prove invaluable as information for calibrating VIs values with GPP inferred from flux towers (Huete et al. 2010). The utilization of EC measurements bridges the gap in GPP assessment, providing a practical means to evaluate and refine the reliability of remotely sensed data against ground-based observations (Xiao et al. 2019).

Various methods exist for estimating GPP through satellite measurements. Satellite-derived VIs serve as commonly employed proxies for GPP, whereas quantifying GPP more directly often involves the use of Light Use Efficiency (LUE) models, and process-based models (Xiao et al. 2019). These diverse methodologies cater to different aspects of GPP estimation, with satellite-derived VIs offering a convenient indirect assessment, while LUE models, and

process-based models provide more detailed and direct quantification (Lin et al. 2021).

The concept of a more detailed and direct quantification of GPP achieved through LUE and process-based models can be attributed to the incorporation of a constant maximal light use efficiency, which is a plant functional type dependent parameter that is downregulated by stress factors such as temperature and vapor pressure deficit (Pabon-Moreno et al. 2022). This approach is then integrated with APAR derived from remote sensing data to estimate GPP in the case of LUE models (Running et al. 2004). In contrast, process-based models distinguish themselves by prioritizing a mechanistic description of the photosynthetic biochemical processes at both leaf and canopy scales, incorporating the comprehensive theoretical underpinnings of photosynthesis (Ryu et al. 2011). This methodology not only refines GPP quantification but also offers a more in-depth understanding of the underlying physiological processes governing vegetation productivity.

However, both Light Use Efficiency (LUE) models and process-based models demand a multitude of inputs, some of which may not be readily available for all sites. In such cases, a data-driven approach becomes particularly advantageous, and VIs serve as effective proxies for GPP estimation. Notably, emerging VIs like the Near-Infrared Reflectance Index (NIRv) demonstrate a capability to account for 68% of the monthly variability in GPP across 105 fluxnet sites (Badgley et al. 2019). Additionally, indices like CCI, a pigment-based index designed to capture carotenoid/chlorophyll ratios during seasonal photosynthetic activity in evergreen leaves (Gamon et al. 2016), have demonstrated proficiency as estimators of GPP in certain forest types. These indices offer valuable insights into the dynamics of vegetation productivity.

Despite the notable progress in VIs development and their user-friendly applications there is a consensus that no single VI can be universally applied to address every GPP estimation across species and periods, even for a single biome (Zeng et al. 2022). VI measurement error poses additional challenges to remote sensing applications, with diverse sources affecting satellite-derived data, including atmospheric effects, retrieval errors, cloud contamination, and sensor degradation (Van Leeuwen et al. 2006; Fang et al. 2012). Thus, a quantitative and rigorous understanding of the specific context and characteristics of multiple sites is essential to understanding the potential and limits of VIs for GPP estimation over large extents of space and time (Zeng et al. 2022).

Together with the advancement of new VIs, there has been an upsurge in data analysis meth-

ods, specifically Machine Learning (ML) approaches, has garnered significance in recent years. The capability of ML methods to navigate non-linear relationships proves crucial, particularly when predicting GPP from various predictors (Meyer et al. 2019). However, despite their simplicity and effectiveness, ML approaches encounter challenges, and the resulting fluxes estimates bear various sources of uncertainty. The distribution and quantity of EC sites play a pivotal role in influencing the accuracy and variability of regional flux estimates derived from machine learning methods (Papale et al. 2015).

The goal of this study is to assess the advantages and constraints associated with the utilization of 4 widely employed VIs for estimating GPP. The study is limited to temperate broadleaf forests of the Eastern United States of America and Canada as it required consistent multi-annual EC GPP measurements over multiple sites to quantify uncertainty related to the spatial location of calibration data. The VIs are derived from the Moderate Resolution Imaging Spectroradiometer (MODIS), which has long-standing temporal records and acceptable spatial resolution to study extended forest areas. The analysis employs both traditional regression methods and ML techniques to evaluate the performance of these indices in estimating GPP. This methodological approach seeks to provide a nuanced understanding of the efficacy and limitations of the selected indices in capturing the complexities of ecosystem processes, particularly in the context of temperate broadleaf forests.

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2. Assessing uncertainties related to satellite remote sensing indices to estimate Gross Primary Production

2.1. Introduction

Vegetation Gross Primary Production (GPP) is the total amount of carbon fixation by plants through photosynthesis (Badgley et al. 2019). Quantifying GPP is essential for understanding land-atmosphere carbon exchange (Köhler et al. 2018), ecosystem function, and ecosystem responses to climate change (Guan et al. 2022; Brown et al. 2021; Myneni and Williams 1994). However, terrestrial GPP cannot be directly measured due to the contribution of respiration to land carbon fluxes (Anav et al. 2015). Instead, GPP can be inferred in a non-destructive manner by the net carbon exchange measurements at the ecosystem level, or at broader scales using models that incorporate various assumptions and limitations (Reichstein et al. 2005; M. Jung, Reichstein, and Bondeau 2009).

GPP estimations can be grouped into two broad categories: Eddy Covariance (EC) techniques, and satellite data-driven methods (Guan et al. 2022; Xie et al. 2020). EC is the primary in-situ non-destructive method for measuring terrestrial fluxes, and specifically for quantifying the exchange of CO₂ between land and the atmosphere, using advanced field instrumentation (Baldocchi 2020b; Badgley et al. 2019; Ryu, Berry, and Baldocchi 2019; Tramontana et al. 2016). However, EC measurements come with certain limitations, such as their relatively low spatial resolution, typically less than $< 1\text{km}^2$, which constrains the accuracy of estimating ecosystem carbon and water fluxes at regional and global scales. (Badgley, Field, and Berry 2017).

Additionally, it's important to note that EC techniques directly measure Net Ecosystem Exchange (NEE), not GPP. Subsequently, GPP must be estimated by both subtracting respiration using models and ancillary measurements while accounting for the removal or deposition of carbon stocks due to natural or anthropogenic transport processes such as water flow, fires, or harvest (Beer et al. 2010; Reichstein et al. 2005).

The satellite data-driven models constitute the second category of methods to estimate GPP. Because of their dependency on earth observation platforms, they are not spatially constrained but can have greater uncertainties than the EC techniques (Ryu, Berry, and Baldocchi 2019; Wang et al. 2011). Satellite data-driven models can be classified into process-based models, Light Use Efficiency models (LUE), and Vegetation Index models (Xie et al. 2020).

Process-based models integrate climate, canopy, and soil information derived from multiple sources, including satellite EO, into biophysical models of carbon, water, energy, and nutrient cycles with varying levels of detail (Running and Coughlan 1988; Harris et al. 2021). While these models can be scaled globally (Beer et al. 2010; M. Jung, Reichstein, and Bondeau 2009) they require many parameters that may not be readily available in changing landscapes or for fine-scale studies.

LUE models are based on the concept of radiation conversion efficiency and take into consideration ecological processes (Liu et al. 1997; Heinsch et al. 2006). This efficiency signifies the amount of carbon a specific vegetation type can fix per unit of solar radiation (J. L. Monteith 1972). Initially used for NPP estimation (Prince 1991), LUE models were subsequently adapted for GPP and respiration calculations (Goetz et al. 1999). These models explicitly account for the impact of environmental stress on plant physiological responses. The imposition of environmental stressors may lead to a reduction in the rates of daily carbon assimilation, thereby diminishing overall efficiency. (Prince 1991; Running and Coughlan 1988).

The GPP product from the Moderate Resolution Imaging Spectroradiometer (MODIS) employs an algorithm based on the radiation conversion efficiency concept. This algorithm establishes a connection between absorbed photosynthetically active radiation (APAR) and the LUE term (Heinsch et al. 2006) as shown in Equation 2.1.

$$GPP = PAR \times fAPAR \times LUE \quad (2.1)$$

Where PAR is the incident photosynthetically active radiation (John Lennox Monteith and

Unsworth 2013) and fAPAR is the fraction of the PAR that is effectively absorbed by plants (GCOS 2016). The LUE term depends on vegetation type but also physiological conditions are driven by water availability, temperature stress, and vapour pressure deficit (Goetz et al. 1999; Running and Coughlan 1988). Obtaining these variables for every vegetation type on Earth can be challenging, introducing assumptions that amplify uncertainties (Goetz et al. 1999). Nevertheless, under unstressed conditions, LUE remains constant for a given vegetation type, requiring only PAR and fAPAR to assess primary productivity (Running et al. 2004).

VIs are the Satellite data-driven models' third approach. VIs are a summary of non-linear functions of surface bi-directional reflectance spectra (Myneni and Williams 1994) derived from optical sensors that are combined with climate variables to calculate GPP (C. Wu, Chen, and Huang 2011). This is usually done with some form of regression and physical methods that associate interactions between vegetation and incoming radiation (Fernández-Martínez et al. 2019).

VIs have been used to provide inputs to Equation 2.1 related to fAPAR from regional to global extents or to estimate GPP (Sellers et al. 1994; Running et al. 2004). Some of the most common VIs to estimate GPP are the Normalized Difference Vegetation Index (NDVI), the Enhanced Vegetation Index (EVI), the Near-Infrared Reflectance Index (NIRv), or the Chlorophyll/Carotenoid Index (CCI) among others. (Balzarolo, Peñuelas, and Veroustraete 2019; Rahman et al. 2005, 2005; Xie et al. 2020; Badgley et al. 2019; Zhang et al. 2020; Sellers et al. 1994). These VIs are based on a spectral reflectance ratio between the red and near-infrared regions of the electromagnetic spectrum (Glenn et al. 2008) which tracks an integrated impact of fraction of photosynthetically active radiation (fAPAR) and LUE on productivity (Myneni and Williams 1994).

An index such as NDVI is good for detecting structural vegetation changes in seasonal variability, but it becomes saturated with high biomass conditions (Badgley, Field, and Berry 2017). Other indices such as EVI can overcome the soil and atmospheric effects by adding the blue band (Huete 1988) and may be better suited for predicting GPP in large biomass forests (Badgley, Field, and Berry 2017). Nonetheless, it has been evaluated in a narrow range of ecosystems and needs inputs of start and end dates of the growing seasons which can increase uncertainties (Shi et al. 2017). For specific types of ecosystems such as evergreen conifers, CCI can track the seasonality of daily GPP due to its sensibility to the chlorophyll/carotenoid

pigment ratios (Gamon et al. 2016).

Other indices such as the near-infrared reflectance index (NIRv) have been formulated to address the mixed-pixel effect (pixel with vegetated and non-vegetated features) and to determine the vegetation photosynthetic capacity (Badgley et al. 2019). NIRv is defined as the fraction of reflected NIR light that originates from vegetation. NIRv was originally proposed as a replacement for fPAR in LUE models in that the NIR and PAR reflectance of vegetation are correlated and the scaling by NDVI corrects for soil contributions in the signal (Badgley et al. 2019). NIRv has been shown to have a stronger correlation to GPP at flux towers and on a regional basis than fAPAR notwithstanding the fact that GPP is directly related to fAPAR. NIRv has been extended to include weighting with PAR (Dechant et al. 2022) and replacing the NIR reflectance with NIRv radiance (G. Wu et al. 2020). Both of these approaches have been shown to have even stronger correlations with GPP than NIRv as could be expected since they either directly or indirectly weight the NIRv with PAR.

The stronger correlation between the NIRv index to GPP in comparison to the correlation between VIs related to fAPAR and GPP seemingly contradicts the hypotheses in the LUE model that GPP should be linearly related to fAPAR. There are two explanations: i. Many of the reported comparative studies use fAPAR based on VIs and not APAR. We hypothesize that the NIR indices are simply better estimators of APAR than these VI approximations due to lower measurement error for the NIR indices or a stronger physical relationship between them and APAR versus the historical VIs. ii. The strength of the correlation between APAR and GPP depends on LUE having a linear relationship to observed APAR. While this may hold in some circumstances (e.g. early seasonal measurements for vegetation such as crops where leaf chlorophyll concentration increases during the growth phase) it is not the case in general and definitely during stress conditions (J. L. Monteith 1972).

Despite these efforts, relying solely on remote satellite VIs presents a challenge. The assumptions accompanying VI models suggest the importance of systematically quantifying their predictive capacity for GPP to validate and improve their accuracy (Anav et al. 2015; Brown et al. 2021) across various growing seasons and locations. This is particularly crucial because photosynthesis regulation can occur with no major changes in canopy structure or leaf pigments that can undergo without being detected with reflectance data (Pabon-Moreno et al. 2022; Pierrat et al. 2022), implying that temporal aggregation may be critical for VI models. In-situ eddy-covariance flux measurements coupled with locally calibrated models for

respiration (Baldocchi 2020b) represent a suitable reference GPP for validation solution given that they represent site-level observations (Chu et al. 2021).

As such, the main objective of this M.Sc. thesis chapter is to quantify and compare the uncertainty associated with the VI/LUE models to estimate GPP. To control for variability in environmental conditions (E), sites are selected that share the same land cover, climate, and biome characteristics. VIs models are evaluated by comparison to EC based estimates of GPP for multiple growing seasons from the Bartlett Experimental Forest (USA), the Borden Forest Research Station flux-site (Canada), and the University of Michigan Biological Station (USA). Considering the reviewed studies, we hypothesize that the uncertainty of this approach will depend on the nature of the VI and the spatial and temporal aggregation of application. Specifically, we expect that i. the NIRv and CCI indices will consistently demonstrate a stronger correlation and lower prediction uncertainty for GPP compared to NDVI and EVI across the tested sites. ii. larger temporal aggregations will result in improved predictions due to the reduction in observation variability.

2.2. Methods

2.2.1. Eddy Covariance sites

We used three deciduous broadleaf forest sites located in the northern hemisphere (see Figure 2.1) with eddy covariance (EC) data collected by Ameriflux. For each site, we used the daily, weekly, and monthly GPP values (GPP_DT_VUT_REF variable) estimated using the ONEFlux workflow (Pastorello et al. 2020). The ONEFlux processing does the estimation of the CO₂ fluxes into GPP and Ecosystem Respiration (RECO) from Net Ecosystem Exchange (NEE) through two methods known as daytime and nighttime. Here we selected the daytime method (DT) which uses daytime and nighttime to parameterize a model with two components: one based on light response curve and vapour pressure deficit and a second one using a respiration-temperature relationship to estimate RECO which in turn is used to obtain the difference with NEE and provide GPP (Pastorello et al. 2020).

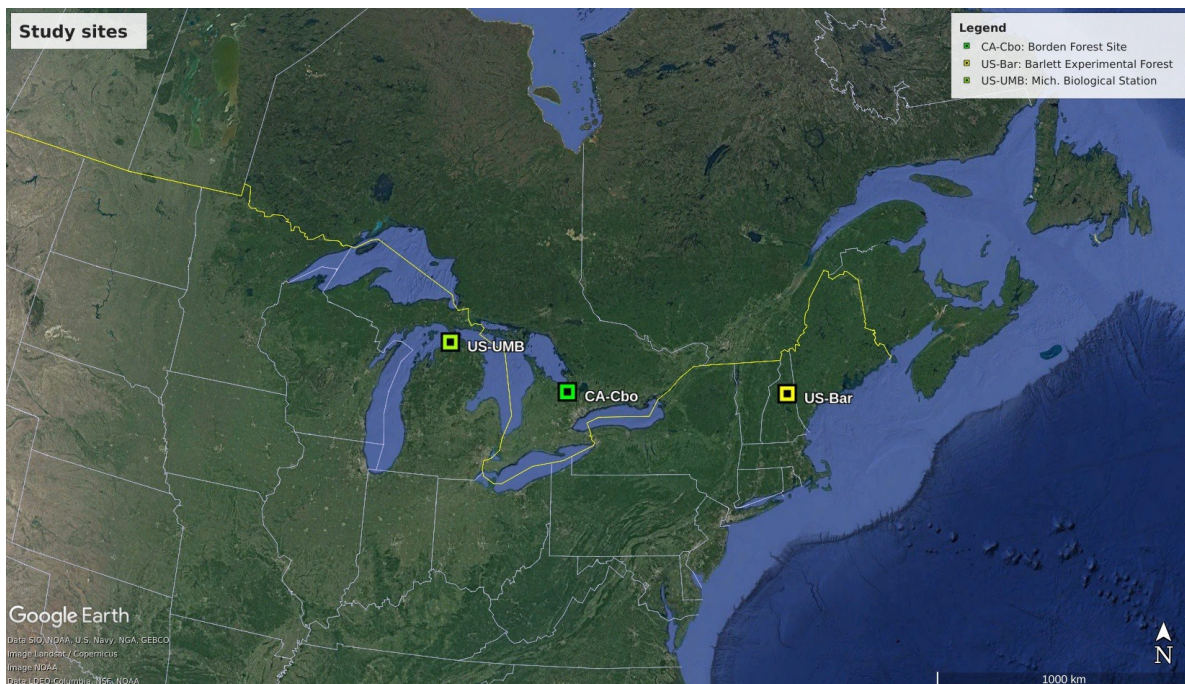


Figure 2.1.: Sites locations

Figure 2.2 displays the GPP trends for the University of Michigan Biological Station, Bartlett Experimental Forest, and the Borden Forest Research Station. Additional details regarding the characteristics of these datasets can be found in Table Table 2.1.

Table 2.1.: ONEFlux sites datasets description

| Site | Data range available | Dataset name | Reference |
|----------|----------------------|--|--|
| Bartlett | Jan 2015 to Dec 2017 | US-Bar: Barlett Experimental Forest (version: beta-3) | (Staebler 2019) |
| Borden | Jan 2015 to Jan 2022 | CA-Cbo: Ontario - Mixed Deciduous, Borden Forest Site | (A. Richardson and Hollinger 2016) |
| Michigan | Jan 2015 to Jan 2018 | US-UMB: Univ. of Mich. Biological Station (version: beta-4) | (C. Gough, Bohrer, and Curtis 2016) |

The Bartlett experimental forest is located in New Hampshire, USA (44°06 N, 71°3 W). This site is characterized by a forest with an average canopy height ranging from 20 to 22 meters with a mean annual temperature of 6°C. Despite events such as a hurricane in 1938 and small scale forest management, the forest's mean stand age is around 120-125 years (Ouimette et al. 2018).

The second flux site is Borden Forest Research Station located in Ontario (44°19 N, 79°56 W), Canada. This is one of the largest patches of forest in Southern Ontario which has been collecting EC data since 1996 (Rogers et al. 2020). This site has a forest cover of over 60% with a height of approximately 22 m. It's a deciduous broadleaf natural re-growth forest since 1916 dominated by woody vegetation (Lee et al. 1999).

The third site is the University of Michigan Biological Station which is located in northern Michigan, USA (45°350 N 84°430 W). The site has a forest with different succesional stages, with an average stand age of 90 years (C. M. Gough et al. 2010), and a mean height of 22m. The mean annual temperature is around 5.5°C (C. M. Gough et al. 2021).

The three sites exhibit similar characteristics, indicating their representation of a specific ecosystem type. This uniformity enables meaningful comparisons and offers valuable insights into the relationship between GPP and VIs. A tabular summary of site characteristics, guided by insights detailed in Teets (2022), is presented in Table 2.2.

Table 2.2.: ONEFlux Site characteristics overview

| Variable | Site | | |
|--------------------------------|---------------------|-------------------------|----------------------|
| | Bartlett | Michigan | Borden |
| Mean annual temperature (°C) | 6 | 5.5 | 7.4 |
| Mean annual precipitation (mm) | 1246 | 803 | 784 |
| Elevation (m) | 272 | 234 | 209 |
| Dominant genera | Acer, Fagus, Betula | Populus, Quercus, Pinus | Acer, Pinus, Populus |
| Climate Koeppen ¹ | Dfb | Dfb | Dfb |

¹**D** stands for the warm-summer continental or hemiboreal climate. **f** indicates that this climate has significant precipitation in all seasons. **b** indicates that the warmest month has an average temperature between 22°C and 28°C.

2.2.2. Satellite imagery

We used data from the Terra Moderate Resolution Imaging Spectroradiometer (MODIS), specifically the collection MOD09GA Version 6.1 product (MODIS/Terra Surface Reflectance Daily L2G Global 1 km and 500 m SIN Grid) chosen for its daily sampling and broad temporal coverage. Data retrieval was performed using Google Earth Engine (GEE). For each site, a square polygon with an area of 3 km surrounding the EC tower was defined, and the pixel values within this polygon were extracted for comprehensive analysis.

The MODIS contains the surface spectral reflectance from bands 1 through 7 with a spatial resolution of 500m, with corrections for atmospheric conditions such as aerosols, gasses, and Rayleigh scattering (Vermote 2021), and validation of cloud-free pixels as well. Bands used to derive the vegetation indices are shown in Table 2.3

Table 2.3.: MODIS (MOD09GA.061 product) bands used to calculate the VIs

| Name | Description | Resolution | Wavelength |
|-------------|-------------|------------|------------|
| sur_refl_01 | Red | 500 meters | 620-670nm |
| sur_refl_02 | NIR | 500 meters | 841-876nm |
| sur_refl_03 | Blue | 500 meters | 459-479nm |
| sur_refl_04 | Green | 500 meters | 545-565nm |

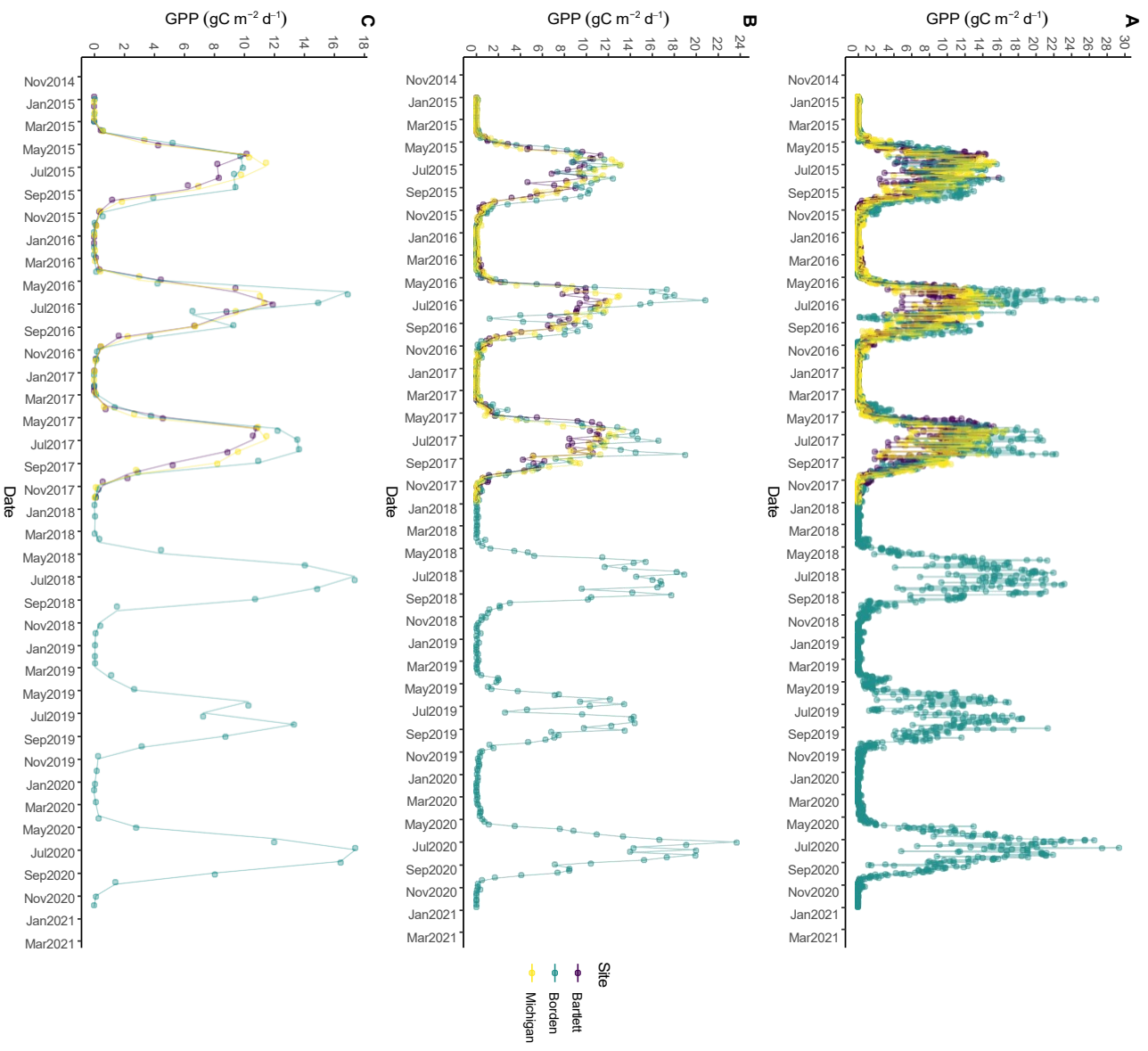


Figure 2.2.: Reference in-situ GPP time series from the study sites on a daily (a), weekly (b), and monthly (c) basis for the University of Michigan Biological Station, Bartlett experimental forest, and the Borden Forest Research Station

The data processing involved three main steps: selecting high-quality pixels, scaling band values, and calculating vegetation indices. MODIS product data have 4 bit-encoded variables which provide information about the observation quality. From those variables, only the 1km Reflectance Data State QA (state_1km) and Surface Reflectance 500m Quality Assurance (qc_500m) variables were used along with each of the band's bits quality indicators, as 250m scan value information (q_scan) was not informative and Geolocation flags (g_flags) had the same value for all observations. The bit-encoded variables were transformed into categorical strings, and only the categories indicating the best quality were selected to filter the pixels (Figure 2.3). The specific bit strings selected for state_1km are shown in Table A.1 and for qc_500m in Table A.2. Subsequently, the surface reflectance for each filtered pixel was determined by scaling the digital number recorded by 0.0001.

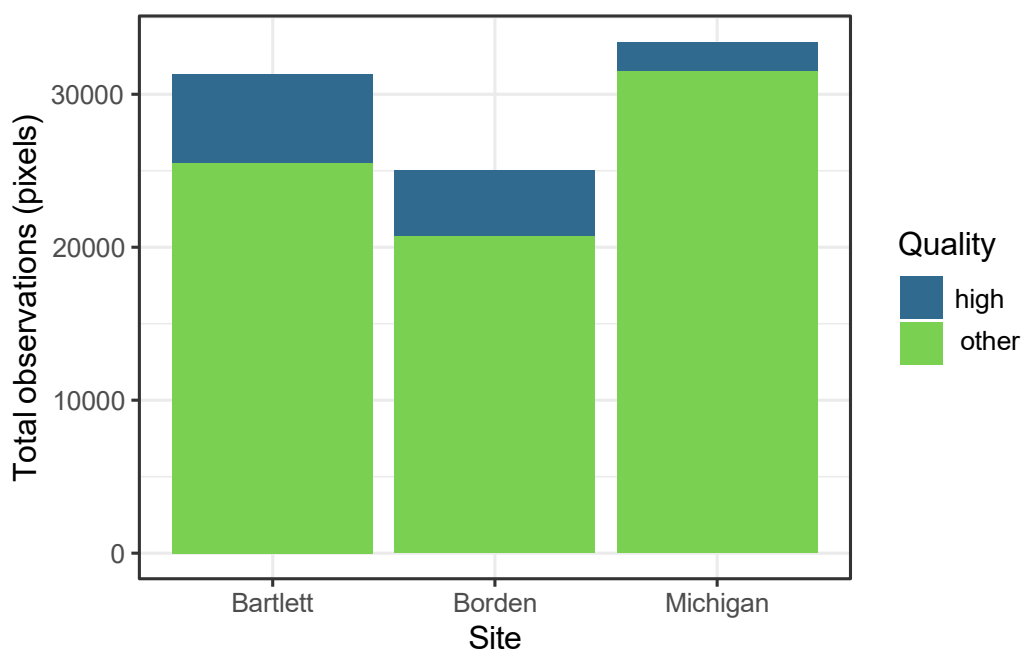


Figure 2.3.: Total number of observations (pixels) from MODIS classified as high quality (used in the analysis) or other quality (filtered out from the analysis) per site.

Following the MODIS Collection 6.1 (C61) LSR Product User Guide (Vermote 2021), any scaled value that fell outside the range of 0 to 1 was considered a fill value or uncorrected Level 1B data and was subsequently discarded. These values were deemed unreliable or lacking meaningful information for the analysis. The number of high-quality surface reflectance observations was summarized on a monthly basis for each site (Figure 2.4)

The VIs NDVI (Equation 2.2), NIRv (Equation 2.3), EVI (Equation 2.4), and CCI (Equation 2.5) were calculated and then matched with the corresponding date in the flux datasets (Green,

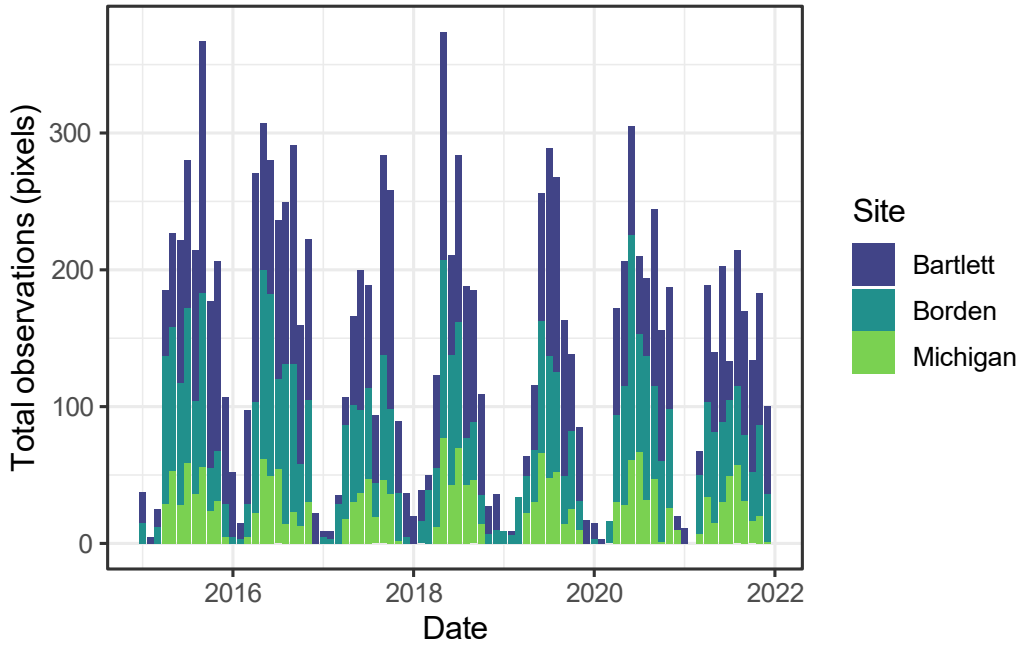


Figure 2.4.: Total number of observations (pixels) from MODIS classified as high quality (used in the analysis) or other quality (filtered out from the analysis)

Red, and NIR correspond to the bands defined in Table 2.3).

$$NDVI = \frac{NIR - Red}{NIR + Red} \quad (2.2)$$

$$NIRv = NIR \times \frac{NIR - Red}{NIR + Red} \quad (2.3)$$

$$EVI = 2.5 \times \frac{NIR - Red}{(NIR + 6 \times Red - 7.5 \times Blue + 1)} \quad (2.4)$$

$$CCI = \frac{Green - Red}{Green + Red} \quad (2.5)$$

2.2.3. Data Preparation

Three datasets were prepared for each site: a daily, a weekly, and a monthly dataset. These datasets were generated from the satellite imagery data with the selected high-quality pixels and the ONEFluxprocess data in order to capture variations in vegetation indices (VI), band values, and GPP over different time scales.

The daily dataset included the VI values, band values only from the high-quality pixels, and GPP measurements derived from the ONEFlux process collected on a daily basis for the time period of available GPP at each site Table 2.1. This dataset provided a high-resolution representation of the variables, allowing for a detailed analysis of their daily fluctuations.

The weekly and monthly datasets were derived from the corresponding daily dataset. These datasets contained summarized values of the VIs and band values, aggregated over the weekly and monthly time frames, respectively. The aggregation process involved calculating an average for the VIs and band values within each week or month.

For the GPP values in the weekly and monthly datasets, rather than summarizing the daily GPP values, the GPP measurements for the weekly and monthly time frames were obtained directly from the ONEFlux process, which provided a reliable estimation of GPP for these longer time intervals.

By creating these three datasets (daily, weekly, and monthly), the study allowed for a comprehensive analysis of the VIs, band values, and GPP at different temporal resolutions. This approach provided insights into the temporal dynamics and patterns of the variables, enabling a more thorough understanding of the processes and relationships under investigation.

For each site and across all time scales (daily, weekly, and monthly), GPP values below $1 \text{ gC m}^{-2} \text{ d}^{-1}$ were excluded. These values were deemed either below the detection limit or insufficient to make a significant contribution to the overall analysis, particularly in representing meaningful vegetation productivity. This process aimed to refine the dataset and focus on values within a range considered more pertinent to the growing season. The final number of observations per site after selecting the high-quality pixels and GPP observations are shown in Figure 2.5.

2.2.4. Data Analysis

To address the potential non-linearity of the lowest uncertainty data-driven model across different temporal aggregation scales or for specific VIs, we employed two distinct modeling approaches: a Linear Model (LM) and a Generalized Additive Model (GAM). GAM models allow a better fit for those cases where the distribution and variability observed in the data is greater, due to the varying temporal scales. We aimed to evaluate whether this variance could be better explained by this type of model, which might not be optimal to capture with a linear

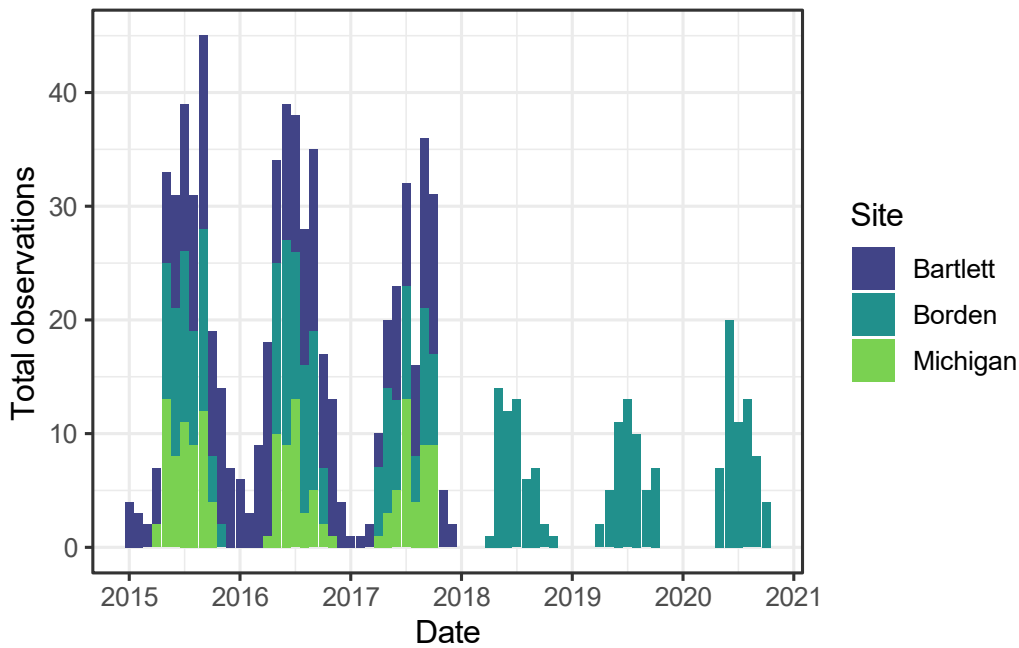


Figure 2.5.: Monthly high-quality MODIS observations after joining with flux observations containing Gross Primary Productivity (GPP) values higher than 1.

relationship, potentially leading to greater residuals. Both models' approaches were applied individually to test our hypothesis regarding VI predictive uncertainty. Additionally, a single model was applied to assess whether a combination of VIs could improve GPP estimation uncertainty.

This facilitated a comparative analysis of how each individual index independently explains the variation in GPP per site. Additionally, the performance of each VI was also evaluated using models calibrated using all sites to assess their robustness to spatial variability within the selected temperate broadleaf biome across various temporal scales. Another step of the analysis was the examination of the relationship between GPP and all indices functioning as covariates, both on an individual site basis and without site distinction.

In total, per each site plus the category of all-sites we created 5 linear models and 5 GAM models for every timescale (daily, weekly, and monthly): one per each index (NDVI, CCI, EVI, NIRv) and one with all the indices as covariates (NDVI + CCI + EVI + NIRv), resulting in a total of 120 models. To assess and compare the models' performance, we calculated the coefficient of determination (R^2) to measure the correlation between the actual observations and predictions. Additionally, we used the Root Mean Squared Error (RMSE) and the Mean Absolute Error (MAE) as indicators of the model estimation error.

2.3. Results

2.3.1. Analysis of GPP-Vegetation Index Relationships Using Linear Models

The Table 2.4 provides a summary of linear models used for GPP estimation at each site, employing the vegetation indices as predictors. For each site and predictor, the table includes the relevant model summary statistics, such as R^2 , MAE, and RMSE as indicators of the statistical significance of the model fit. All models add p -values < 0.05 . More metrics such as the p -value, adjusted r -squared, the Akaike Information Criterion (AIC), and Bayesian Information Criterion (BIC) are displayed in the Table A.3 for monthly results, Table A.4 for weekly results, and the Table A.5 for daily outputs. A residuals distribution for each of the models is in Figure A.1

Our findings show that when using all the indices as covariates within the linear model, the model's performance demonstrates better outcomes compared to using any single index alone across all scenarios. In the assessment of individual VI performance on a monthly basis, CCI tend to perform better than EVI, NDVI, and NIRv. Although all models were statistically significant ($p < 0.05$), it is important to note that for the Bartlett and Michigan sites, while CCI shows favourable predictive accuracy, its advantage over the NIRv and EVI is slightly better and mostly due to the MAE ($0.86 \text{ gC m}^{-2} \text{ d}^{-1}$ for Bartlett and $1.02 \text{ gC m}^{-2} \text{ d}^{-1}$ for Michigan) and RMSE ($1.05 \text{ gC m}^{-2} \text{ d}^{-1}$ for Bartlett and $1.29 \text{ gC m}^{-2} \text{ d}^{-1}$ for Michigan) results.

The NDVI displays relatively diminished performance, indicating a 9% and 15% reduction in the R^2 compared to CCI in the Bartlett and Michigan sites, respectively. In contrast, EVI exhibits less favourable predictive results ($R^2 = 0.75$, $\text{MAE} = 1.92 \text{ gC m}^{-2} \text{ d}^{-1}$, $\text{RMSE} = 2.54 \text{ gC m}^{-2} \text{ d}^{-1}$) for the Borden site and the aggregated sites category, although in the specific context of the Borden site, its performance aligns closely with NDVI and NIRv. Among the individual sites, Bartlett has the most favourable predictive outcomes in terms of R^2 , MAE, and RMSE for every individual VI model, while the aggregated sites category yields the least favourable predictive results in the same scenarios.

In the case of the weekly models, NDVI records the least favourable results in terms of R^2 , MAE, and RMSE when assessed at the three individual sites, while EVI demonstrated its weakest predictive capabilities when all sites were treated as a single entity. However, for this case, EVI shows marginal differences for individual indices such as NDVI and NIRv ($R^2 =$

0.01 variance explained and no more than 0.02 $\text{gC m}^{-2} \text{d}^{-1}$ in RMSE). The best performing individual indices were CCI in terms of R^2 , MAE, and RMSE for Borden and the aggregated sites, NIRv for Michigan and EVI for Bartlett. Nonetheless, those superior performances are subtle when compared with the other individual indices. On a weekly basis, differences in variability explained between the best and least performing models range from $R^2 = 0.11$ to $R^2 = 0.2$. Bartlett and Michigan sites consistently yield the most accurate predictive models.

On a daily basis, CCI outperforms the rest of the individual indices in 3 cases: for Bartlett by $R^2 = 0.02$ in variance explanation compared with EVI, for Borden by $R^2 = 0.01$ compared with EVI, and for the combined sites dataset by $R^2 = 0.1$. Nonetheless it is worth mentioning that the variance explained in any of the models by individual indices in Borden or the combined site is less than $R^2 = 0.5$ in most of the cases and the error scores are the highest. In the case of Michigan, the best performing individual index was EVI which outperformed the next best-performing individual index NIRv by $R^2 = 0.03$. Generally, the Bartlett and Michigan sites consistently yielded the most accurate predictive models across various configurations. Conversely, the Borden site consistently exhibited the poorest model performance across all scenarios.

Overall, when comparing individual indices, CCI consistently performed better across different timeframes in terms of variance explainability and error metrics. Nonetheless, those differences are subtle compared to EVI and NIRv. In contrast, NDVI consistently performs less mentioning across all evaluated timeframes. Notably, models based on monthly values consistently exhibit better performance than those based on weekly or daily values as illustrated in Figure 2.6.

2.3.2. Analysis of GPP-Vegetation Index Relationships Using GAM Models

In Table 2.5, we present a summary of the results obtained from the GAM models used for GPP estimation at each site, employing the vegetation indices as predictors. To compare between the models, we include in the table relevant model summary statistics, such as R^2 , MAE, and RMSE. Furthermore, additional metrics such as the p-value, the F statistic (f), effective degrees of freedom (edf), and the Akaike Information Criterion (AIC), are displayed in Table A.10, and Table A.11 for daily models, Table A.8, and Table A.9 for weekly outputs, and for monthly results the Table A.6, and Table A.7. A residuals distribution for each of the models is in Figure A.2

Table 2.4.: Summary of Linear models for GPP estimation using the vegetation indices on a monthly (a), weekly (b), and daily (c) basis. MAE and RMSE metrics units are $\text{gC m}^{-2} \text{d}^{-1}$

(a)

| Site | EVI | | | NDVI | | | NIR _v | | | CCI | | | All | | |
|----------|------|------|------|------|------|------|------------------|------|------|------|------|------|------|------|------|
| | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE |
| Bartlett | 0.89 | 0.92 | 1.08 | 0.80 | 1.23 | 1.44 | 0.89 | 0.89 | 1.08 | 0.89 | 0.86 | 1.05 | 0.92 | 0.72 | 0.93 |
| Michigan | 0.86 | 1.05 | 1.35 | 0.72 | 1.62 | 1.92 | 0.86 | 1.05 | 1.33 | 0.87 | 1.02 | 1.29 | 0.94 | 0.72 | 0.90 |
| Borden | 0.75 | 1.92 | 2.54 | 0.76 | 1.95 | 2.46 | 0.75 | 1.90 | 2.50 | 0.79 | 1.82 | 2.32 | 0.80 | 1.72 | 2.24 |
| All | 0.57 | 2.21 | 2.87 | 0.64 | 1.95 | 2.62 | 0.59 | 2.18 | 2.81 | 0.74 | 1.69 | 2.26 | 0.78 | 1.63 | 2.04 |

(b)

| Site | EVI | | | NDVI | | | NIR _v | | | CCI | | | All | | |
|----------|------|------|------|------|------|------|------------------|------|------|------|------|------|------|------|------|
| | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE |
| Bartlett | 0.79 | 1.16 | 1.54 | 0.64 | 1.62 | 2.03 | 0.77 | 1.21 | 1.61 | 0.79 | 1.23 | 1.56 | 0.82 | 1.06 | 1.43 |
| Michigan | 0.78 | 1.50 | 1.79 | 0.61 | 1.99 | 2.39 | 0.78 | 1.50 | 1.78 | 0.71 | 1.72 | 2.09 | 0.81 | 1.43 | 1.67 |
| Borden | 0.57 | 2.62 | 3.54 | 0.50 | 2.91 | 3.83 | 0.57 | 2.64 | 3.58 | 0.58 | 2.62 | 3.52 | 0.61 | 2.53 | 3.39 |
| All | 0.45 | 2.65 | 3.53 | 0.46 | 2.64 | 3.50 | 0.46 | 2.61 | 3.51 | 0.56 | 2.35 | 3.16 | 0.56 | 2.34 | 3.15 |

(c)

| Site | EVI | | | NDVI | | | NIR _v | | | CCI | | | All | | |
|----------|------|------|------|------|------|------|------------------|------|------|------|------|------|------|------|------|
| | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE |
| Bartlett | 0.81 | 1.42 | 1.98 | 0.66 | 2.01 | 2.68 | 0.80 | 1.44 | 2.05 | 0.83 | 1.42 | 1.88 | 0.87 | 1.20 | 1.67 |
| Michigan | 0.70 | 1.87 | 2.33 | 0.52 | 2.31 | 2.96 | 0.67 | 1.92 | 2.48 | 0.61 | 2.06 | 2.67 | 0.72 | 1.81 | 2.27 |
| Borden | 0.43 | 3.42 | 4.48 | 0.28 | 3.88 | 5.04 | 0.38 | 3.57 | 4.66 | 0.44 | 3.31 | 4.42 | 0.53 | 3.07 | 4.08 |
| All | 0.49 | 3.07 | 4.12 | 0.45 | 3.22 | 4.30 | 0.48 | 3.14 | 4.19 | 0.59 | 2.71 | 3.70 | 0.61 | 2.65 | 3.63 |

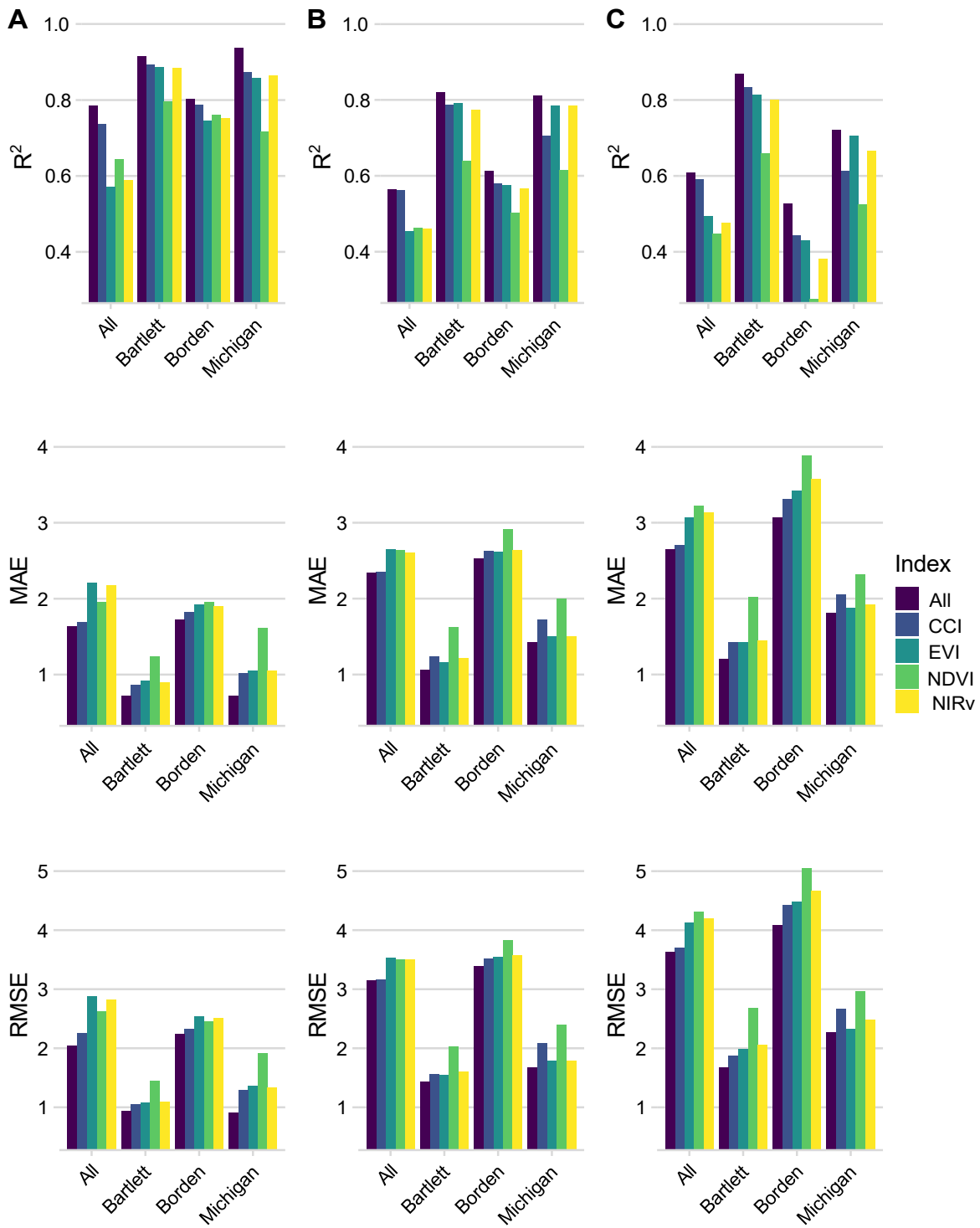


Figure 2.6.: Summary of Linear models for GPP estimation using the vegetation indices on a monthly (a), weekly (b), and daily (c) basis. MAE and RMSE metrics units are $\text{gC m}^2 \text{d}^{-1}$

When using GAM models on a monthly basis, no single VI demonstrates consistent superiority over the others. For the all sites category, CCI has an $R^2 = 0.03$ better variance explanation than NDVI, which is the second best model with an $R^2 = 0.75$. In the case of Michigan EVI and NIRv were the best individual indices with an $R^2 = 0.96$ of the variance in GPP, but EVI had slightly lower error metrics values in MAE ($0.01 \text{ gC m}^{-2} \text{ d}^{-1}$) and RMSE ($0.02 \text{ gC m}^{-2} \text{ d}^{-1}$). Bartlett had a better model performance when using CCI with an $R^2 = 0.91$ of variance explained with the lowest error metrics among all the GAM monthly models. It's important to highlight that for Michigan and Bartlett, implementing a GAM model using all the VIs as covariates posed challenges due to limited observations for the model parameters, raising concerns of potential overfitting.

Finally, it's noteworthy that NDVI on a monthly basis displayed suboptimal performance in two of the sites (Michigan and Bartlett) with a difference of $R^2 = 0.1$ and $R^2 = 0.13$ with the best performing models respectively, while EVI exhibited less favourable results in just the Borden site with MAE $1.92 \text{ gC m}^{-2} \text{ d}^{-1}$) and RMSE $2.54 \text{ gC m}^{-2} \text{ d}^{-1}$).

On a weekly basis, models incorporating all VIs as covariates consistently obtained better performance compared to any individual VI, irrespective of the site. Specifically when evaluating the all sites category, the inclusion of all indices as covariates yielded an $R^2 = 0.07$ increase in variance explanation compared with the best individual VI result CCI. Further, for Michigan, this improvement amounted to $R^2 = 0.02$ compared to EVI, $R^2 = 0.01$ for Bartlett in contrast to EVI, and an $R^2 = 0.02$ enhancement for Borden when compared with EVI.

Conversely, NDVI showed a diminished performance as an individual index when compared with all the other individual VIs. In the context of all sites, NDVI yielded $R^2 = 0.07$ less variance explanation than CCI. Notably, for Michigan, NDVI's performance lagged by $R^2 = 0.17$ compared to EVI, Bartlett an $R^2 = 0.18$ reduction relative to EVI, and Borden exhibited an $R^2 = 0.09$ deficit when contrasted with EVI.

On a daily basis, when considering the all sites category, the model with all the VIs as covariates explained $R^2 = 0.04$ more variance in GPP when compared with the best performing individual VI CCI. In the case of Bartlett, the increase was also $R^2 = 0.04$ but in this case, the best individual performing VI was EVI. For Michigan, the model using all VIs as covariates outperformed EVI by $R^2 = 0.05$ in GPP prediction, and for Borden, it was an $R^2 = 0.09$ improvement when compared with NIRv.

Table 2.5.: Summary of GAM models for GPP estimation using the vegetation indices on a monthly (a), weekly (b), and daily (c) basis. MAE and RMSE metrics units are $\text{gC m}^{-2} \text{d}^{-1}$

(a)

| Site | EVI | | | NDVI | | | NIR _v | | | CCI | | | All | | |
|----------|------|------|------|------|------|------|------------------|------|------|------|------|------|------|------|------|
| | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE |
| All | 0.66 | 1.91 | 2.50 | 0.72 | 1.72 | 2.27 | 0.67 | 1.89 | 2.47 | 0.75 | 1.60 | 2.16 | 0.79 | 1.47 | 1.91 |
| Michigan | 0.96 | 0.55 | 0.64 | 0.86 | 0.91 | 1.23 | 0.96 | 0.56 | 0.66 | 0.91 | 0.79 | 0.97 | NA | NA | NA |
| Bartlett | 0.88 | 0.92 | 1.08 | 0.78 | 1.23 | 1.44 | 0.88 | 0.89 | 1.08 | 0.91 | 0.67 | 0.87 | NA | NA | NA |
| Borden | 0.74 | 1.92 | 2.54 | 0.77 | 1.83 | 2.37 | 0.75 | 1.90 | 2.50 | 0.78 | 1.80 | 2.30 | 0.78 | 1.72 | 2.24 |

(b)

| Site | EVI | | | NDVI | | | NIR _v | | | CCI | | | All | | |
|----------|------|------|------|------|------|------|------------------|------|------|------|------|------|------|------|------|
| | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE |
| All | 0.53 | 2.35 | 3.25 | 0.51 | 2.40 | 3.34 | 0.52 | 2.37 | 3.29 | 0.58 | 2.17 | 3.06 | 0.59 | 2.17 | 3.02 |
| Michigan | 0.84 | 1.20 | 1.49 | 0.67 | 1.61 | 2.16 | 0.82 | 1.30 | 1.61 | 0.75 | 1.46 | 1.88 | 0.86 | 1.05 | 1.33 |
| Bartlett | 0.81 | 1.03 | 1.45 | 0.63 | 1.62 | 2.03 | 0.79 | 1.10 | 1.51 | 0.79 | 1.13 | 1.50 | 0.82 | 0.98 | 1.39 |
| Borden | 0.59 | 2.51 | 3.41 | 0.50 | 2.90 | 3.82 | 0.58 | 2.57 | 3.49 | 0.58 | 2.62 | 3.52 | 0.61 | 2.47 | 3.35 |

(c)

| Site | EVI | | | NDVI | | | NIR _v | | | CCI | | | All | | |
|----------|------|------|------|------|------|------|------------------|------|------|------|------|------|------|------|------|
| | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE | R2 | MAE | RMSE |
| All | 0.58 | 2.70 | 3.73 | 0.55 | 2.79 | 3.87 | 0.56 | 2.74 | 3.83 | 0.61 | 2.55 | 3.60 | 0.65 | 2.39 | 3.37 |
| Bartlett | 0.85 | 1.22 | 1.79 | 0.74 | 1.63 | 2.30 | 0.83 | 1.25 | 1.85 | 0.85 | 1.26 | 1.76 | 0.89 | 1.04 | 1.51 |
| Michigan | 0.74 | 1.67 | 2.13 | 0.58 | 2.01 | 2.73 | 0.67 | 1.84 | 2.43 | 0.64 | 1.92 | 2.55 | 0.79 | 1.45 | 1.92 |
| Borden | 0.50 | 3.22 | 4.17 | 0.42 | 3.41 | 4.47 | 0.46 | 3.33 | 4.33 | 0.45 | 3.27 | 4.36 | 0.55 | 2.94 | 3.92 |

Among the individual VIs, NDVI consistently demonstrated the poorest performance across all four cases. As an individual VI, EVI performed better in Bartlett, Michigan, and Borden. However is worth noting that for Borden the variance explained was limited to $R^2 = 0.5$.

In summary, among the three individual sites, Bartlett consistently produced the most favourable results in terms of models explaining variance and yielding lower residuals, followed by Michigan. In the case of Borden, when employing individual indices, the models struggled to achieve a variance explanation exceeding $R^2 = 0.5$.

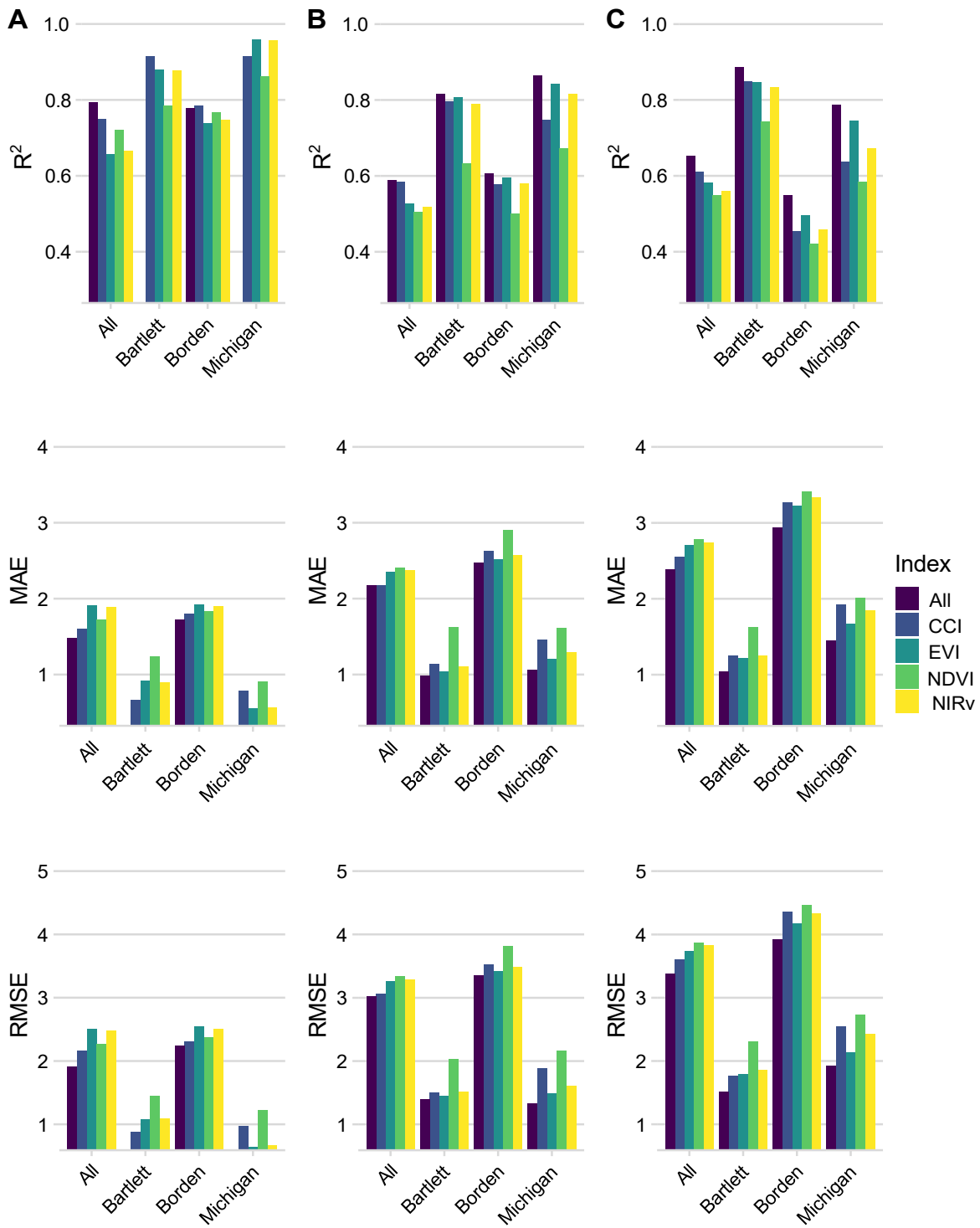


Figure 2.7.: Summary of GAM models for GPP ($\text{gC m}^{-2} \text{d}^{-1}$) estimation using the vegetation indices. Column A represents the metrics for the monthly models, B the weekly, and C the daily metrics. MAE and RMSE metrics units are $\text{gC m}^{-2} \text{d}^{-1}$

2.3.3. LM vs GAM

Monthly GAM applied in the context of Michigan exhibited a better performance when employing EVI and NIRv in comparison to LMs with all vegetation indices as covariates. Regarding weekly GAM models, they demonstrated improved performance compared to LM across all sites, except for Borden, where both models yielded equivalent R^2 values. Notably, despite similar R^2 values, the GAM model consistently displayed lower error metrics, suggesting its capacity to better accommodate potential nonlinear relationships and produce more accurate predictions. This adaptability was particularly evident when addressing the daily, weekly, or monthly variations in GPP. Furthermore, it is crucial to highlight that monthly models, both in LM and GAM frameworks, consistently exhibited superior metrics compared to their weekly and daily counterparts, as illustrated in Figure 2.6 and Figure 2.7. This can be an effect of reducing variance when summarizing values as it is shown in the daily Figure 2.8, weekly Figure 2.9, and monthly Figure 2.10

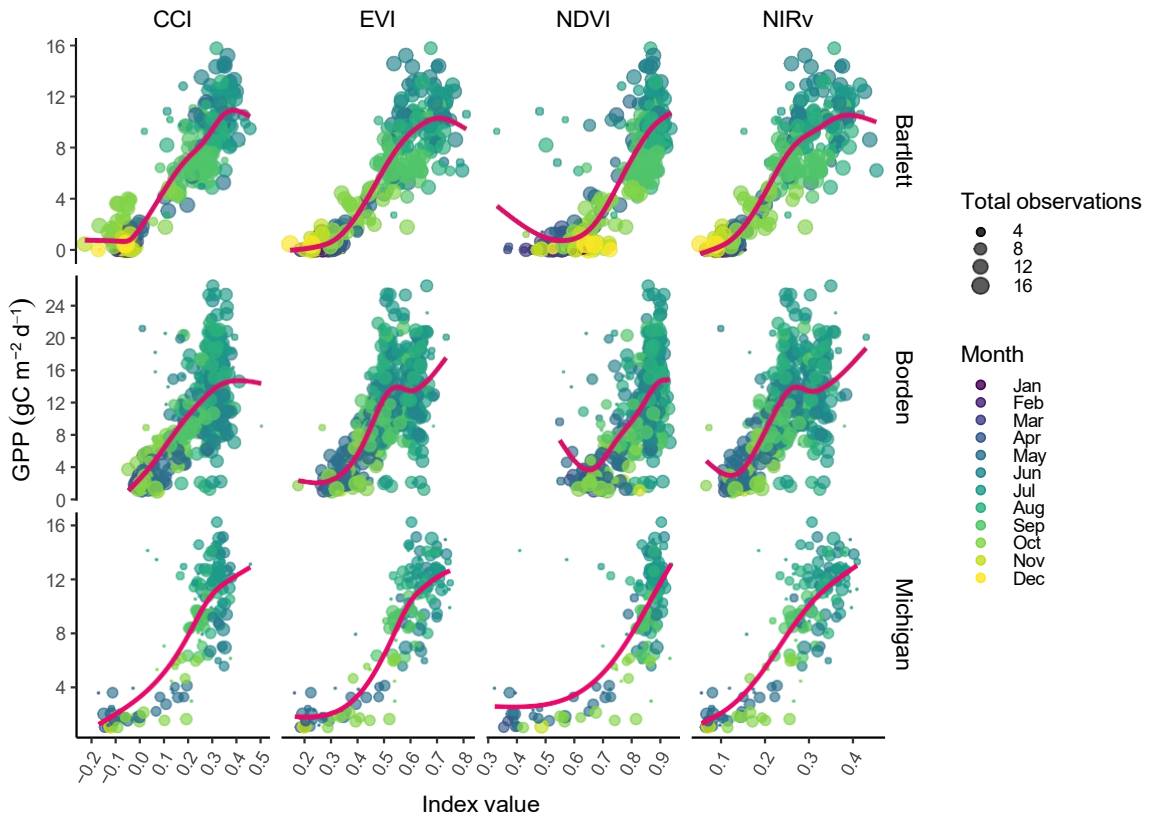


Figure 2.8.: Scatterplot of MODIS 500m derived VIs and GPP with daily values. Every observation corresponds to the observed GPP from a flux tower site. Total observations corresponds to the number of observations used to obtain the mean of the vegetation index (NDVI, NIRv, CCI and EVI). The red line indicates the GAM fit.

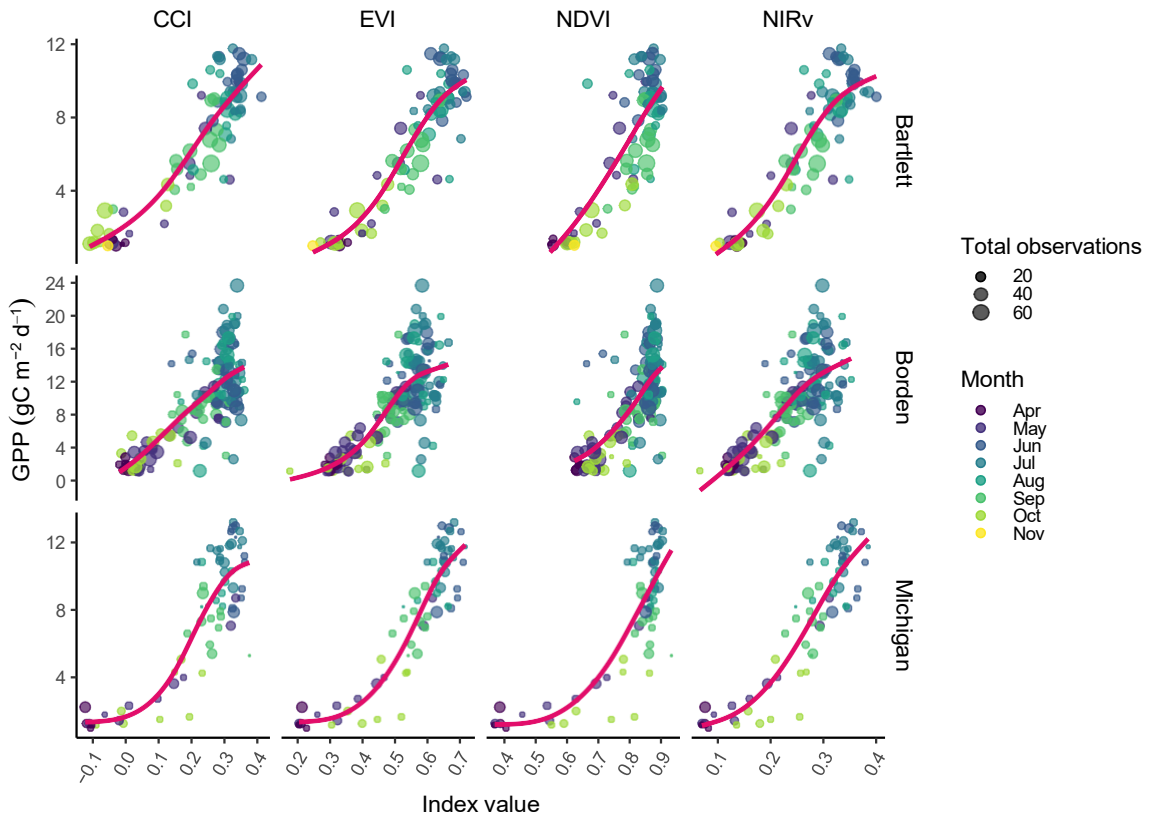


Figure 2.9.: Scatterplot of MODIS 500m derived VIs and GPP with weekly values. Every observation corresponds to the observed GPP from a flux tower site. Total observations corresponds to the number of observations used to obtain the mean of the vegetation index (NDVI, NIRv, CCI and EVI). The red line indicates the GAM fit.

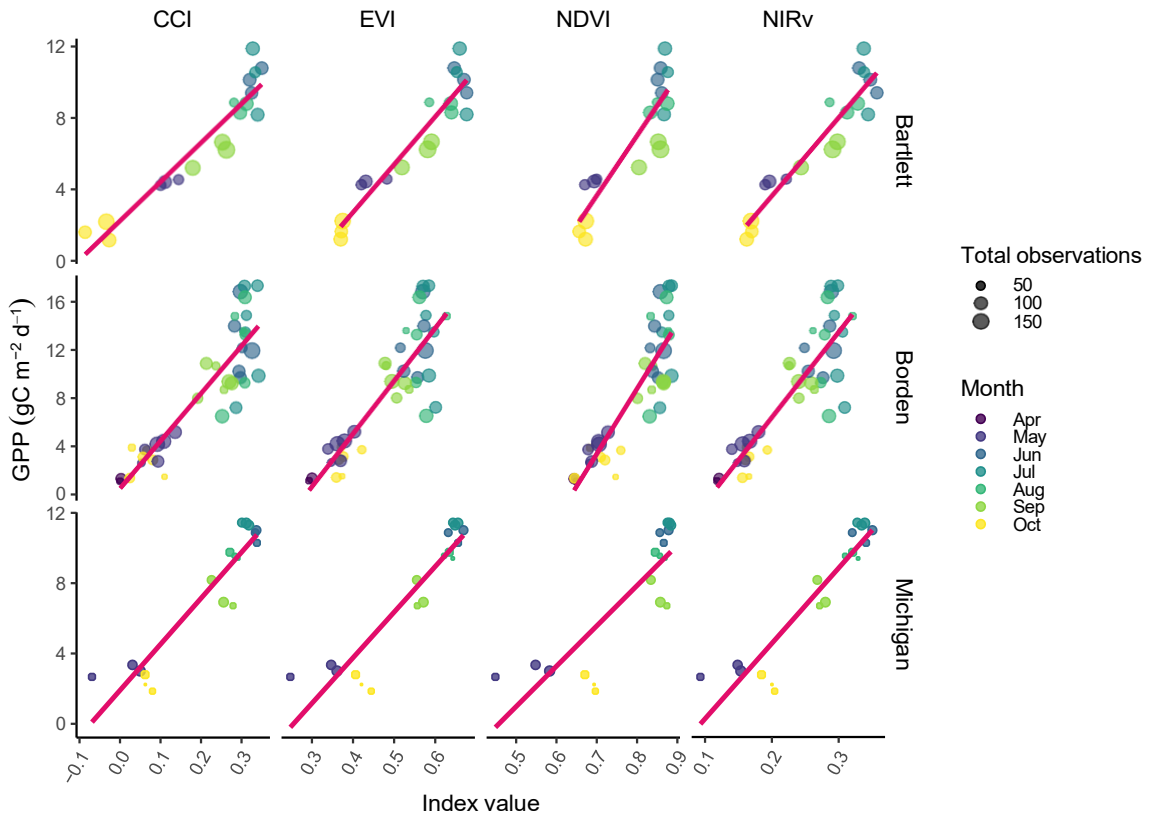


Figure 2.10.: Scatterplot of MODIS 500m derived VIs and GPP with monthly values. Every observation corresponds to the observed GPP from a flux tower site. Total observations corresponds to the number of observations used to obtain the mean of the vegetation index (NDVI, NIRv, CCI and EVI). The red line indicates the GAM fit.

2.4. Discussion

In summary, our LM results demonstrate that incorporating all VIs as covariates in the model enhances predictive accuracy for GPP compared to models using single VIs. This holds true across the three temporal scales and for each of the sites individually or when considered as part of the same biotype. This observation suggests that the relationship is really non-linear, and combining different VIs allows to use of an LM to capture non-linear patterns throughout the entire study duration, contrasting with the predictive power of any individual VI. Additionally, VIs exhibit different noise sensitivities (Zeng et al. 2022), which explains why using a single VI is insufficient to capture nuanced variations in GPP.

Conversely, when aggregating all study sites indiscriminately to represent a unified ecosystem type, model metrics exhibit diminished performance compared to employing distinct models for each individual site. This decline in performance may be attributed to the introduction of variability inherent in the Borden site. The Borden site presents larger ranges of GPP values (see Figure Figure 2.2) than the Bartlett and Michigan sites. Incorporating the Borden values together with those from the Bartlett and Michigan sites to represent a single ecosystem type in a model introduces higher GPP values that contribute to variability. This variability cannot be accurately tracked by the vegetation indices, resulting in a reduction in RMSE and R².

If such discrepancies arise when constructing models with identical specifications for various sites categorized under the same biotype ecosystem, it may indicate that, for global GPP models, exploring the use of diverse VIs is crucial. An approach like this one could open the possibility for a more accurate estimation of GPP by leveraging specific indices that might yield superior results for each distinct type of ecosystem potentially contributing to more robust and accurate global GPP estimations. For example, the study by (Lin et al. 2019) showed that Chlorophyll Index Red (CRI) had a betterThis hypothesis could be further elucidated through standardized processes for calculating GPP with in-situ data across all sites with flux towers available; a task currently in development by FLUXNET (Pastorello et al. 2020). Standardized GPP values obtained through this process could then be used to train models and assess which VIs have the lowest errors when predicting GPP.

Considering the availability of data with daily, weekly, and monthly values, a pertinent question arises about the temporal scale's impact on GPP estimation. Our results indicate that models based on monthly data demonstrated better model fit for each model and smaller resid-

uals metrics. This superiority may be attributed to reduced data variation, as daily values are aggregated into monthly summaries, leading to values more centralized around a mean which can reduce the measurement noise and concurrently mitigate the saturation effect of VIs for daily peaks. The performance of model fits and residuals metrics declined when transitioning from monthly to weekly and, subsequently, to daily values (See Figure A.1 for LM residuals and Figure A.2 for GAM residuals). Although the decrease in model fit was not substantial, the increase in residuals metrics suggests a potential for larger errors in predictions.

While NDVI is one of the most used VIs in EO (Pabon-Moreno et al. 2022), its performance, as measured by the GPP RMSE, was comparatively less favorable on two out of the three sites when compared to other indices. However, the performance variation among the sites was not significant, with the Bartlett experimental forest showing slightly better results in all GPP ~ VIs relationships compared to the other sites. These findings suggest that, although NDVI may not be the most optimal vegetation index for GPP estimation at these specific sites, variations in performance across different locations are still noteworthy.

When examining the data trends for GPP at each site and its relationship with VIs, it becomes evident that at higher GPP values, the dispersion of the data is more pronounced. This observed pattern may be ascribed to the inherent limitation of these indices, which predominantly track the presence of green leaves. It is important to note, however, that the mere presence of green leaves does not consistently signify active photosynthesis. This discrepancy can arise for two reasons. Firstly, the photosynthetic process may undergo temporary suspension without any manifest changes in chlorophyll content or leaf abscission (Camps-Valls et al. 2021) Secondly, VIs relying on the NIR band face challenges in detecting photosynthesis in instances when a higher amount of healthy plant biomass is present. The increased biomass leads to greater scattering and reflection of NIR radiation, resulting in a saturation effect (Camps-Valls et al. 2021).

This observation holds significance as, despite VIs serving as indicators of vegetation capacity rather than vegetation physiology, historical data records are predominantly derived from sensors equipped with bands for generating these VIs. Conversely, newer sensors capable of capturing additional bands to derive indices such as Solar-Induced Fluorescence (SIF) lack comprehensive historical records, hindering the feasibility of long-term studies that extend back to years preceding the 2000s.

The application of both LM and GAM models to the dataset revealed some differences in

model performance. The GAM model exhibited a slightly superior performance in capturing the underlying patterns within the weekly and daily data compared to the LM counterpart, but not in the monthly data. This was evident from assessments of model fit and residuals metrics. The GAM model's ability to flexibly capture non-linear relationships allowed for a more accurate representation of the complex structure inherent in the data. Consequently, the results suggest that the GAM framework may be more suitable for capturing the nuances present in the dataset, emphasizing the importance of considering model flexibility when analyzing non-linear relationships.

It is worth noticing that using a GAM model with only CCI as a predictor, performs almost as well as the GAM model with all the VIs as covariates. This could imply that the limitation in using a single VI with a LM arises from its inability to capture the non-linear nature of the relationship. A selection of a proficient VI with a non-linear model could better estimate GPP without the need to create models with multiple predictors. In this scenario, the GAM model with CCI to estimate GPP emerges as a potentially adequate choice among generalized linear fits. However, this assessment does not address the question of whether combining diverse models could offer more robustness, as such an approach might implicitly incorporate spatial variability to a certain degree.

2.5. Conclusions

In conclusion, our analysis demonstrates that incorporating all VIs as covariates in our models yields a substantial improvement in predictive accuracy for GPP compared to using any single VI. Additionally, our research highlights the impact of time aggregation on prediction accuracy across different models. Monthly LM models exhibit the best performance metrics, while weekly and daily LM models present lower metrics, attributable to higher variability in observations that makes tracking GPP challenging. A similar pattern is observed when using GAM; however, weekly and daily GAM models outperform LM models.

Notably, no single VI emerges as the universal best predictor for every site or time aggregation, but CCI with a GAM model emerges as a potentially adequate choice among generalized linear fits. Moreover, the impact of large variations in GPP ranges within a site is evident in the quality of predictions. These variations introduce complexities and uncertainties, emphasizing the necessity of accounting for local site characteristics and inherent heterogeneity when

aiming for accurate predictions of GPP. Growing seasons exhibit the most pronounced variability, posing challenges for vegetation indices that saturate at high biomass concentrations, making it more difficult to track changes in GPP.

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3. A data driven approach to predict GPP from VIs through machine learning methods

3.1. Introduction

The field of Earth Science has witnessed a transformative shift with the integration of Machine Learning (ML) methods, which has led to a deeper understanding of our planet's complex ecosystems and processes (Reichstein et al. 2019). ML methods are now well established in environmental sciences (Lary et al. 2016), including their use within studies aimed at mapping and quantifying vegetation characteristics and ecological processes such as vegetation cover, structure, and disturbances, among others (Lehnert et al. 2015; Jochem Verrelst et al. 2012).

The increasing number of EC sites (Tramontana et al. 2016) coupled with the continuously growing amount of Earth system data surpassing dozens of petabytes (Reichstein et al. 2019), has led to an emergence of purely data-driven methodologies for quantifying ecosystem status and fluxes. These approaches have shown promise for the quantification of global terrestrial photosynthesis (Martin Jung et al. 2011; Tramontana et al. 2016) and have resulted in good progress in the estimation of biogeo-physical parameters using remotely sensed reflectance data, both at local and global scales (Coops et al. 2003; J. Verrelst et al. 2012).

Furthermore, these data-driven approaches have contributed significantly to the scientific community by providing spatial, seasonal, and interannual variations in predicted fluxes. These predictions, generated through machine learning methodologies, are now serving as important benchmarks for evaluating the performance of physical land-surface and climate models (Martin Jung et al. 2010; Bonan et al. 2011; Anav et al. 2015).

Some of the differences between data driven models and process-based methods are the inherent observational character of data driven models and that functional relationships emerge

from the patterns found in the data, rather than being stated before (Tramontana et al. 2016). So functional relationships between in-situ measured fluxes with the explanatory variables can emerge (Tramontana et al. 2016). This paradigm shift toward data-driven modeling to extract patterns represents an opportunity to come up with new ideas and question established theories in earth system models.

In contrast to process based models, data driven models inherently possess an observational nature where functional relationships emerge from the patterns found in the data, rather than being predefined, such as the relationships between in-situ measured fluxes and the explanatory variables. (Tramontana et al. 2016). For example, the application of spatially explicit global data driven methods, has unveiled discrepancies in the estimation of photosynthesis within tropical rainforests when compared to climate models (Beer et al. 2010). This overestimation, has led to the creation of hypothesis for a better understanding of radiative transfer in vegetation canopies (Bonan et al. 2011) which can result in better photosynthesis estimates. This paradigm shift toward data-driven modeling to extract patterns, represents an opportunity to explore novel ideas and question established theories in earth system models (Reichstein et al. 2019).

Predicting dynamics in the biosphere is challenging due to biologically mediated processes (Reichstein et al. 2019). The term “prediction” should not be confused with forecasting, as most models are not aiming at predicting into the future. Instead, the focus of these data-driven models is to improve historical estimates or enable the use of reflectance values to predict GPP when no in-situ data is available in the present times (Meyer et al. 2018).

These predictions can include many forms of uncertainty (Reichstein et al. 2019). One form is that individual ML methods can have different responses, especially when these models are applied beyond the conditions presented in the training dataset. (M. Jung, Reichstein, and Bondeau 2009; Papale et al. 2015). Another form is related to the explanatory variables used in ML methods derived from satellite remote sensing, which are partial in providing information about the vegetation state (Tramontana et al. 2016). Consequently, they lack the information required to explain the complete variability in fluxes (Tramontana et al. 2015). For instance, if a model is created using only reflectance data to estimate GPP without meteorological data, phenomena such as drought could lead to predicted values with large errors, given that stomata closure during water deficit have an immediate effect on the fluxes that can not be detected by the reflectance values only until later when the stress conditions

persist (Tramontana et al. 2015).

This chapter evaluates data driven approaches for estimating GPP given MODIS surface reflectance data for temperate broadleaf forests in North America. Two state-of-the-art approaches previously used for GPP estimation are tested: regression random forests (James et al. 2022) and AutoML (LeDell and Poirier 2020) to understand how the ML algorithm impacts prediction uncertainty. Both approaches were tested locally and using pooled data from three sites to quantify their ability to be applied over larger spatial extents.

3.2. Methods

3.2.1. Eddy Covariance sites

For this study, we selected three deciduous broadleaf forest sites: University of Michigan Biological Station located in northern Michigan, USA (45°35' N 84°43' W), Bartlett experimental forest in New Hampshire, USA (44°06' N, 71°3' W), and the Borden Forest Research Station (44°19' N, 79°56' W) in Ontario, Canada. These sites were selected to ensure they represented a single ecosystem type, characterized by shared environmental features. This approach allowed us to treat the dataset as a representation of a specific vegetation type terrestrial ecosystem.

In-situ data such as GPP was obtained utilizing the ONEFlux estimation processing by Ameriflux. Here, we selected GPP estimation done by the daytime method (Pastorello et al. 2020) on a daily, weekly, and monthly basis.

To capture seasonal variations and long-term trends, GPP data was collected over a minimum of 2 years. Specifically, University of Michigan Biological Station collected data spanned from January 2015 to January 2018, Bartlett experimental forest data ranges from January 2015 to December 2018, and Borden Forest Research Station from January 2015 to January 2022.

3.2.2. Satellite imagery

We used Google Earth Engine (GEE) to retrieve the Terra Moderate Resolution Imaging Spectroradiometer (MODIS), specifically the collection MOD09GA Version 6.1 daily 500m resolution surface reflectance products (MODIS/Terra Surface Reflectance Daily L2G Global 1 km and 500 m SIN Grid). A square polygon with an area of 3km surrounding the EC tower was defined for each of the study sites, and all daily data pixel values within this polygon were extracted for analysis.

MOD09GA contains the surface spectral reflectance from bands 1 through 7 (Table 3.1) with a spatial resolution of 500m, with corrections for atmospheric conditions such as aerosols, gasses, and Rayleigh scattering (Vermote 2021).

We selected the highest quality pixels according to the 1km Reflectance Data State QA (state_1km) (Table A.1) and Surface Reflectance 500m Quality Assurance (qc_500m) (Table A.2) variables. Once we had just the highest quality pixels, all the band values were scaled

by a factor of 0.0001. If any value fell outside the range of 0 to 1 after the scaling, it was discarded.

Once all the band values were scaled, we calculated 4 Vegetation Indices: NDVI (Equation 2.2), NIRv (Equation 2.3), EVI (Equation 2.4), and CCI (Equation 2.5). Then all the MODIS bands values and VIs were summarized on a daily, weekly, and monthly basis to be merged with the GPP values from ONEFlux.

Table 3.1.: MODIS (MOD09GA.061 product) bands used for ML methods

| Name | Description | Resolution | Wavelength |
|-------------|-------------|------------|-------------|
| sur_refl_01 | Red | 500 meters | 620-670nm |
| sur_refl_02 | NIR | 500 meters | 841-876nm |
| sur_refl_03 | Blue | 500 meters | 459-479nm |
| sur_refl_04 | Green | 500 meters | 545-565nm |
| sur_refl_05 | Red Edge | 500 meters | 1230-1250nm |
| sur_refl_06 | SWIR 1 | 500 meters | 1628-1652nm |
| sur_refl_07 | SWIR 2 | 500 meters | 2105-2155nm |

3.2.3. Random Forests

A random forest is an ensemble learning technique that leverages the power of multiple decision trees to improve predictive accuracy and robustness. It combines regression and classification trees, constructing each tree from random subsets of both trained data and features (James et al. 2022). The final prediction is obtained by aggregating individual tree predictions, which are then averaged to produce the ultimate estimate. Additionally, at each split, the best predictor from the random subset is selected to effectively partition the data and in the case of the regression random forest, the process is adapted to predict a continuous numeric outcome (James et al. 2022; Meyer et al. 2018).

Regression random forests were used as an approach to predict GPP as a function of all available MOD09GA bands values (from B01 to B07) and the calculated VIs. Three distinct models were developed, each one tailored for a specific time scale (daily, weekly, and monthly). This approach allowed us to assess the prediction performance of GPP at different time scales.

Each model was calibrated using a random data splitting procedure, dividing the data into a training set with 70% of the observations and a test set with the remaining 30%. Due to the varying number of observations for each site, to avoid an imbalanced training dataset, we employed a stratified data split to ensure a proportional representation of each site category in both sets. To ensure reproducibility, we used a consistent random number generator state throughout the process.

To implement the RF models, we use the ranger package in R (Wright and Ziegler 2017), utilizing 1000 trees within the forest ensemble. The models were trained using the bootstrap resampling technique with 100 folds, which helps to improve the robustness and accuracy of the predictions.

We calculated the variable of importance (VIP) to understand which MODIS bands or VIs are driving the predictions in each of the regression random forest models. To measure the influence of each feature on the overall model's predictive performance, we quantify how this performance deteriorates when a particular variable is permuted while keeping others constant.

To understand which features contributed the most on average to a particular GPP prediction in different coalitions (Molnar 2020), we calculated the Shapley values (Lundberg and Lee 2017). These values were computed with the DALEX package in R (Biecek 2018) for both low GPP and high GPP scenarios within each of the temporally aggregated models.

3.2.4. AutoML

The AutoML approach is designed to identify the most optimal ML pipeline for a specific problem and available training data by evaluating different combinations of data processing steps, ML models, and hyperparameters settings (Gaber et al. 2023). To create a stacked ensemble of models and evaluate the performance, we utilized the H2O AutoML framework (LeDell and Poirier 2020).

For the data training and evaluation, we randomly split the dataset into a training and a test set, allocating 80% and 20% of the observations, respectively. We excluded the total observations and site variables from both datasets, while the remaining features (bands and VIs), were considered predictors. The response variable for prediction was GPP. We transformed the data into an H2O frame to ensure compatibility with the AutoML process.

After generating machine learning models and hyperparameter configurations, we imposed a time constraint of a maximum of 2 minutes for model training. The best performing models were then selected in terms of lower predictions errors and used to form an ensemble. This trained ensemble model was used to generate predictions on the test set. Model performance was evaluated using metrics such as the coefficient of determination and the RMSE.

To assess the variable importance, we determined the influence of each variable on the model's predictions. The results of this analysis were visually represented as heatmaps for each of the models within the ensemble.

3.3. Results

3.3.1. Data-Driven GPP Prediction: A Regression Random Forest Approach

The best performing model, based on R^2 and RMSE was the monthly aggregated model, achieving a R^2 of 0.81 and an RMSE of $2.03 \text{ gC m}^{-2} \text{ d}^{-1}$ (see Figure 3.1). The daily model exhibited the second best performance model with a RMSE of $3.20 \text{ gC m}^{-2} \text{ d}^{-1}$ and a R^2 of 0.69 (see Figure 3.2). However, the weekly model displayed comparatively lower performance, with an R^2 of 0.56 and an RMSE of $3.23 \text{ gC m}^{-2} \text{ d}^{-1}$ (see Figure 3.3)

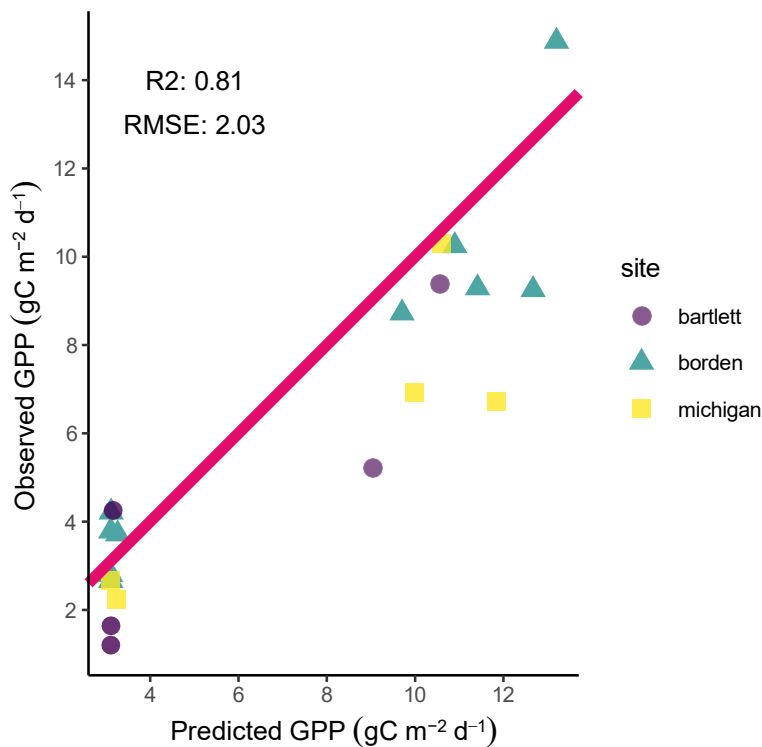


Figure 3.1.: GPP observed and predicted values from the Random Forest for all the sites at a monthly basis. The red line represents a 1:1 relation. Metrics units are $\text{gC m}^{-2} \text{ d}^{-1}$

Moreover, in assessing the importance of predictor variables within each model, VIs variables held top positions across all models, surpassing the importance of any other spectral bands alone. Specifically, the CCI contributed most significantly to the predictive performance in two of the models, the monthly and weekly models (Figure 3.4 and Figure 3.5). In the case of the daily model, CCI emerged as the second most influential predictor variable (as shown in Figure 3.6).

Despite NDVI being ranked as the second variable in importance for the monthly model, it

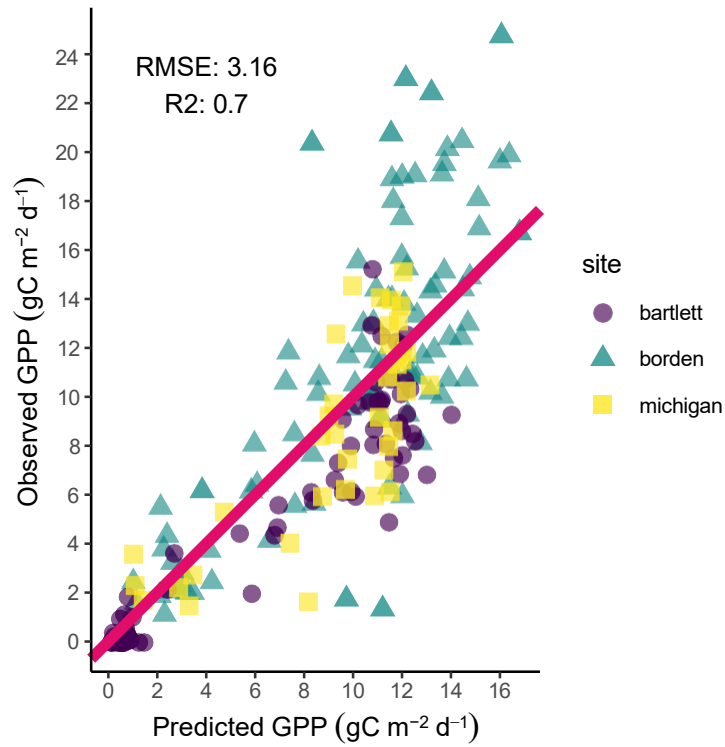


Figure 3.2.: GPP observed and predicted values from the Random Forest model for all the sites at a daily basis. The red line represents a 1:1 relation. Metrics units are gC m² d¹

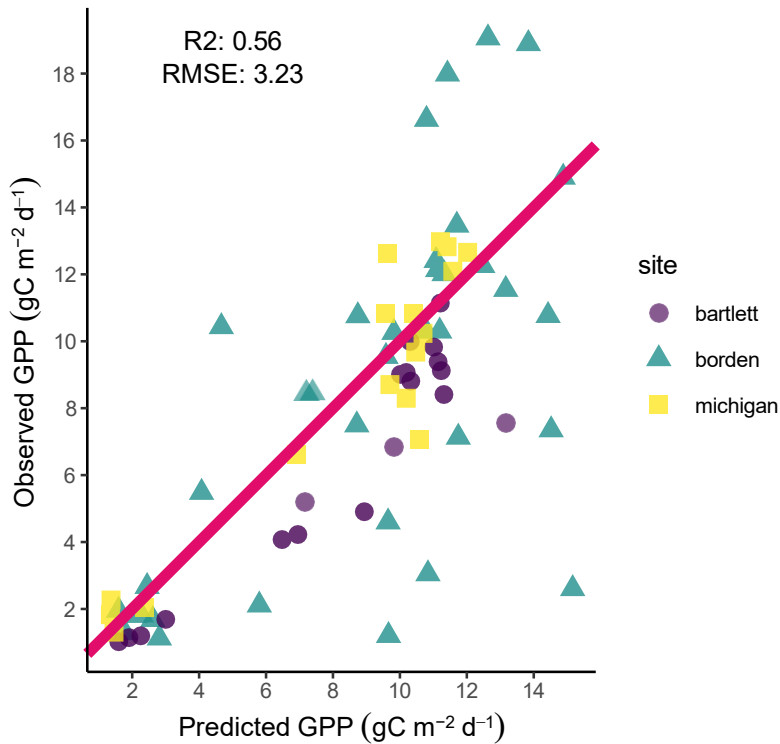


Figure 3.3.: GPP observed and predicted values from the Random Forest for all the sites at a weekly basis. The red line represents a 1:1 relation. Metrics units are gC m² d¹

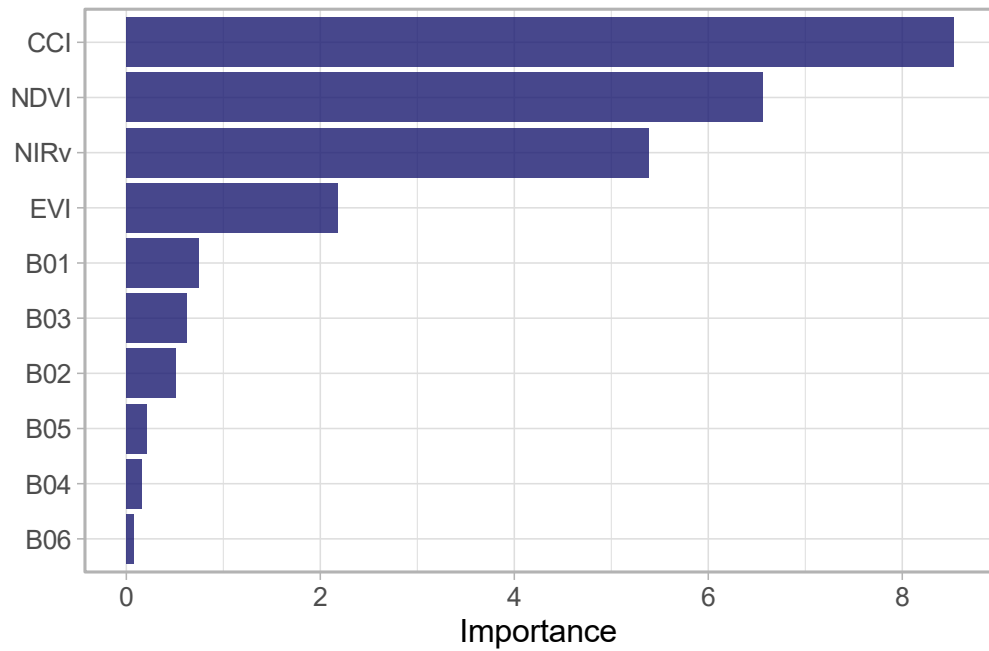


Figure 3.4.: Variable of importance derived from the Random forest model for the monthly values at 500 m spatial resolution model.

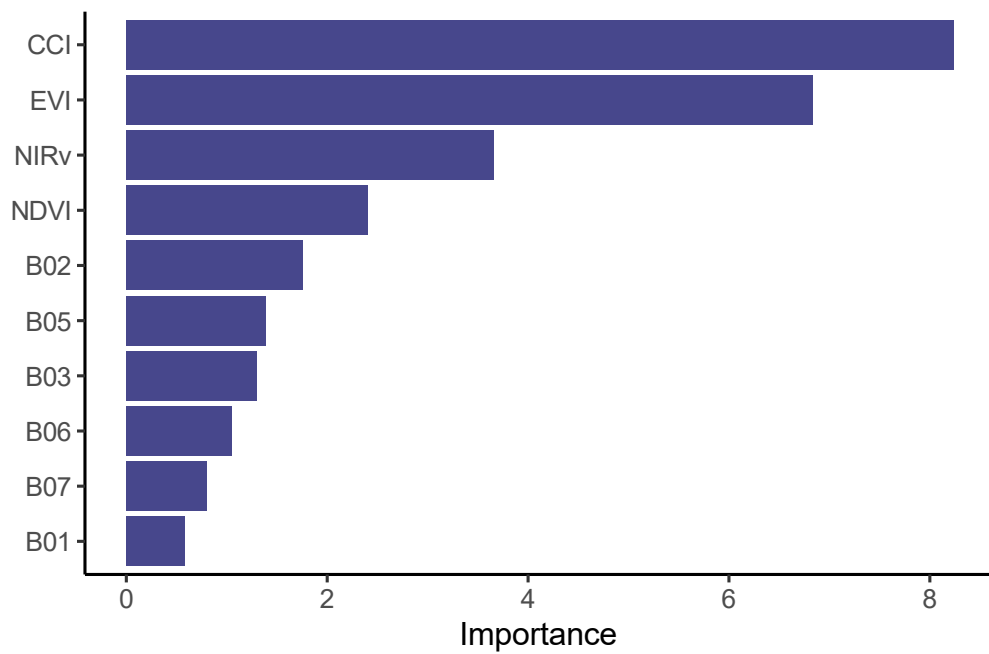


Figure 3.5.: Variable of importance derived from the Random forest model for the weekly values at 500 m spatial resolution model.

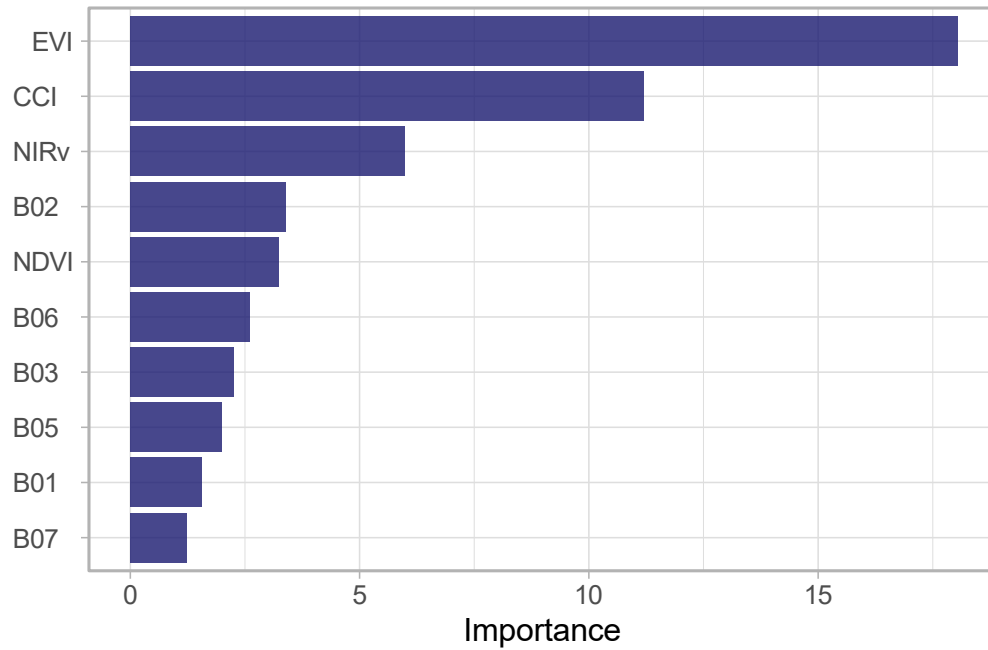


Figure 3.6.: Variable of importance derived from the Random forest model for the daily values at 500 m spatial resolution model.

was the fifth variable in importance for the daily model and the fourth one for the weekly model. Meanwhile, the NIRv consistently held the third most influential position across all models. EVI was found to be most important for the daily model, the second most important variable in the weekly model, but fourth in the monthly model.

To evaluate how the predictor variables had an impact on the model’s predictions, we employed the Shapley values. For each of the models created, we calculated the contributions to each variable in predicting GPP for two cases: for a known high GPP value and a known low GPP value from the test data set.

In the case of the daily model, we chose a low GPP value from the test dataset, specifically $0.01 \text{ gC m}^{-2} \text{ d}^{-1}$ and obtained a model prediction of $0.60 \text{ gC m}^{-2} \text{ d}^{-1}$. Our Shapley value analysis showed that EVI, CCI and NIRv were the most influential attributes affecting the model prediction (See Figure 3.7 A). These VIs contribute negatively to the prediction, reducing the predicted GPP.

Conversely, the high GPP value selected ($16.21 \text{ gC m}^{-2} \text{ d}^{-1}$) from the test dataset, had a model prediction of $12.70 \text{ gC m}^{-2} \text{ d}^{-1}$. In this case the Shapley values analysis indicated that the most influential variables were CCI and NIRv (See Figure 3.7 B). These variables contributed positively to the prediction.

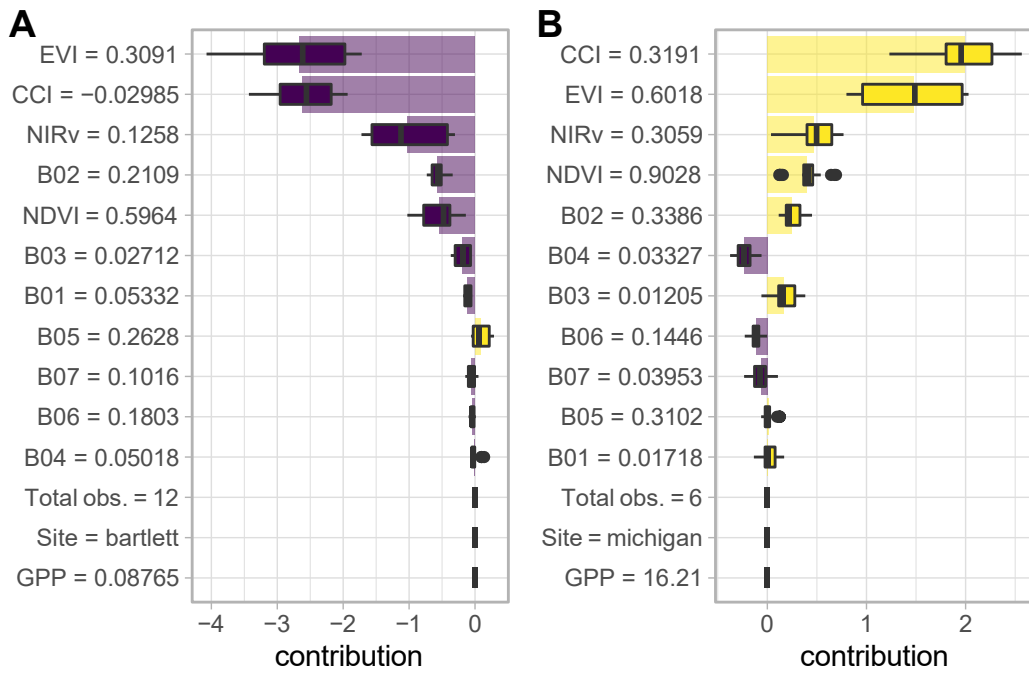


Figure 3.7.: Shapley values derived from the Random forest model for the daily values at 500 m spatial resolution model. Predicted value for the low GPP value is 0.59 (A) and 12.7 for the selected high GPP value (B).

For the weekly model, we selected high and low GPP values of $1.21 \text{ gC m}^{-2} \text{ d}^{-1}$ and $13.17 \text{ gC m}^{-2} \text{ d}^{-1}$, respectively, resulting in predictions of $1.60 \text{ gC m}^{-2} \text{ d}^{-1}$ and $11.0 \text{ gC m}^{-2} \text{ d}^{-1}$. For the low GPP value, the most influential variables were CCI, EVI, NIRv, and NDVI, all contributing negatively to the prediction (see Figure 3.8 A). Conversely, for the high GPP value, CCI, EVI, NIRv, B02 and NDVI had the most influence on the prediction positively (refer to Figure 3.8 B).

In the case of the monthly model, the low GPP value selected was $1.86 \text{ gC m}^{-2} \text{ d}^{-1}$ and the high GPP value was $11.87 \text{ gC m}^{-2} \text{ d}^{-1}$, leading to predictions of $3.31 \text{ gC m}^{-2} \text{ d}^{-1}$ and $10.60 \text{ gC m}^{-2} \text{ d}^{-1}$ respectively. In both scenarios, CCI, NDVI, and NIRv emerged as the most influential variables, albeit contributing negatively to the low GPP value prediction (See Figure 3.9 A) and positively to the high GPP value prediction (See Figure 3.9 B).

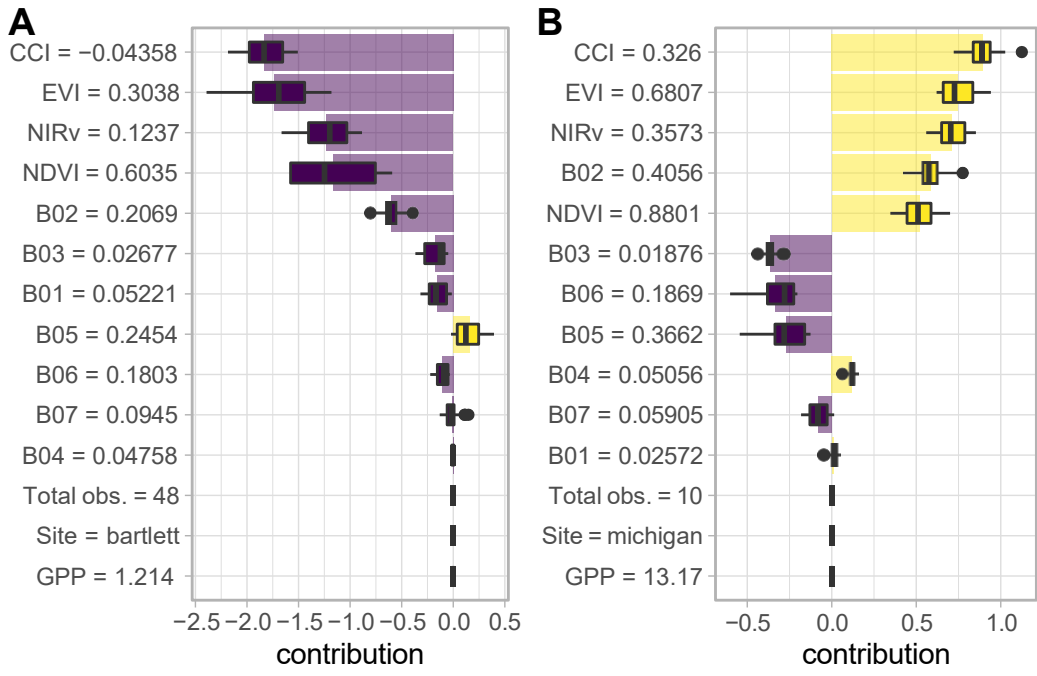


Figure 3.8.: Shapley values derived from the Random forest model for the weekly values at 500 m spatial resolution model. Predicted value for the low GPP value is 1.59 (A) and 11.0 for the selected high GPP value (B).

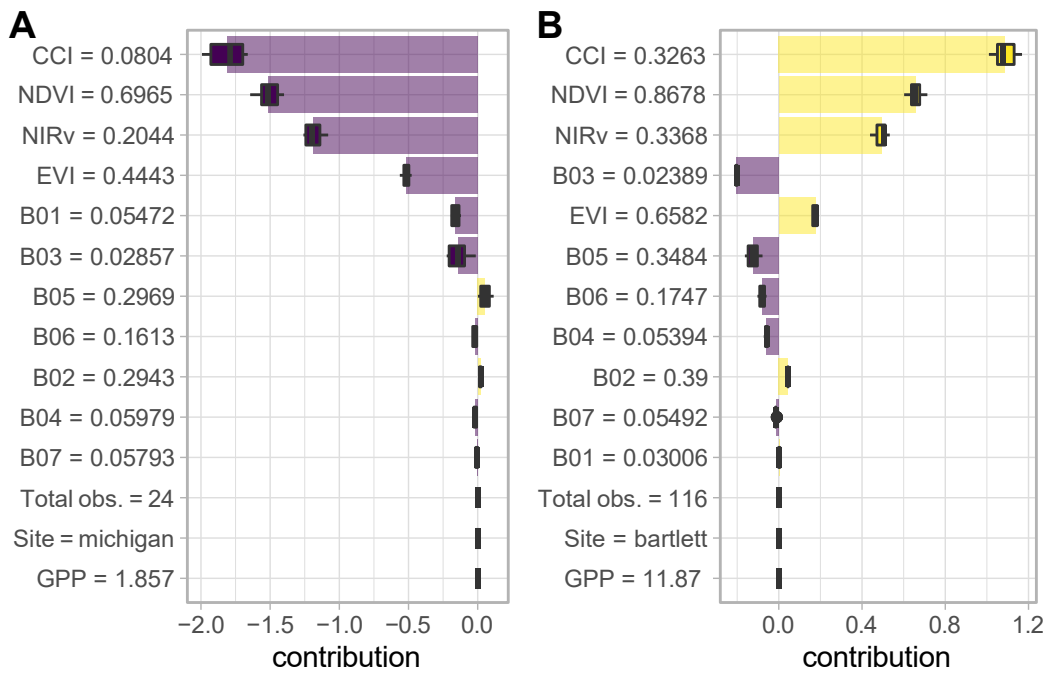


Figure 3.9.: Shapley values derived from the Random forest model for the monthly values at 500 m spatial resolution model. Predicted value for the low GPP value is 3.31 (A) and 10.6 for the selected high GPP value (B).

3.3.2. The potential of AutoML approaches for GPP predictions

The AutoML approach yielded varying performance outcomes across different temporal scales as measured by R^2 and RMSE. The monthly prediction model (Figure 3.10) emerged as the top performer, with an R^2 of 0.76 and a low RMSE of $1.84 \text{ gC m}^{-2} \text{ d}^{-1}$. Following this, the weekly model (Figure 3.11) demonstrated the second-highest explanatory power, capturing $R^2 = 0.72$ of the GPP variability, albeit with a slightly elevated RMSE of $3.08 \text{ gC m}^{-2} \text{ d}^{-1}$. Conversely, the daily model (Figure 3.12) exhibited comparatively diminished performance, explaining 0.67 of the variability with the highest RMSE of $3.11 \text{ gC m}^{-2} \text{ d}^{-1}$.

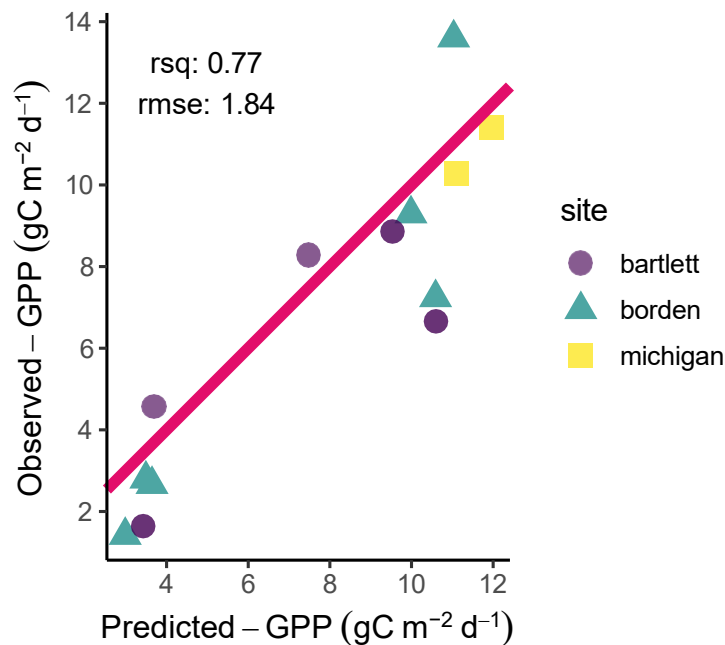


Figure 3.10.: GPP observed and predicted values from the autoML for all the sites at a monthly basis. The red line represents a 1:1 relation. Metrics units are $\text{gC m}^{-2} \text{ d}^{-1}$

An examination of variable importance in the AutoML model revealed distinctive patterns in significant contributors to GPP prediction across temporal scales. In the daily prediction model, EVI and CCI emerged as the most important (Figure 3.13). Moving to a weekly time frame, the importance of EVI and CCI persisted, with the addition of Band 02 (NIR band) (Figure 3.14). In the monthly prediction model, EVI once again took precedence, accompanied by Band 02 and NIRv (Figure 3.15), highlighting the enduring importance of these variables.

Additionally, a comparative assessment between the regression random forest model and the

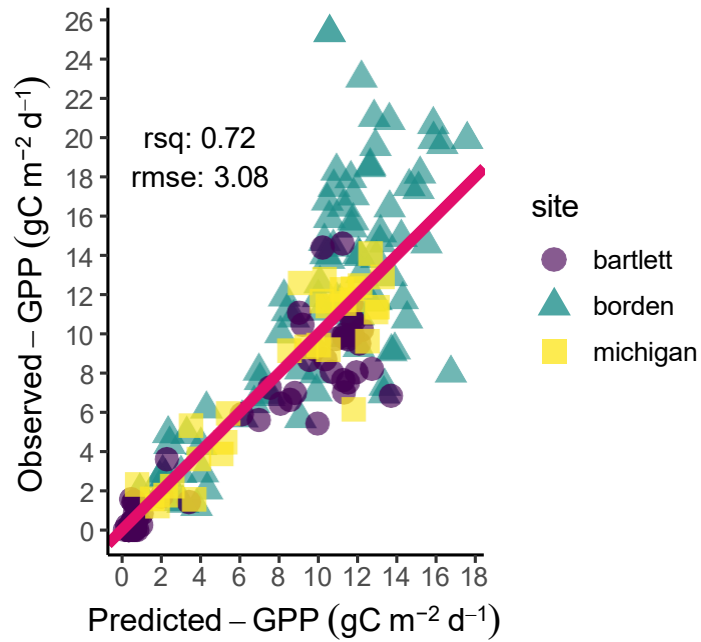


Figure 3.11.: GPP observed and predicted values from the autoML for all the sites at a weekly basis. The red line represents a 1:1 relation. Metrics units are $\text{gC m}^{-2} \text{d}^{-1}$

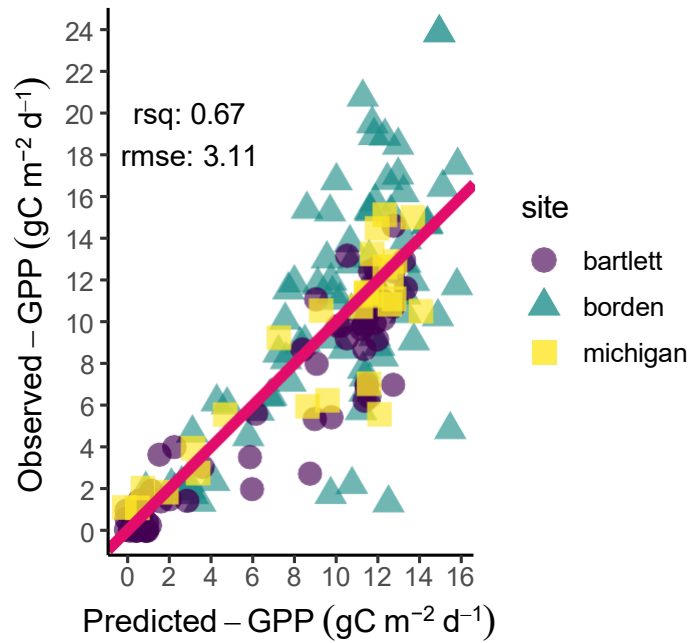


Figure 3.12.: GPP observed and predicted values from the autoML for all the sites at a daily basis. The red line represents a 1:1 relation. Metrics units are $\text{gC m}^{-2} \text{d}^{-1}$

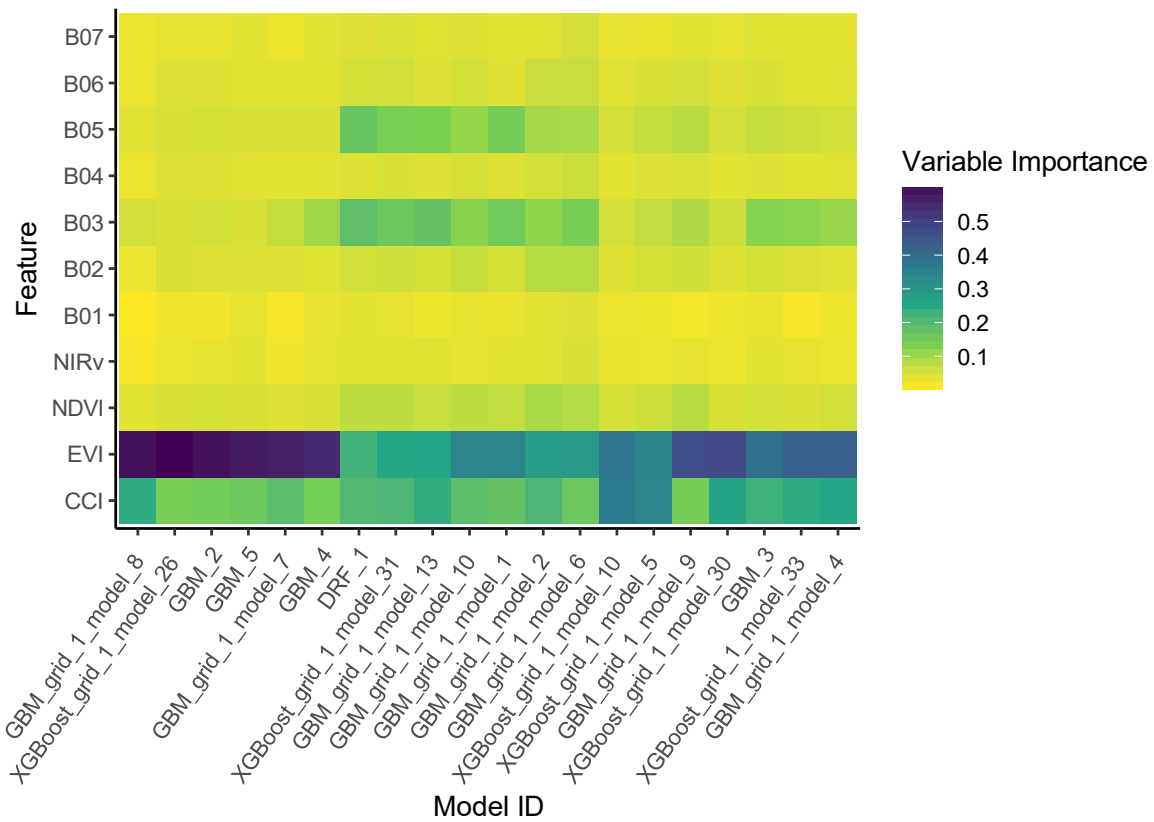


Figure 3.13.: Variable of importance derived from the autoML model for the daily values at 500 m spatial resolution model.

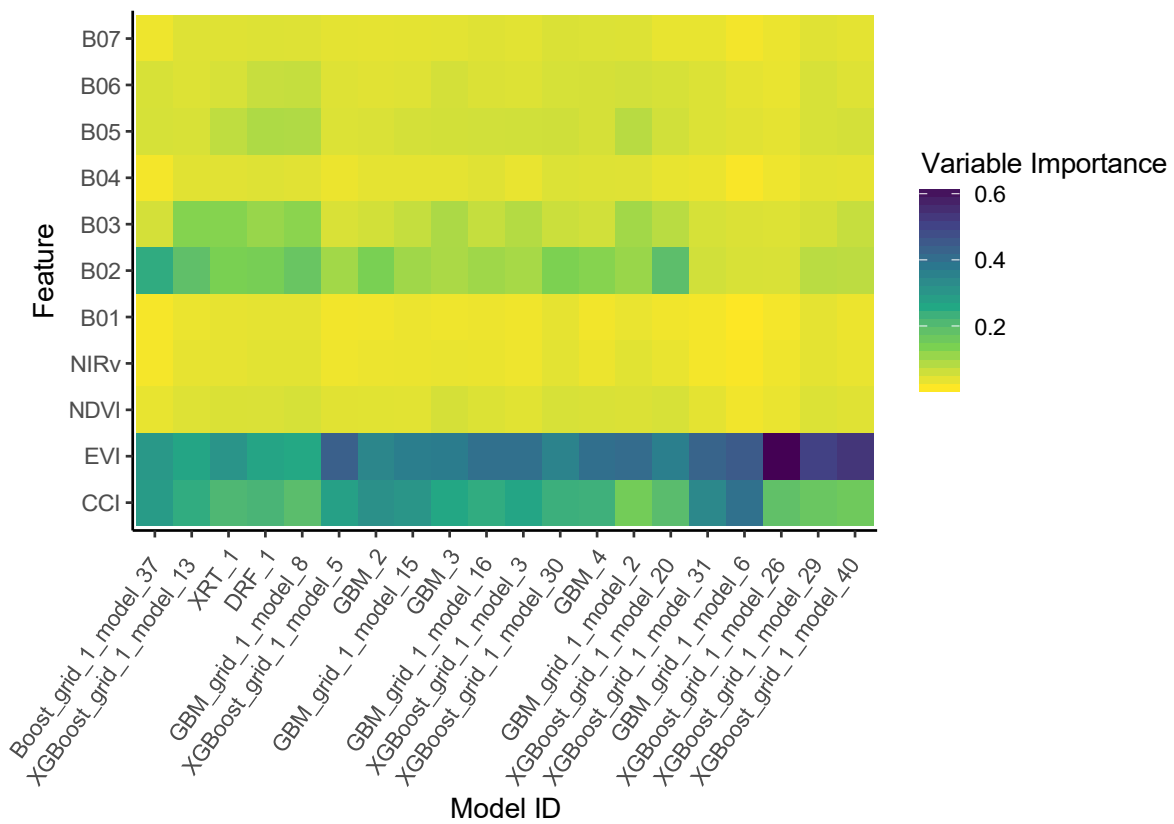


Figure 3.14.: Variable of importance derived from the autoML model for the weekly values at 500 m spatial resolution model.

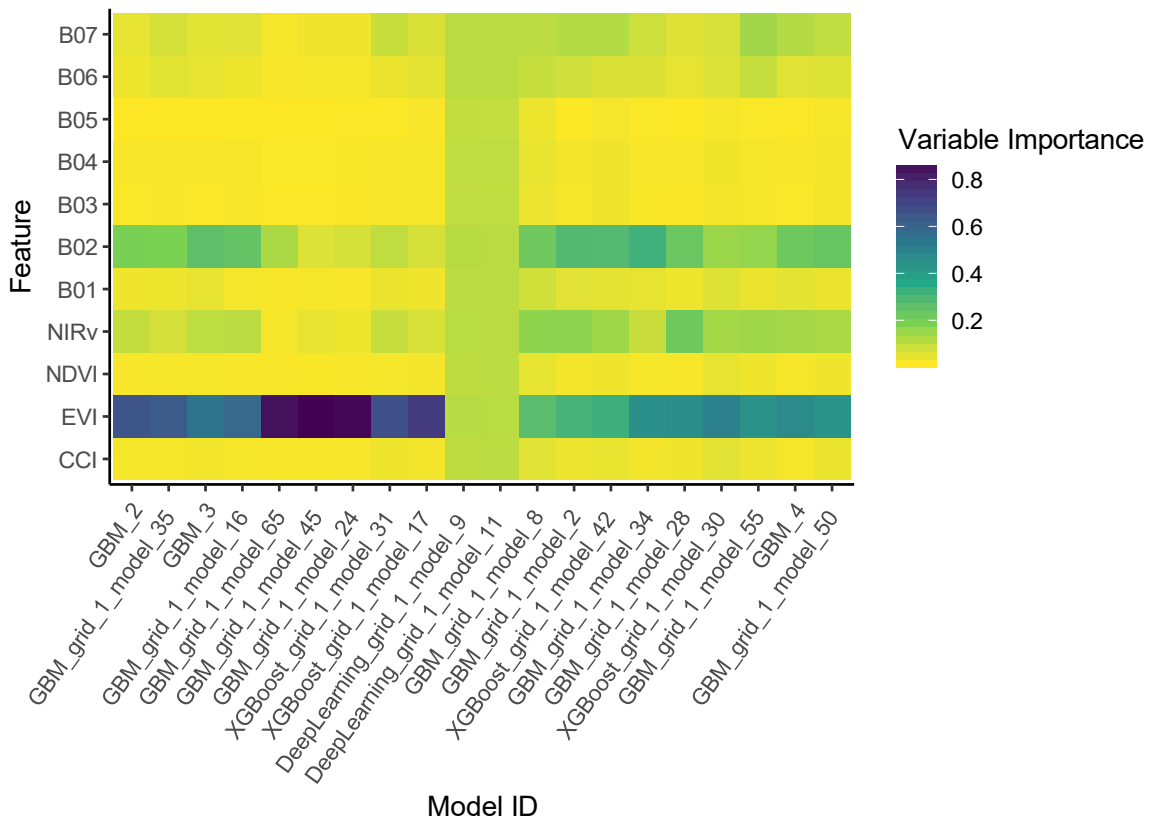


Figure 3.15.: Variable of importance derived from the autoML model for the monthly values at 500 m spatial resolution model.

AutoML model, both applied to the same datasets, revealed nuanced differences in their predictive performance Table 3.2. While the regression random forest model exhibited superior R^2 values in daily and monthly predictions, indicating a better overall fit to the data, the AutoML model demonstrated lower RMSE. Conversely, for weekly predictions, the AutoML model outperformed in both metrics. These findings underscore the importance of considering multiple metrics and temporal scales when evaluating and selecting models for GPP predictions.

Table 3.2.: Summary ML metrics in $\text{gC m}^{-2} \text{d}^{-1}$

| Variable | RF R^2 | RF RMSE | automl R^2 | automl RMSE |
|----------|----------|---------|--------------|-------------|
| Daily | 0.70 | 3.17 | 0.67 | 3.11 |
| Weekly | 0.55 | 3.23 | 0.72 | 3.08 |
| Monthly | 0.81 | 2.03 | 0.76 | 1.84 |

3.4. Discussion

These results underscore the temporal variations in model performance, with the monthly models having the best performance metrics for both, the regression random forest and the autoML. The lower RMSE of the AutoML model indicates it can potentially result in lower uncertainty for monthly predictions than RF. Given the higher R^2 , regression RF captures a larger proportion of the variance in the dependent variable than the autoML. This discrepancy underscores the importance of considering multiple evaluation metrics when assessing model performance. The choice between these models may depend on the specific goals of the analysis, weighing the trade-off between explaining variability and achieving precision in predictions.

Upon examining the six specific predictions analyzed through Shapley values for the Random Forest models across all timeframes, it became evident that high-predicted values of Gross Primary Productivity (GPP) were consistently underestimated, while low GPP values were overestimated. It is crucial to clarify that these particular predictions and Shapley values are specific to each instance and do not represent the entirety of potential predictions. However, when scrutinizing the graphs of predicted GPP values against observed GPP values, especially in the case of the daily model, it is noteworthy that the maximum predicted values hover around $16 \text{ gC m}^{-2} \text{ d}^{-1}$, whereas observed values can reach up to $24 \text{ gC m}^{-2} \text{ d}^{-1}$ (See Figure 3.2). Importantly, these values all belong to the Borden site, where the range of observed GPP is notably higher compared to the other two sites.

The discrepancy between the observed values and the predicted values may arise from a potential lag between the change in photosynthetic rate and the concentration of photosynthetic pigments, particularly the change in chlorophyll. Since the predicted GPP values for the constructed models are solely based on reflectance values, the primary changes related to GPP that they can capture are APAR and, to some extent, chlorophyll concentration (Pabon-Moreno et al. 2022). Given that photosynthesis can change rapidly without significant alterations in pigment concentrations, there might be an overestimation of predicted values compared to observed and estimated GPP values at the site mediating EC. In situations of stress, the dynamics of GPP could shift unnoticed solely based on satellite reflectance values.

Conversely, underestimation may be attributed to the well-known phenomenon of satura-

tion. When using indices created with the NIR band for GPP estimation, challenges arise in scenarios where there is a substantial increase in vegetation biomass, which happens during summers in the study sites. In such situations, the dense biomass leads to increased scattering and reflection of radiation. While the NIR band is sensitive to changes in vegetation structure and density, it encounters limitations as the amount of biomass intensifies. This results in a phenomenon known as saturation, where the sensor reaches its maximum capacity to detect changes in reflectance values.

Saturation occurs because the dense vegetation causes a greater proportion of the incoming radiation to be scattered or reflected, particularly in the NIR spectrum (Camps-Valls et al. 2021). As a consequence, the sensor becomes saturated, meaning that further increases in biomass or productivity do not translate proportionally into higher measured reflectance values. This limits the sensor's ability to capture and differentiate changes in the productivity of vegetation beyond a certain threshold.

While, for the monthly model, AutoML demonstrates an acceptable percentage of explained variability, with Random Forest surpassing it by 0.05, for the weekly and daily models, it appears that capturing the inherent variability in the data is challenging solely with the utilized indices and the entirety of MODIS bands. There seems to be a necessity to incorporate additional variables that possibly could further enhance predictive capabilities.

Regarding the question of whether it is preferable to employ a Random Forest (RF) model or an AutoML model, the advantages revealed by the results of this study do not exhibit marked distinctions. The performances are comparable, with the only instance of a notable increase, a $R^2 = 0.17$ improvement in explaining variance, observed in the weekly model by the AutoML model. In the monthly model, Random Forest demonstrates a $R^2 = 0.05$ superiority in terms of explained variance compared to AutoML, although AutoML achieves a marginal improvement in error reduction by 0.198. These nuanced differences suggest that the choice between RF and AutoML may depend on specific considerations, emphasizing the importance of assessing both explanatory power and error metrics for comprehensive model evaluation.

3.5. Conclusions

In summary, the monthly models generated using both methods (RF and autoML) exhibited superior performance based on various metrics, surpassing the outcomes from weekly or daily temporal aggregations. Specifically, in the context of monthly modelling, although RF demonstrated a slightly higher variance explanation compared to autoML (by 0.05), autoML showcased a lower RMSE, signifying more accurate predictions with minimal error. When incorporating VIs alongside all bands into the GPP prediction models (both RF and autoML), the VIs consistently demonstrated their significance as the most influential variables in the predictions, with CCI and EVI consistently having pivotal roles. Notably, among the bands, B02 (NIR) emerged as the most crucial for predictions, surpassing the importance of the remaining bands.

Analyzing the RF model revealed interesting insights through Shapley values, indicating a tendency to underestimate high GPP values and overestimate low GPP values. This observation was consistent across scatterplots depicting predicted versus actual values for diverse temporal aggregations. When comparing both methods (RF and autoML), no definitive superiority emerges, as both exhibit nuanced distinctions, except for the weekly models where autoML outperforms with a 0.17 better variance explanation and a lower RMSE.

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4. Conclusions and Future Work

4.1. Synthesis

This research evaluated the advantages and limitations associated with estimating GPP using traditional VIs like NDVI, EVI, and recent ones like CCI and NIRv. Specifically, the study focused on VIs derived from MODIS, chosen for its extensive time records. To assess GPP estimates through the VIs, we employed traditional methods such as linear regression and GAM, as well as ML techniques.

Traditional methods, such as LM and GAM, offer the advantage of being more straightforward to interpret in relation to their predictions and associated errors. In contrast, ML techniques like regression RF and autoML pursue predictions through methods that are harder to elucidate, a phenomenon referred to as “black model boxes.” Despite their complexity, these ML methods have the potential to achieve superior accuracy in their predictions.

To evaluate the capacity of VIs to estimate GPP, Productivity (GPP), our analysis spanned three temporal aggregations: daily, weekly, and monthly. This assessment considered the utilization of VIs individually or collectively as covariates in the model. For ML methods, the same temporal aggregations were employed, with an assessment involving all indices and the entire set of MODIS bands as predictors.

The results indicate that, across all models, monthly temporal aggregation consistently yields better outcomes in comparison to weekly and daily aggregations. This superiority can be attributed to the reduced variation present in monthly aggregated data, resulting in lower errors in the models. According to LM and GAM, incorporating all indices as covariates in the models enhances predictive capability, suggesting that a single VI in isolation may fall short of capturing the entire variation in the data.

Examining the outcomes of models utilizing individual VIs reveals minimal differences. Nonetheless, VIs consistently displaying a slightly superior performance in terms of explained variation and lower error are typically EVI and CCI, whereas NDVI tends to exhibit lower performance. This trend is also evident in ML models, with EVI and CCI frequently occupying top positions in variable of importance rankings. In ML models, the inclusion of all bands as predictors highlights that B02 (NIR) often ranks among the top variables in importance over other bands.

In comparing the various models and their configurations (LM and GAM with VIs as co-variates; RF and autoML with VIs and complete bands as predictors), it becomes evident that ML methods exhibit a better performance across different time aggregations. For daily models, assessed by R^2 and RMSE metrics, RF surpassed autoML, GAM, and LM, in that sequence. For the weekly models, autoML led, followed by GAM, LM, and RF. Regarding the monthly models, RF had better performance, followed by LM, autoML, and GAM.

4.2. Limitations

The uncertainties in VIs, in this case, for GPP estimation, may be due to their inability to capture variations in GPP or inherent limitations falling into two categories: artifacts or external factors. Artifacts originate from the sensors used for deriving VIs, encompassing calibration, quality control, or sensor degradation (Zeng et al. 2022). In this study, we addressed these effects by excluding pixels with suboptimal quality using data from quality assurance (QA) and quality control (QC). These variables offer insights into the quality of each pixel, aiding in the selection of pixels potentially harboring erroneous values. However, the accuracy of these data may be compromised by the presence of pixels with poor quality not identified as such by postprocessing algorithms, or influenced by external factors such as clouds, cloud shadows, inaccuracies in snow detection, atmospheric pollution, or technical issues in image acquisition (Pesquer, Domingo-Marimon, and Pons 2019).

Another potential source of uncertainty in GPP estimates is related to the representation of the flux tower footprint within the pixel area. EC towers exhibit a flux footprint that can vary over time. The footprint is defined as the extent to which measurements taken at a specific time and location accurately reflect the actual flows in time and specific area. The monthly climatologies of the footprint can exhibit variations, usually falling within the range

of 100 m to 450 m. (Chu et al. 2021). The heterogeneity of the landscape beyond this footprint may affect flow conditions differently than those measured within the footprint. In our methodology, we established a square area of 3 km² around the EC tower to ensure that the flux footprint climatology aligns with this region, considering the homogeneity in land cover type. However, the resolution of MODIS might not be adequate for capturing nuanced fine-scale spatial variations (Robinson et al. 2018).

The limitations of the study include the unavailability of data from a sensor like Sentinel-2, which offers better spatial resolution compared to the MODIS product employed. This data could have allowed us to explore whether, despite estimating in areas with homogeneous land cover, it is feasible to capture nuanced fine-scale spatial variations. Such an exploration could potentially lead to improved GPP estimates, thereby reducing uncertainty. The restriction in using Sentinel-2 data stems from the fact that calibrated data is accessible only from 2017 onward. However, the GPP data processed by ONEFlux for each site commenced in 2015, concluding for Bartlett and Michigan in 2017 and 2018, respectively, thus resulting in the exclusion of a significant portion of the data.

Finally, it is important to acknowledge that the sites selection can constitute a potential limitation in this study. The focus on deciduous broadleaf forest sites in the northern hemisphere, encompassing only three locations, restricts the generalization of the findings to this specific ecosystem type. Expanding the scope to include more sites representing other various ecosystem types could further contribute to understanding which indices have limitations or can perform better in estimating GPP across a broader range of scenarios.

4.3. Future work

As uncertainties persist in GPP estimation models relying on VIs as predictors, the potential for improvement lies in models that can consider critical stress factors influencing photosynthesis (Rogers et al. 2020; Ryu, Berry, and Baldocchi 2019; Xiao et al. 2019). While incorporating meteorological data can address this, it poses challenges in various regions around the globe. In such circumstances, a more advantageous approach involves the utilization and validation of VIs that demonstrate sensitivity to stress factors affecting photosynthetic activity, not just photosynthetic capacity. Remote sensing provides a means to achieve this by integrating data products such as thermal information, which can indirectly address water deficit

in ecosystems, for example (Pabon-Moreno et al. 2022). Despite potential uncertainties associated with these products, there is a prospect for enhanced estimations by incorporating considerations beyond traditional reflectance values.

Addressing the inherent variability across ecosystems, it is essential to recognize that no single VI will be the best, and their representation of vegetation functioning has limitations (Zeng et al. 2022). Nonetheless, the continual introduction of new VIs derived from both traditional and advanced sensors (Montero et al. 2023) presents an opportunity to enhance the estimation of GPP on a global level. A potentially effective strategy would involve establishing a standardized guide of VIs or combinations tailored to different ecosystem types, providing researchers with a reliable foundation for accurate assessments. Furthermore, the field could benefit from adopting a qualitative rating of model performance, such as the one used by for the FAO model AquaCrop (Raes et al. 2022). This practice involves standardizing model evaluations based on their performance within specific research domains. Such standardized benchmarks can provide valuable guidance for researchers seeking optimal model selection and interpretation within their respective contexts.

Given the growing reliance on ML methods to enhance GPP predictions in recent years, a notable challenge arises from their tendency to obscure the understanding of the underlying mechanisms explaining the relationship between VIs and GPP (Molnar 2020). Hence, it becomes crucial for new studies to articulate their research objectives with clarity and consistency. Specifically, researchers should explicitly state whether their primary focus is to explore mechanistic relationships between VIs or if their aim is to improve GPP predictions. This distinction is vital, whether the goal is to predict GPP in sites lacking validation data (flux towers) or to advance forecasting capabilities (Meyer et al. 2019).

The implementation of explainable ML methods could enhance our understanding of the variables influencing predictions (Molnar 2020). Additionally, it is well-known that a prediction model's ability to generalize the rules it has learned from the training data set to a new (unseen) data set is poor. Diverse scenarios, such as droughts, intense rainfall, and vegetation mortality, among others, can affect predictive capacity. Therefore, consistent model predictive power assessment over time and retraining as necessary are crucial (Kuhn and Silge 2022). This underscores the significance of having flux towers and extending their presence to underrepresented sites to acquire validation data. These efforts contribute to refining future global GPP estimation models on a broader scale (Meyer et al. 2019).

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A. Appendices

A.1. Bitstrings tables

A.1.1. MODIS State 1km

Table A.1.: state_1km bit strings

| Bit | State |
|---------------------|--------------------------------------|
| Cloud State | |
| 00 | clear |
| 01 | cloudy |
| 10 | mixed |
| 11 | not set, assumed clear |
| Cloud Shadow | |
| 1 | yes |
| 0 | no |
| Land Water | |
| 000 | shallow ocean |
| 001 | land |
| 010 | ocean coastlines and lake shorelines |
| 011 | shallow inland water |
| 100 | ephemeral water |
| 101 | deep inland water |
| 110 | continental/moderate ocean |
| 111 | deep ocean |

Aerosol Quantity

| | |
|----|-------------|
| 00 | climatology |
| 01 | low |
| 10 | average |
| 11 | high |

Cirrus Detected

| | |
|----|---------|
| 00 | none |
| 01 | small |
| 10 | average |
| 11 | high |

Cloud Flag

| | |
|---|----------|
| 1 | cloud |
| 0 | no cloud |

Fire Flag

| | |
|---|---------|
| 1 | fire |
| 0 | no fire |

Snow Ice Flag

| | |
|---|-----|
| 1 | yes |
| 0 | no |

Pixel adjacent Cloud

| | |
|---|-----|
| 1 | yes |
| 0 | no |

Salt Pan Cloud

| | |
|---|-----|
| 1 | yes |
| 0 | no |

Snow Mask

| | |
|---|-----|
| 1 | yes |
| 0 | no |

A.1.2. MODIS QC Scan

Table A.2.: qc_scan bit strings

| Bits | State |
|---|---|
| Modland | |
| 00 | ideal quality - all bands |
| 01 | less than ideal quality - some or all bands |
| 10 | product not produced due to cloud effects |
| 11 | product not produced for other reasons |
| Band quality (apply for all bands) | |
| 0000 | highest_quality |
| 0111 | noisy detector |
| 1000 | dead detector, data interpolated in L1B |
| 1001 | solar zenith ≥ 86 degrees |
| 1010 | solar zenith ≥ 85 and < 86 degrees |
| 1011 | missing input |
| 1100 | internal constant used |
| 1101 | correction out of bounds |
| 1110 | L1B data faulty |
| 1111 | not processed due to deep ocean or clouds |
| Atmospheric Correction | |
| 0 | no |
| 1 | yes |
| Adjacency Correction | |
| 0 | no |
| 1 | yes |

A.2. Complete LM metrics

A.2.1. Monthly LM metrics

Table A.3.: Summary of Linear models for GPP estimation using the vegetation indices on a monthly basis (per site).

| | adj.r.squared | statistic | AIC | BIC | p.value |
|----------|---------------|-----------|---------|---------|-------------------------|
| Michigan | | | | | |
| All | 0.918 | 48.285 | 59.460 | 64.802 | 1.120×10^{-7} |
| CCI | 0.864 | 109.210 | 66.171 | 68.842 | 1.483×10^{-8} |
| NIRv | 0.856 | 102.166 | 67.214 | 69.885 | 2.366×10^{-8} |
| EVI | 0.850 | 97.219 | 67.983 | 70.654 | 3.341×10^{-8} |
| NDVI | 0.699 | 40.436 | 80.515 | 83.186 | 9.481×10^{-6} |
| Bartlett | | | | | |
| All | 0.889 | 34.995 | 60.577 | 65.920 | 7.627×10^{-7} |
| CCI | 0.887 | 133.808 | 58.692 | 61.363 | 3.495×10^{-9} |
| EVI | 0.880 | 125.716 | 59.691 | 62.363 | 5.466×10^{-9} |
| NIRv | 0.878 | 123.297 | 60.001 | 62.672 | 6.280×10^{-9} |
| NDVI | 0.784 | 62.655 | 70.289 | 72.960 | 6.363×10^{-7} |
| Borden | | | | | |
| CCI | 0.782 | 133.796 | 177.703 | 182.616 | 1.107×10^{-13} |
| All | 0.778 | 33.452 | 181.071 | 190.897 | 3.485×10^{-11} |
| NDVI | 0.755 | 114.882 | 182.191 | 187.104 | 9.428×10^{-13} |
| NIRv | 0.746 | 109.643 | 183.534 | 188.447 | 1.790×10^{-12} |
| EVI | 0.738 | 105.340 | 184.673 | 189.586 | 3.086×10^{-12} |
| All | | | | | |
| All | 0.771 | 62.572 | 327.544 | 341.368 | 3.117×10^{-22} |
| CCI | 0.732 | 200.153 | 336.514 | 343.426 | 1.766×10^{-22} |
| NDVI | 0.638 | 129.875 | 358.619 | 365.531 | 8.818×10^{-18} |
| NIRv | 0.583 | 103.147 | 369.129 | 376.041 | 1.528×10^{-15} |
| EVI | 0.566 | 96.172 | 372.136 | 379.048 | 6.694×10^{-15} |

A.2.2. Weekly LM metrics

Table A.4.: Summary of Linear models for GPP estimation using the vegetation indices on a weekly basis (per site).

| | adj.r.squared | statistic | AIC | BIC | p.value |
|----------|---------------|-----------|-----------|-----------|-------------------------|
| Bartlett | | | | | |
| All | 0.809 | 77.188 | 271.583 | 285.326 | 1.513×10^{-24} |
| EVI | 0.787 | 267.599 | 276.529 | 283.400 | 8.712×10^{-26} |
| CCI | 0.783 | 260.798 | 278.010 | 284.881 | 1.795×10^{-25} |
| NIRv | 0.770 | 241.637 | 282.352 | 289.224 | 1.495×10^{-24} |
| NDVI | 0.633 | 125.324 | 316.317 | 323.189 | 2.444×10^{-17} |
| Michigan | | | | | |
| All | 0.799 | 64.530 | 263.238 | 276.284 | 4.684×10^{-21} |
| NIRv | 0.782 | 229.980 | 265.761 | 272.284 | 1.060×10^{-22} |
| EVI | 0.780 | 228.274 | 266.140 | 272.663 | 1.275×10^{-22} |
| CCI | 0.701 | 151.227 | 286.110 | 292.633 | 2.138×10^{-18} |
| NDVI | 0.609 | 100.592 | 303.638 | 310.161 | 1.116×10^{-14} |
| Borden | | | | | |
| All | 0.600 | 53.920 | 761.297 | 779.032 | 3.178×10^{-27} |
| CCI | 0.577 | 193.515 | 766.307 | 775.174 | 3.589×10^{-28} |
| EVI | 0.572 | 189.314 | 768.107 | 776.974 | 8.756×10^{-28} |
| NIRv | 0.563 | 182.981 | 770.864 | 779.732 | 3.434×10^{-27} |
| NDVI | 0.499 | 141.470 | 790.399 | 799.267 | 5.537×10^{-23} |
| All | | | | | |
| CCI | 0.560 | 356.273 | 1,445.058 | 1,455.963 | 1.022×10^{-51} |
| All | 0.558 | 89.013 | 1,449.447 | 1,471.255 | 1.971×10^{-48} |
| NDVI | 0.461 | 239.572 | 1,501.991 | 1,512.895 | 2.115×10^{-39} |
| NIRv | 0.459 | 237.796 | 1,502.954 | 1,513.858 | 3.417×10^{-39} |
| EVI | 0.452 | 230.689 | 1,506.838 | 1,517.743 | 2.370×10^{-38} |

A.2.3. Daily LM metrics

Table A.5.: Summary of Linear models for GPP estimation using the vegetation indices on a daily basis (per site).

| | adj.r.squared | statistic | AIC | BIC | p.value |
|----------|---------------|-----------|-----------|-----------|--------------------------|
| Bartlett | | | | | |
| All | 0.866 | 485.890 | 1,178.271 | 1,200.533 | 6.162×10^{-129} |
| CCI | 0.832 | 1,489.205 | 1,243.242 | 1,254.374 | 2.354×10^{-118} |
| EVI | 0.813 | 1,310.145 | 1,275.087 | 1,286.219 | 1.762×10^{-111} |
| NIRv | 0.800 | 1,201.359 | 1,296.213 | 1,307.345 | 6.406×10^{-107} |
| NDVI | 0.657 | 576.639 | 1,458.698 | 1,469.830 | 7.912×10^{-72} |
| Michigan | | | | | |
| All | 0.713 | 91.565 | 669.582 | 687.525 | 2.499×10^{-38} |
| EVI | 0.702 | 345.680 | 671.830 | 680.802 | 3.248×10^{-40} |
| NIRv | 0.664 | 289.199 | 689.807 | 698.778 | 2.366×10^{-36} |
| CCI | 0.610 | 229.492 | 711.553 | 720.524 | 1.121×10^{-31} |
| NDVI | 0.521 | 159.932 | 741.759 | 750.730 | 3.569×10^{-25} |
| All | | | | | |
| All | 0.606 | 320.497 | 4,522.636 | 4,550.986 | 1.640×10^{-166} |
| CCI | 0.591 | 1,201.855 | 4,550.671 | 4,564.846 | 1.355×10^{-163} |
| EVI | 0.494 | 812.661 | 4,727.696 | 4,741.871 | 3.301×10^{-125} |
| NIRv | 0.475 | 755.095 | 4,757.393 | 4,771.569 | 9.120×10^{-119} |
| NDVI | 0.447 | 672.798 | 4,801.777 | 4,815.952 | 3.873×10^{-109} |
| Borden | | | | | |
| All | 0.521 | 104.996 | 2,181.484 | 2,205.188 | 4.203×10^{-60} |
| CCI | 0.441 | 303.652 | 2,237.254 | 2,249.106 | 1.838×10^{-50} |
| EVI | 0.428 | 287.124 | 2,246.624 | 2,258.475 | 1.973×10^{-48} |
| NIRv | 0.380 | 235.703 | 2,277.328 | 2,289.180 | 8.977×10^{-42} |
| NDVI | 0.273 | 144.920 | 2,338.369 | 2,350.221 | 1.617×10^{-28} |

A.3. Complete GAM metrics

A.3.1. Monthly GAM metrics

Table A.6.: Summary of the GAM models output for monthly GPP estimation using the every VI as an individual non-linear term

| site | index | edf | f | p_value | AIC |
|----------|-----------|----------|-----------|---------------|-----------|
| Michigan | evi_mean | 3.817030 | 87.35043 | 0.000000e+00 | 46.78342 |
| Michigan | nirv_mean | 4.044975 | 75.37774 | 0.000000e+00 | 48.41677 |
| Bartlett | cci_mean | 2.239036 | 66.85647 | 0.000000e+00 | 54.60874 |
| Michigan | cci_mean | 2.602984 | 57.48746 | 0.000000e+00 | 59.36252 |
| Bartlett | evi_mean | 1.000019 | 125.71069 | 0.000000e+00 | 59.69151 |
| Bartlett | nirv_mean | 1.000030 | 123.28934 | 0.000000e+00 | 60.00142 |
| Michigan | ndvi_mean | 2.806127 | 31.20277 | 5.690552e-06 | 68.06940 |
| Bartlett | ndvi_mean | 1.000069 | 62.64577 | -3.515872e-07 | 70.28904 |
| Borden | cci_mean | 1.190879 | 96.46348 | 0.000000e+00 | 177.64856 |
| Borden | ndvi_mean | 1.809503 | 53.64331 | 0.000000e+00 | 181.16487 |
| Borden | nirv_mean | 1.000106 | 109.62110 | 0.000000e+00 | 183.53397 |
| Borden | evi_mean | 1.000118 | 105.31638 | 0.000000e+00 | 184.67361 |
| All | cci_mean | 2.404895 | 72.32912 | 0.000000e+00 | 333.05822 |
| All | ndvi_mean | 2.954201 | 51.51928 | 0.000000e+00 | 341.40009 |
| All | nirv_mean | 3.770256 | 31.19524 | 0.000000e+00 | 355.55683 |
| All | evi_mean | 3.789326 | 29.98655 | 0.000000e+00 | 357.45536 |

Table A.7.: Summary of the GAM models output for monthly GPP estimation using the all VIs per site category as non-linear terms covariates

| site | index | term | edf | f | p_value | AIC |
|--------|-------|--------------|----------|-----------|--------------|-----------|
| Borden | All | s(evi_mean) | 1.000007 | 1.561538 | 0.2202315969 | 181.0712 |
| Borden | All | s(nirv_mean) | 1.000006 | 1.832159 | 0.1850688800 | 181.0712 |
| Borden | All | s(cci_mean) | 1.000100 | 2.667602 | 0.1118925771 | 181.0712 |
| All | All | s(evi_mean) | 5.154191 | 4.079875 | 0.0004261942 | 4411.7779 |
| All | All | s(nirv_mean) | 4.545057 | 2.844737 | 0.0125690226 | 4411.7779 |
| All | All | s(cci_mean) | 6.218002 | 13.246895 | 0.0000000000 | 4411.7779 |

A.3.2. Weekly GAM metrics

Table A.8.: Summary of the GAM models output for weekly GPP estimation using the every VI as an individual non-linear term

| site | index | edf | f | p_value | AIC |
|----------|-----------|----------|-----------|---------|-----------|
| Michigan | evi_mean | 3.663726 | 75.39276 | 0 | 247.2844 |
| Michigan | nirv_mean | 3.209376 | 71.49144 | 0 | 256.7204 |
| Bartlett | evi_mean | 3.336909 | 71.86998 | 0 | 272.0930 |
| Bartlett | cci_mean | 2.583077 | 86.30588 | 0 | 275.5742 |
| Michigan | cci_mean | 3.520166 | 43.76791 | 0 | 277.5737 |
| Bartlett | nirv_mean | 3.272579 | 65.98375 | 0 | 278.0024 |
| Michigan | ndvi_mean | 2.675203 | 39.90565 | 0 | 293.9057 |
| Bartlett | ndvi_mean | 1.000087 | 125.30174 | 0 | 316.3173 |
| Borden | evi_mean | 3.560463 | 46.50232 | 0 | 762.7106 |
| Borden | cci_mean | 1.000274 | 193.41184 | 0 | 766.3073 |
| Borden | nirv_mean | 2.853504 | 54.04690 | 0 | 767.5482 |
| Borden | ndvi_mean | 1.406799 | 80.37055 | 0 | 790.2338 |
| All | cci_mean | 4.538741 | 69.58626 | 0 | 1433.6512 |
| All | evi_mean | 4.748375 | 53.58396 | 0 | 1468.9599 |
| All | nirv_mean | 4.334480 | 55.98519 | 0 | 1473.8814 |
| All | ndvi_mean | 3.223579 | 71.73012 | 0 | 1480.0666 |

Table A.9.: Summary of the GAM models output for weekly GPP estimation using the all VIs per site category as non-linear terms covariates

| site | index | term | edf | f | p_value | AIC |
|----------|-------|--------------|----------|------------|--------------|-----------|
| Michigan | All | s(evi_mean) | 4.488458 | 5.8177458 | 0.0001315731 | 240.5091 |
| Michigan | All | s(nirv_mean) | 1.000195 | 9.7009559 | 0.0028789998 | 240.5091 |
| Michigan | All | s(cci_mean) | 1.000043 | 0.0175387 | 0.8952502747 | 240.5091 |
| Bartlett | All | s(evi_mean) | 1.000051 | 3.1454363 | 0.0806776572 | 269.6656 |
| Bartlett | All | s(nirv_mean) | 1.000129 | 0.5997498 | 0.4413933495 | 269.6656 |
| Bartlett | All | s(cci_mean) | 2.128825 | 4.2538207 | 0.0095109230 | 269.6656 |
| Borden | All | s(evi_mean) | 1.001795 | 3.0976955 | 0.0804966509 | 760.4175 |
| Borden | All | s(nirv_mean) | 1.976616 | 0.6214367 | 0.5421077236 | 760.4175 |
| Borden | All | s(cci_mean) | 1.000123 | 2.6980022 | 0.1027452161 | 760.4175 |
| All | All | s(evi_mean) | 5.154191 | 4.0798752 | 0.0004261942 | 4411.7779 |
| All | All | s(nirv_mean) | 4.545057 | 2.8447368 | 0.0125690226 | 4411.7779 |
| All | All | s(cci_mean) | 6.218002 | 13.2468950 | 0.0000000000 | 4411.7779 |

A.3.3. Daily GAM metrics

Table A.10.: Summary of the GAM models output for daily GPP estimation using the every VI as an individual non-linear term

| site | index | edf | f | p_value | AIC |
|----------|-----------|----------|-----------|---------|-----------|
| Michigan | evi_mean | 4.457179 | 77.72707 | 0 | 652.5681 |
| Michigan | nirv_mean | 2.807163 | 86.23596 | 0 | 687.8352 |
| Michigan | cci_mean | 4.218361 | 48.29421 | 0 | 704.7992 |
| Michigan | ndvi_mean | 2.980300 | 55.52978 | 0 | 722.8092 |
| Bartlett | cci_mean | 5.975808 | 236.42420 | 0 | 1215.1703 |
| Bartlett | evi_mean | 5.365946 | 250.56443 | 0 | 1222.4305 |
| Bartlett | nirv_mean | 5.461424 | 226.11564 | 0 | 1244.9206 |
| Bartlett | ndvi_mean | 5.198974 | 137.44546 | 0 | 1374.8503 |
| Borden | evi_mean | 5.569955 | 55.49361 | 0 | 2202.2988 |
| Borden | nirv_mean | 5.804760 | 46.01083 | 0 | 2230.5977 |
| Borden | cci_mean | 3.191225 | 78.29535 | 0 | 2231.2904 |
| Borden | ndvi_mean | 4.912790 | 45.99191 | 0 | 2235.6639 |
| All | cci_mean | 6.266321 | 173.70000 | 0 | 4515.9138 |
| All | evi_mean | 6.642583 | 148.32392 | 0 | 4574.2505 |
| All | ndvi_mean | 5.308406 | 155.72087 | 0 | 4615.5740 |
| All | nirv_mean | 6.317347 | 140.36124 | 0 | 4618.3302 |

Table A.11.: Summary of the GAM models output for daily GPP estimation using the all VIs per site category as non-linear terms covariates

| site | index | term | edf | f | p_value | AIC |
|----------|-------|--------------|----------|-----------|--------------|----------|
| Michigan | All | s(evi_mean) | 3.012297 | 15.313284 | 0.000000e+00 | 630.891 |
| Michigan | All | s(nirv_mean) | 1.953289 | 11.098293 | 1.465150e-05 | 630.891 |
| Michigan | All | s(cci_mean) | 3.316744 | 2.334852 | 5.228136e-02 | 630.891 |
| Bartlett | All | s(evi_mean) | 5.075549 | 6.261981 | 3.136046e-06 | 1139.659 |
| Bartlett | All | s(nirv_mean) | 3.933390 | 2.321316 | 5.093414e-02 | 1139.659 |
| Bartlett | All | s(cci_mean) | 5.523614 | 9.113037 | 0.000000e+00 | 1139.659 |
| Borden | All | s(evi_mean) | 1.001559 | 22.542139 | 3.092503e-06 | 2147.143 |
| Borden | All | s(nirv_mean) | 4.720920 | 4.428707 | 2.739846e-04 | 2147.143 |
| Borden | All | s(cci_mean) | 3.652172 | 2.062896 | 9.147704e-02 | 2147.143 |
| All | All | s(evi_mean) | 5.154191 | 4.079875 | 4.261942e-04 | 4411.778 |
| All | All | s(nirv_mean) | 4.545057 | 2.844737 | 1.256902e-02 | 4411.778 |
| All | All | s(cci_mean) | 6.218002 | 13.246895 | 0.000000e+00 | 4411.778 |

A.4. Residuals distributions

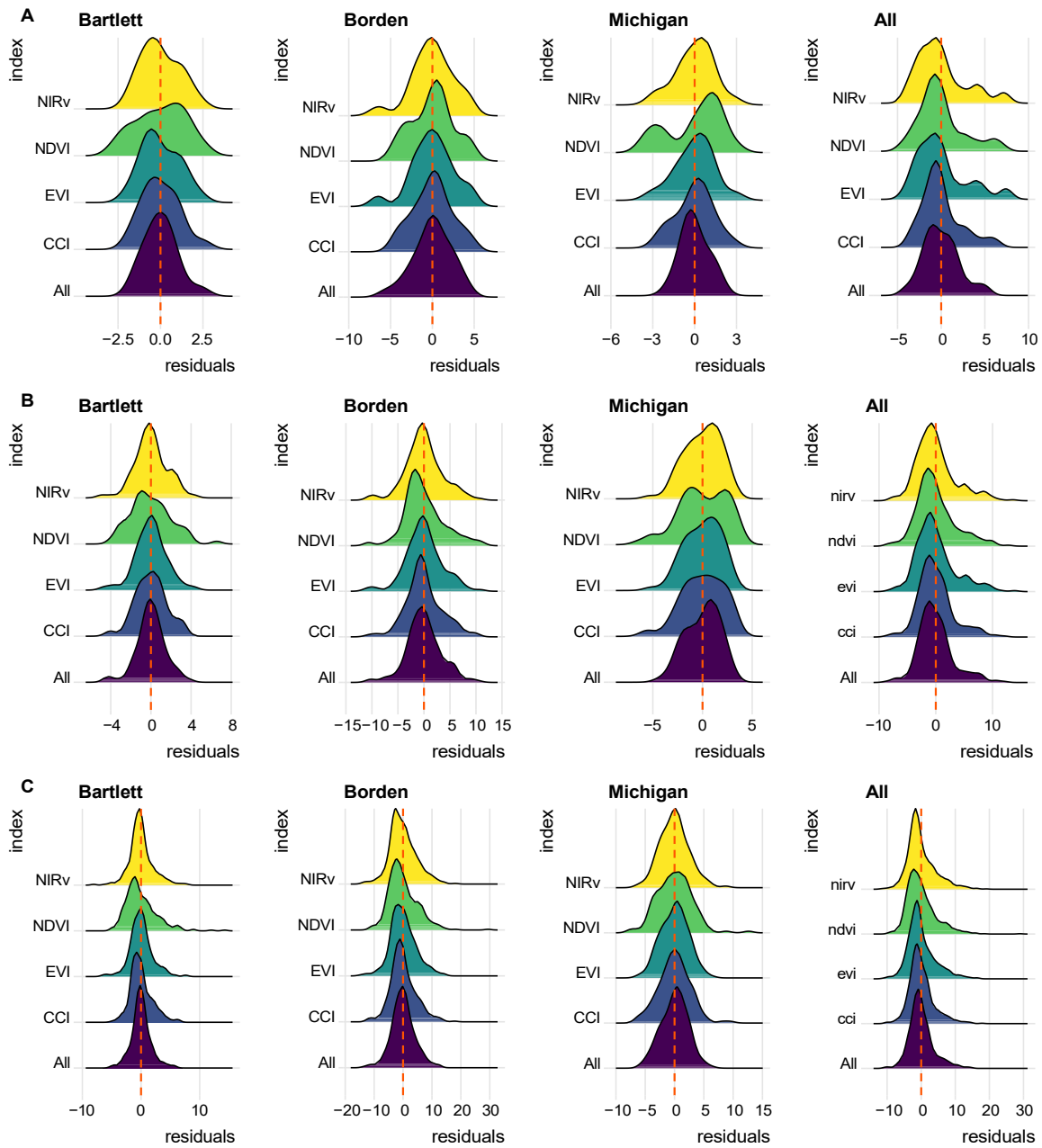


Figure A.1.: Residuals distributions for each of the LMs for GPP estimation using the vegetation indices on a monthly (a), weekly (b), and daily (c) basis.

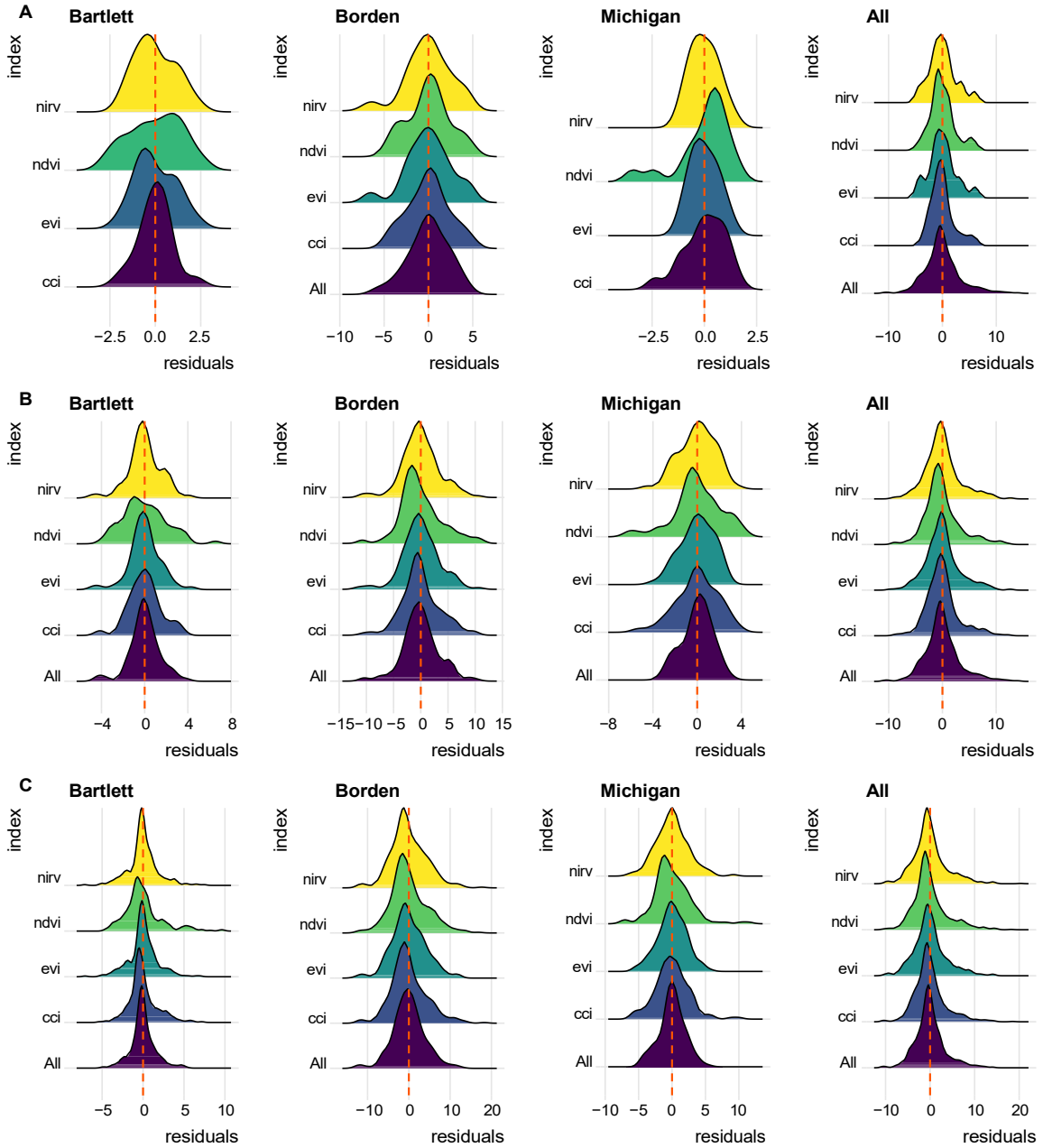


Figure A.2.: Residuals distributions for each of the GAMs for GPP estimation using the vegetation indices on a monthly (a), weekly (b), and daily (c) basis.